

Appendix H

SWAT – Soil and Water Assessment Tool

**Predicting Phosphorus and TSS Export with the Soil and Water
Assessment Tool (SWAT) to Evaluate Alternative Agricultural
Management Practices in the Big Green Lake Watershed, Wisconsin**

4/15/2000, Final Draft
Prepared by Fox-Wolf Basin 2000
Paul Baumgart

Objective

The primary objective of this modeling project was to provide a predictive tool that could be used to estimate the potential for phosphorus and TSS load reductions in the Big Green Lake Watershed by assessing the impact of alternative management scenarios on total phosphorus and TSS loads to Big Green Lake. To accomplish this objective, the Soil and Water Assessment Tool (SWAT) was applied to the Big Green Lake Watershed. SWAT was developed by USDA-ARS to improve the technology used in the SWRRBWQ model (Arnold et al. 1996). SWAT is a distributed parameter, daily time step model that was developed to assess non-point source pollution from watersheds and large river basins. SWAT simulates hydrologic and related processes to predict the impact of management on water, sediment, nutrient and pesticide export from rural basins. A more detailed description of this model can be found in Appendix A.

This report describes: (1) the derivation of SWAT inputs; (2) model set up, calibration and assessment; and (3) the predicted impacts of alternative management scenarios on simulated loads of phosphorus and TSS to Big Green Lake.

Watershed description

The Big Green Lake Watershed is located primarily in Green Lake and Fond du Lac Counties, but a small portion of the watershed is located in Winnebago County (Figure 1). Big Green Lake is the deepest lake in Wisconsin, and it is the primary surface water feature in the watershed with an area of 7,325 acres (29.6 km²). Other lakes in the watershed include Spring, Big Twin and Little Twin. As shown in Figure 1, the dominant land cover in the 244 km² Big Green Lake Watershed is agriculture (area without Big Green Lake).

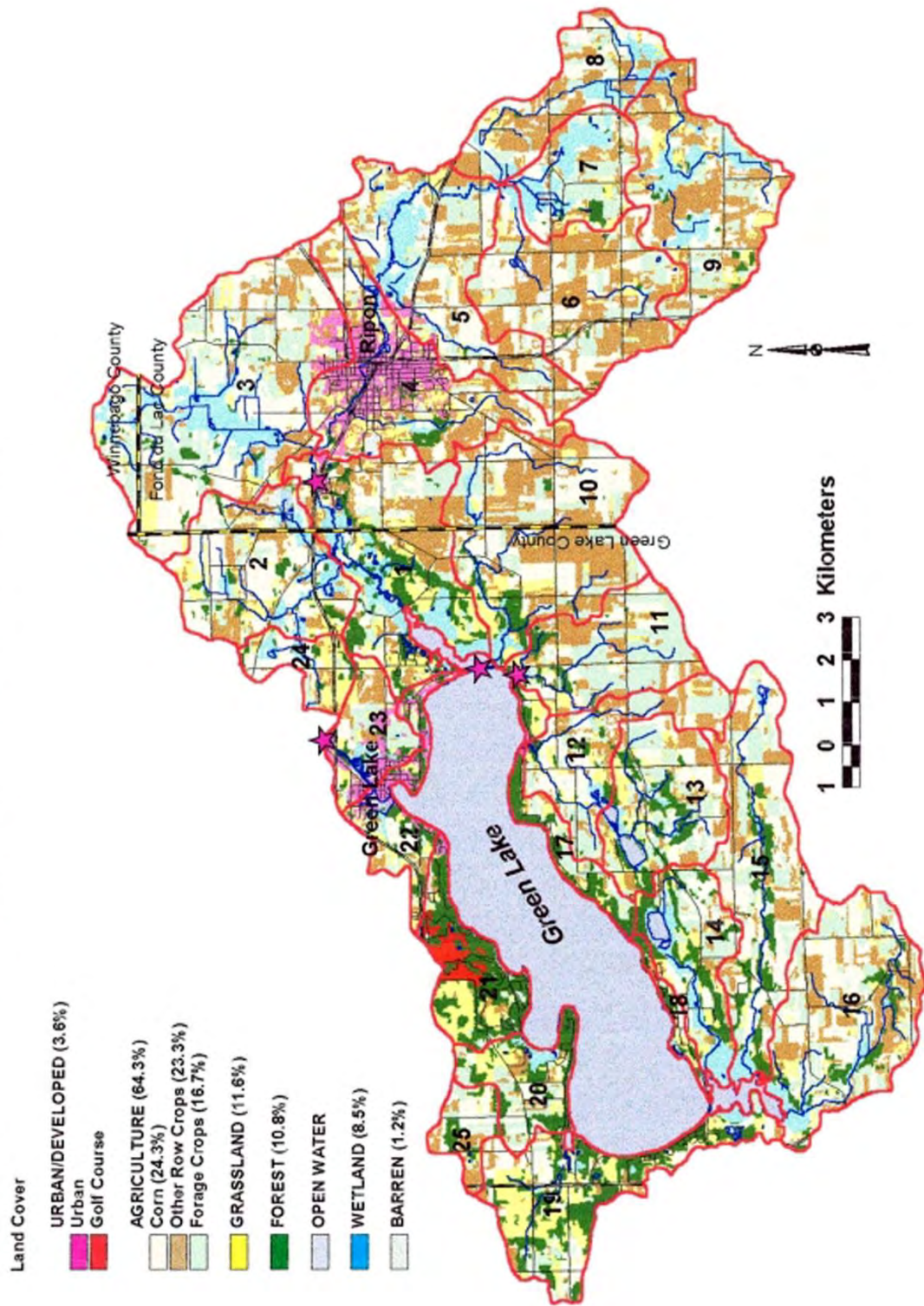
SWAT Model Inputs

GIS layers: The following GIS data layers were used to provide inputs to the SWAT model and to prepare a various GIS-based maps and analyzes:

1. 1:24k WDNR watershed boundaries; subwatersheds were added as part of this project
2. USGS 1:24k Quadrangle Images - digital topo maps (used to delineate subwatersheds)
3. WISCLAND 1992 Land Cover from WDNR
4. NRCS certified digital soil surveys from Fond du Lac and Green Lake counties, and the Winnebago County digital soil surveys (combined into a single watershed coverage)
5. 30 meter digital elevation model (DEM), primarily used to derive overland slope
6. 1:24k surface water hydrology
7. Miscellaneous: roads, county boundaries, etc.

All the GIS coverages and images were obtained from the WDNR, except the county soil surveys. All GIS coverages were projected into WTM-NAD83/91 coordinates. The watershed was divided into 25 subwatersheds by using the DEM and USGS 1:24k digital quadrangle images to directly digitize the boundaries from the computer screen. Through this project, it was determined that subwatersheds 24 and 25 do not appear to drain to Big Green Lake. Subwatershed names and areas are provided later in the report in Table 2.

Figure 1. Big Green Lake Land Cover (based on WISCLAND 1992).



Land cover/use: Land cover within the watershed (Figure 1) was determined from the Level 3 classification of the 1992 WISCLAND land cover image, which was based on LANDSAT Thematic Mapper images. Because of the nature of interpretation and classification, forested urban areas that surround Green Lake were classified as forest in the WISCLAND coverage. The actual proportion of urban area in the entire watershed is therefore somewhat greater than that shown in Figure 1, which was based on the WISCLAND image. A GIS coverage was created to correct for this problem, and this layer was based in part on digitized USGS 1:24k quadrangle images. Another GIS coverage was later obtained from Big Green Lake county which had these types of urban areas delineated around Big Green Lake. This coverage was then used to refine the other, and the proportion of urban areas within the subwatersheds surrounding Big Green Lake were adjusted accordingly. At this time, Figure 1 does not show these additional urban areas surrounding Big Green Lake.

The WISCLAND classified land cover image was used to assign 6 major land covers/uses which were modeled within the watershed: agriculture, urban, golf course, forest, grassland and wetland. These land covers were further divided into 11 "Hydrologic Response Units" which were directly modeled in the following fashion:

- Agriculture - Dairy
 - 1 Conventional tillage practice
 - 2 Mulch-till
 - 3 No-till
- Agriculture - Cash crop
 - 4 Conventional tillage practice
 - 5 Mulch-till
 - 6 No-till
- 7 Urban
- 8 Grassland
- 9 Forest
- 10 Wetland (or Golf Course in a subwatershed that had no significant amount of wetlands)

HRU's basically represent areas within a subwatershed that are similar in a hydrologic or management sense, but are not necessarily contiguous. No one specific farming practice could be used to model the entire watershed; therefore, various proportions of six possible agricultural practices (6 HRU's) were used to simulate what occurred in each subwatershed. For simplicity, every subwatershed was modeled as though it contained 10 HRU's in the order shown above. Since there were 25 subwatersheds, the total number of modeled HRU's was 250. A GIS overlay operation was used to derive the proportional area of the major HRU's within each of the 25 modeled subwatersheds. The next section describes how the agricultural areas were further divided in 6 agricultural HRU's. Where a subwatershed did not contain all of the landuses, the area of the non-existent landuse was assigned a negligible fraction of the total area (0.000001).

Management Practices and Hydrological Response Units (HRU)

SWAT requires detailed information regarding landuse management practices. For example, the type of crop, the date it was planted and harvested, tillage practices and dates, fertilizer

applications and dates, and NRCS curve number for each period, are just some of the information that is input into SWAT's management files. The following discussion describes how these inputs were obtained.

Farm crops: The Level 3 classification of the 1992 WISCLAND classified land cover image has 3 primary crops classified within the Big Green Lake Watershed: "corn", "forage" and "other row crops". For this project, it was assumed that "other row crops" was either soybeans or another fragile crop. Unfortunately, the relatively small size of the delineated subwatersheds made it unreasonable to assume that the proportions of each crop within each subwatershed, as derived from GIS analysis of the 1992 WISCLAND image, were consistent from year to year. In some areas, the field sizes were sufficiently large that a different phase in the crop rotation would have sharply changed the proportion of the crops classified in the WISCLAND image for some of the subwatersheds. In addition, the image of the Green Lake Watershed was actually based on three separately classified scenes, which decreases the reliability of the data, particularly at the detailed Level 3 classification. Therefore, the subwatershed crop percentages derived from the WISCLAND data were not directly used as inputs. Instead, the proportion of dairy and cash cropping in each of the subwatersheds was derived by generalizing the subwatershed-specific data into two agricultural regions within the watershed: (dairy) 50% dairy and 50% cash crop; and (cash crop) 67% cash crop and 33% dairy. This task was accomplished by looking at all the sources of data including: 1) visual inspection of the WISCLAND image; 2) the proportions of each crop within each subwatershed, as indicated by the WISCLAND image; and 3) a watershed inventory conducted by Fond du Lac and Green Lake counties. The cash crop region encompassed subwatersheds 1-11 (except #3), and the remainder of the subwatersheds were assigned to the dairy regions.

Tillage practices: An inventory of farm practices within the Green Lake Watershed was gathered by the Fond du Lac and Green Lake and LCD's, in part, to support the modeling requirements of SWAT. However, the person gathering this information did not transfer this data into a database before leaving their position for different employment. Therefore, the data could only be roughly translated into a spreadsheet information system, but the accuracy of this rough translation was questionable. Therefore, management practice inputs to SWAT were based on generalizations of the collected data, as input to a spreadsheet, augmented by information obtained from the Fond du Lac and Green Lake LCD's. In addition, the Conservation Technology Information Center (CTIC) Conservation Tillage Reports from Green Lake and Fond du Lac Counties were analyzed to determine the primary tillage practice inputs to SWAT. These "Transect Survey" reports were based on statistical sampling procedures of farm fields to determine residue levels present on farm fields shortly after spring planting, as well as other information. The assumptions about current tillage practices that were utilized as data inputs for the management files in the model are summarized below. This data is based solely on the transect survey data for Green Lake County. There were probably too few data points to use the combined watershed data from both counties, particularly since data from hay/alfalfa fields were included in residue/tillage summaries for Fond du Lac County. The details of how these assumptions were derived will be described in the final report.

Summary of farm crop and management assumptions:

Crop practices

Cash crop subwatersheds: 1-11, except #3

2/3 cash crop rotation

1/3 dairy rotation

Dairy subwatersheds: (remaining subs 12-25, and #3)

50% dairy rotation (corn-grain, corn-silage, a, a, a, a)

50% cash crop rotation (corn, soybean)

Primary tillage practices

tillage	corn	soybeans		
conventional practice (CT)	fall moldboard plow	fall chisel plow		
mulch till (MT)	fall chisel plow	spring field cultivator, or disk		
no-till (NT)	none	none		
	CT	MT	NT	
Dairy - present practices	61.0%	36.0%	3.0%	
Cash Crop - present practices	36.0%	46.0%	18.0%	
Combined Present practices	46.2%	41.7%	12.1%	
Alternative A		87.9%	12.1%	
Alternative B		46.2%	53.8%	
Alternative C			100.0%	
Alternative D		update	%	

Nutrients and Nutrient Management: The following assumptions concerning commercial fertilizer and manure applications were utilized as model inputs.

Dairy rotation (cg,cs, oat/a, a,a,a) - options: moldboard plow, chisel or mulch till, no-till

1 corn grain ---- 250 lbs/acre (9-23-30 prior to planting); 30 tons manure in fall after harvest

1 corn silage ---- 250 lbs/acre (9-23-30 prior to planting)

1 oat/alfalfa

3 alfalfa ---- 2nd & 3rd year 18 lbs/acre of 0-10-60 each year; after 4th year, apply 30 t/acre manure in fall (it was assumed that only 10% of farmers apply 180 lbs/acre of 0-10-60; hence, the rate of 18 lbs/acre of fertilizer applied)

30 ton/acre/yr of dairy manure is applied for two years of this rotation (total 60 tons in 6 years)

Cash crop rotation (corn, soybeans) - options: moldboard plow, chisel mulch till, no-till
1 year corn 125 lbs/acre Anhydrous ammonia prior to planting; 280 lbs of 9-23-30 @planting

1 year soybean (soybeans serve as the legume crop or fragile crop in the cash crop rotation)
200 lbs/acre 9-23-30 @ planting (this could instead be applied during corn season; at this time the model did not show a difference; note that the nitrogen was not necessary of soybeans)

Nutrient management was not modeled at this time.

Climatological inputs: Precipitation data from Ripon, temperature data from Fond du Lac and general weather statistics from Portage were used for climatological inputs to SWAT.

Soils and overland slopes: County soil surveys were processed and combined into a single GIS coverage, and projected into WTM-NAD83 coordinates. This coverage was intersected/combined with both the WISCLAND land cover image (used to delineate HRU's), and the subwatershed delineated GIS layer, to produce soils information that was specific to each of the HRU's within each of the 25 subwatersheds. For each HRU within a subwatershed, and for each of the soil parameters required by SWAT, an area-weighted average value was assigned based on the area of each of the soil series within that HRU. A somewhat simpler procedure was conducted with the 75 meter digital elevation map to produce average slopes for each of the HRU's, within each of the 25 subwatersheds.

HRU-specific information was deemed important because slopes and soils often vary between different landuses. For example, compared to agricultural land (average slope = 3.6%), the average slope of forested land was approximately twice as steep (7.1%), and the average slope of grassland was 4.5%. This level of analysis is even more critical where there is a large proportion of wetlands, for which the average slope would substantially reduce the slope of the other HRU's in the subwatershed (unless other procedures are taken to not include the slopes from wetland areas). However, for other critical model parameters, this procedure was not as important because of the relatively homogeneous nature of the soils in the watershed. For example, the hydrologic soil group (A, B, C, D), which helps determine the NRCS curve number, did not vary much throughout the watershed, nor between the different HRU's (excluding wetlands). This was not the case when this same procedure was used in to determine subwatershed-specific soil parameters in the Lower Fox River Basin.

As previously mentioned, hydrologic soil groups varied little throughout the watershed, and therefore curve numbers also varied little. There are only four categories for hydrologic group, so this outcome is not unexpected. However, saturated conductivity did vary substantially; thereby, indicating that surface runoff and recharge proportions are unlikely to be the same throughout the watershed. Therefore, saturated conductivity was used to differentiate those subwatersheds whose soils were much more permeable than the rest. Essentially, this procedure was akin to assigning a "low B" soil hydrologic group to subwatersheds whose soils were much more permeable than others. Curve numbers for subwatersheds 18-21 were reduced by 3 units, while the curve numbers of 22-23 were reduced by 1.5 units, to reflect the reduced runoff potential expected in areas which had soils that were much more permeable than in other subwatersheds.

Stream characteristics: Geomorphic relationships between drainage area and stream channel characteristics were used to calculate both the main and routing channel depths and widths at different locations in the watershed.

Stream flows and loads: For purposes of calibrating and validating the SWAT model, stream flow, and phosphorus and total suspended solids (TSS) loads were obtained from USGS for the following locations in the watershed: Silver Creek at Koro Rd. (USGS # 040734644; 1987-96), Silver Creek at Big Green Lake inlet (USGS # 04073468; 1987-98), and White Creek at Spring Grove Road (USGS # 04030201; 1982-88; 1997-98). These monitoring sites were jointly funded by USGS, WDNr and the Green Lake Sanitary District. The monitoring locations are shown in Figure 1. Observed data from the Silver Creek-Koro Rd. station, for the 1987-92 period, was chosen to calibrate the model. Observed data from the remaining years (1993-96) at this site, and data from the Silver Creek at Big Green Lake Inlet (1987-98) were used to assess the validity of the model (validation period).

Unfortunately, the observed data from the White Creek station could not be directly utilized for calibration or validation for two reasons: (1) the annual stream flows were unusually high given the drainage area of this subwatershed; and (2) there appears to have been a substantial change in water quality between the 1982-88 period, and the 1997-98 period. Whereas the long-term stream flow on an areal basis for Silver Creek was about 250 mm (annualized over 1987-96 period), the measured flows at White Creek were much higher than could reasonably be expected (Table 1) given the amount of measured precipitation, and assumed evapotranspiration. There are at least two possible explanations for this disparity. First, the surface water drainage area of the White Creek subwatershed may be much greater than the areas delineated by either the USGS or Fox-Wolf Basin 2000. Second, the groundwater drainage area may be substantially greater than the surface water drainage areas delineated by either the USGS or Fox-Wolf Basin 2000. Further review of the 1:24,000 topological maps indicated that the eastern drainage area divide is not clearly defined, and there are also many springs in the White Creek subwatershed. In addition, road ditches may also cross the natural drainage divide at an elevation sufficiently low that large amounts of water are transferred to the White Creek subwatershed from adjacent subwatersheds. Further analysis of the White Creek data also showed that substantial changes may have occurred as result of efforts to reduce stream bank and possible gully erosion. This inference was drawn because phosphorus to TSS ratios have risen markedly, in association with what appears to be a sharp drop in TSS loads (instantaneous TSS concentrations of over 50,000 mg/L were recorded in the 1982-88 period). For example, the volumetric concentration of TSS in 1982 (calendar year) was 905 mg/L compared to a concentration of 47 mg/L in 1998 (water year), even though the total annual flows were nearly identical (1,250 cfs). In addition, the average ratio of phosphorus to TSS in the 1982-88 period was 0.93 lbs of phosphorus per ton TSS, compared to the more recent average in the 1997-98 period of 4.0 lbs of phosphorus per ton TSS.

Given the unusual water budget, and the potential temporal change in water quality over the monitoring period, it was determined that it would not be reasonable to calibrate the SWAT model with observed data from the White Creek monitoring location. This determination was particularly unfortunate because it precluded calibrating the SWAT model in two separate phases, as originally intended: (1) calibrate the subwatershed component of the model first by using a simple approach which involved modeling the White Creek subwatershed as a single subwatershed, without any routing required; and (2) then calibrating the routing component of the SWAT model by modeling the larger Silver Creek watershed, which is composed of many subwatersheds, while adjusting only those parameters that affect routing. By separating the subwatershed load-generating routines from the routing component, a more robust and predictive model may have been developed for the Big Green Lake watershed.

Table 1. Annual flows at White Creek station (areal basis) and precipitation at Ripon.

Year	Precip. mm	Flow mm	Flow % of Precip.
1982*	790	347	44%
1983	826	593	72%
1984	893	527	59%
1985	885	553	62%
1986	998	777	78%
1987	651	192	30%
1988*	574	137	24%

* all are full calendar years, except 1982 and 1988 are incomplete years

SWAT Model Setup - summary

A total of 25 subwatersheds were delineated for this project (# 24 and # 25 do not drain to Green Lake). Each of the subwatersheds contained 10 hydrologic response units (HRU's), which were based on these primary land uses: agriculture (6 HRU's), urban, grassland, forest, and wetland (or golf course). HRU's basically represent areas within a subwatershed that are similar, but are not necessarily contiguous. No one specific farming practice could be used to model the entire watershed; therefore, various proportions of six possible agricultural practices (6 HRU's) were used to simulate what occurred in each subwatershed.

The agricultural HRU's consisted of two potential farming practices:

- 1) Dairy-based (6 year rotation: corn-grain, corn silage, oats/alfalfa, alfalfa, alfalfa, alfalfa)
- 2) Cash crop (2 year rotation: corn, soybeans).

Under each of the two potential farming practices, three tillage practices were simulated: a) conventional tillage with fall moldboard plow as the primary tillage implement for corn and fall chisel plow for soybeans; b) mulch till, or chisel plow tillage in fall; and c) no-till. Hence, a total of six HRU's were used to represent agricultural areas.

SWAT98.2 was tested and modified to suit conditions in Wisconsin. Many of these code modifications were to bring in prior modifications that we made with SWAT97.2. In our testing of SWAT98.2, we have found some additional errors that required fixing. These errors have now been fixed by Fox-Wolf Basin 2000 or the model developers at USDA-ARS in Temple, Texas.

Model Calibration and Assessment

Flow Calibration: The Priestly-Taylor evapotranspiration equation was utilized for this project. The following coefficients were added to the model code which allowed adjustment of the simulated water balance to obtain a reasonable fit with the observed stream flows: Priestley-Taylor ET equation (0.77), NRCS curve number input (0.99), and available water capacity soils input (0.92).

TSS calibration: Parameters in the modified universal soil loss equation (MUSLE) were adjusted to obtain a reasonable fit between observed and simulated TSS loads. MUSLE is shown in Equation 1.

$$\text{MUSLE: } Y = a (Q)^b (q_p)^c (DA)^d [(K) (C) (PE) (LS)] \quad (\text{Eq. 1})$$

where:

Y	=	sediment yield in metric tons/ha (Mg/ha)
Q	=	surface runoff volume in mm
q_p	=	peak flow rate in mm/hr
DA	=	drainage area in hectares
K	=	soil erosion factor
C	=	crop management factor
LS	=	slope-length and slope-steepness factor
PE	=	erosion control practice factor
a,b,c,d	=	constants normally set at a = 1.586, b & c = 0.56, d = 0.12 (user-specified values can be used where there are sufficient data for calibration)

The following values were utilized in the MUSLE equation for this project: a = 0.01, b = 1.6, c = 0.0, and d = 0.0.

Annual flow and loads: Simulated and observed annual flows (mm, on an areal basis), TSS loads (metric ton) and phosphorus loads (kg) are compared in Table 2 and Figures 2a, 2b and 2c respectively, for the monitoring site located on Silver Creek at Koro Road (drainage area of 94.7 km²). The annual totals correspond to USGS water years (October 1 to Sept 30). Annual precipitation (mm) is shown on the second y-axis in Figure 2, so precipitation and stream flow can be compared. Despite wide fluctuations in annual precipitation, observed and simulated annual values for TSS and phosphorus loads as well as annual water balance all coincide fairly well at Silver Creek during the calibration period (1987-92). However, during the validation, or assessment period (1993-96), TSS and phosphorus loads were substantially under predicted for 1993, and somewhat under predicted for 1994. Exceptionally high runoff occurred in 1993. The observed annual flow of 560 mm in 1993 represented 63% of the precipitation measured at Ripon during that year (890 mm); plus, the flow in 1993 was more than two times greater than the second highest annual flow which occurred during the 1987-96 period. Evapotranspiration must have been greatly suppressed to produce such a high flow to precipitation ratio. Spring planting was delayed during 1993, and high soil moisture levels further delayed plant emergence and growth early in the year. As a result, evapotranspiration and protective plant canopy from annual crops were lower than normal during the early growing season. Plant stress due to wet soil conditions (including reduced availability of nitrogen) is not simulated by the model, and only the average planting dates were input to the model. If these factors had been accounted for by model simulations, both of these factors would have increased simulated flows and loads, and produced a closer correspondence between observed and simulated values. Only average planting/tillage dates were input to the model to reduce the number of input files and simplify the model set up.

Silver Creek at Green Lake inlet: The model was not calibrated with data from this site, so the entire 1987-98 period could be considered a validation/assessment period; however, since this site is downstream of the primary calibration site, it may be more appropriate to consider separating the data into two periods (1987-92: calibration period; 1993-98 validation period). Simulated and observed annual flows (mm), TSS loads (metric ton) and phosphorus loads (kg) are compared in Table 3 and Figures 3a, 3b and 3c respectively, for the monitoring site located on Silver Creek

Table 2. Observed and simulated annual flow and loads - Silver Creek at Koro Rd.
Calibration period (1989-92) and Validation period (1993-96).

	Ripon Precip. (mm)	Flow (mm)		TSS (MT)		Total Phosphorus (kg)	
		Observed	SWAT	Observed	SWAT	Observed	SWAT
1987*	640	150	160	280	380	2,640	3,480
1988	570	120	90	330	170	2,780	2,280
1989	680	200	200	1,530	1,620	9,810	10,030
1990	1,000	250	310	700	910	7,480	6,360
1991	790	240	250	780	540	3,990	4,550
1992	750	250	220	540	480	3,380	4,020
1993	890	560	440	1,530	960	10,640	6,760
1994	730	170	110	480	230	4,650	2,670
1995	840	240	220	480	630	3,910	4,810
1996*	540	250	260	510	710	3,980	4,810
Totals							
1987-92	4,440	1,210	1,220	4,150	4,110	30,070	30,710
1993-96	3,000	1,230	1,030	3,000	2,530	23,180	19,060
1987-96	7,440	2,440	2,260	7,150	6,640	53,260	49,770

* values are totals for calendar years, except 1987 (started Feb 1), and 1996 (ended Sept. 30)

Table 3. Observed and simulated annual flow and loads - Silver Cr. at Green Lake Inlet.
Calibration period (1989-92) and Validation period (1993-98).

	Ripon Precip. (mm)	Flow (mm)		TSS (MT)		Total Phosphorus (kg)	
		Observed	SWAT	Observed	SWAT	Observed	SWAT
1987*	640	160	150	520	410	2,610	3,770
1988	570	130	80	320	170	2,700	2,310
1989	680	220	200	1,790	2,170	10,570	11,200
1990	1,000	270	310	1,420	1,150	8,270	6,800
1991	790	230	250	500	610	4,160	4,980
1992	750	230	210	750	460	3,470	4,200
1993	890	560	440	3,290	1,160	12,580	7,390
1994	730	190	110	540	260	6,450	2,720
1995	840	260	210	720	640	4,370	5,150
1996	660	300	280	790	920	5,100	5,650
1997	710	240	160	740	390	4,230	3,580
1998*	620	190	150	870	610	3,230	4,240
Totals							
1987-92	4,440	1,260	1,200	5,300	4,980	31,800	33,260
1993-98	4,460	1,740	1,350	6,950	3,990	35,970	28,720
1987-98	8,900	3,000	2,550	12,260	8,970	67,760	61,990

* values are totals for calendar years, except 1987 (started Feb 1), and 1998 (ended Sept. 30)

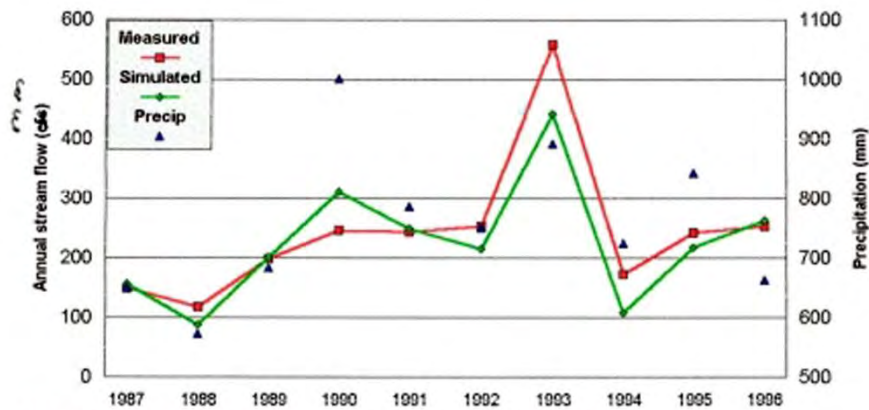


Figure 2a. Observed and simulated annual stream flow - Silver Cr. at Koro Rd.

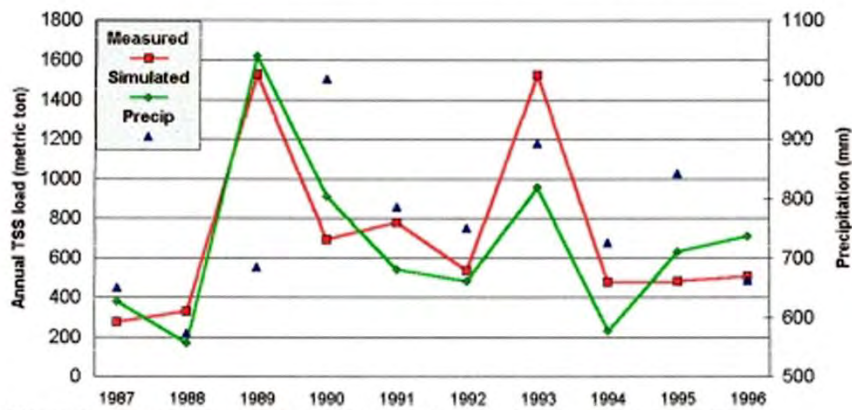


Figure 2b. Observed and simulated annual TSS load - Silver Cr. at Koro Rd.

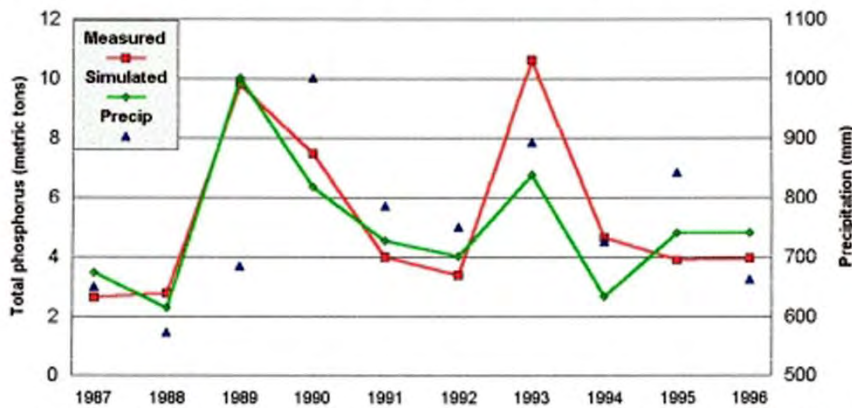


Figure 2c. Observed and simulated annual total phosphorus load - Silver Cr. at Koro Rd.

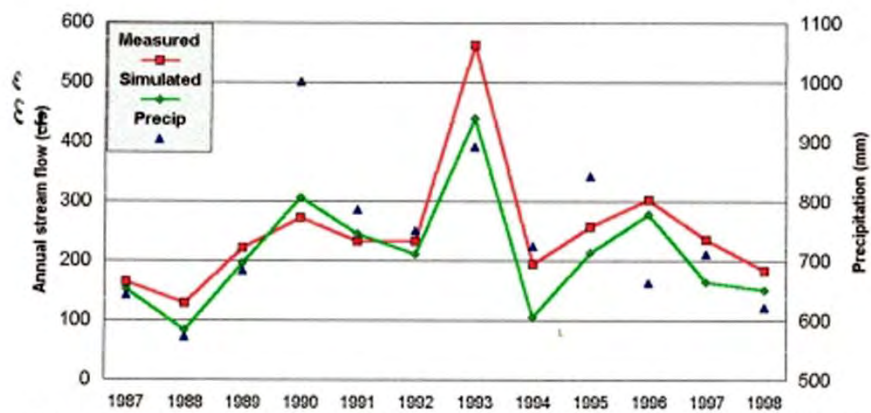


Figure 3a. Observed and simulated annual stream flow - Silver Cr. at Green Lake Inlet.

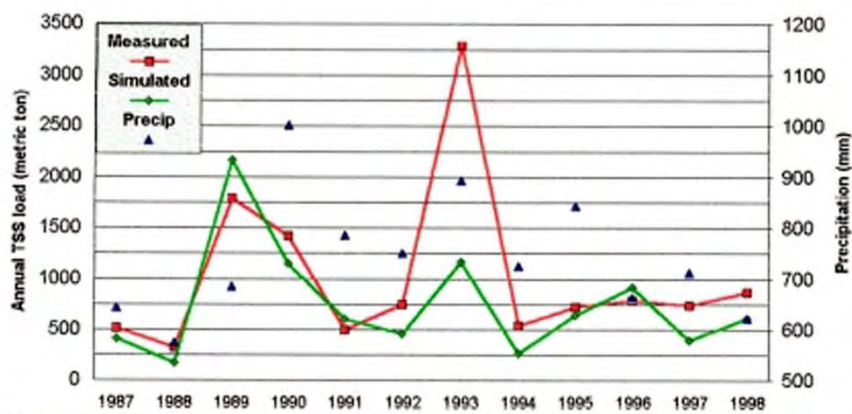


Figure 3b. Observed and simulated annual TSS load - Silver Cr. at Green Lake Inlet.

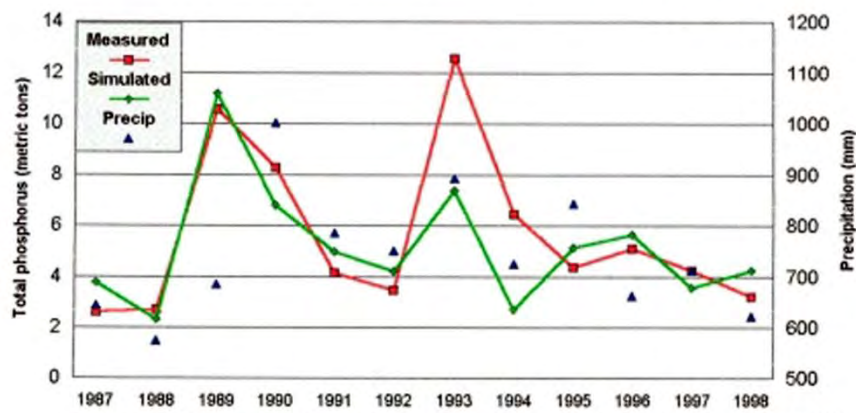


Figure 3c. Observed and simulated annual total phosphorus load - Silver Cr. at Green Lake Inlet.

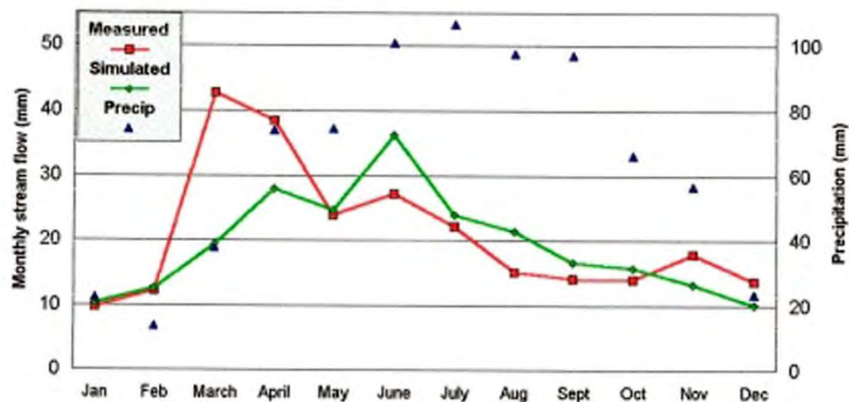


Figure 4a. Observed and simulated monthly stream flow - Silver Cr. at Koro Rd.

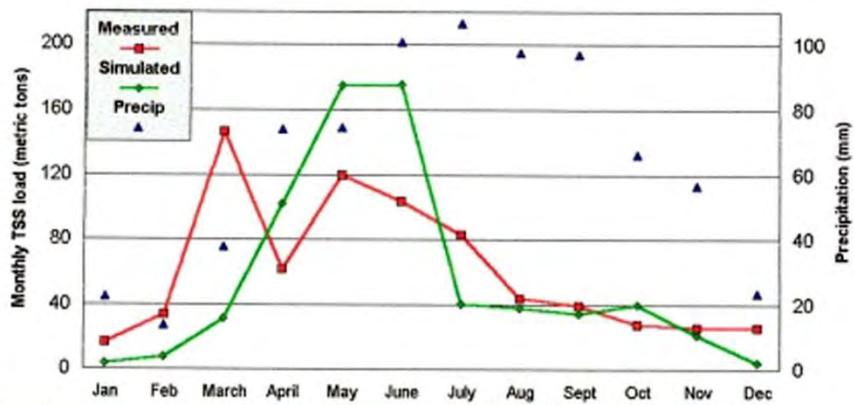


Figure 4b. Observed and simulated monthly TSS load - Silver Cr. at Koro Rd.

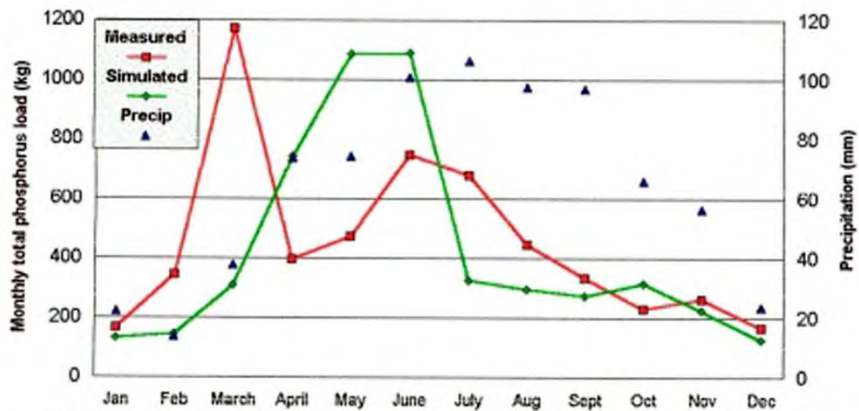


Figure 4c. Observed and simulated monthly total phosphorus load - Silver Cr. at Koro Rd.

at the Green Lake inlet (drainage area of 122 km²). Again, despite wide fluctuations in annual precipitation, observed and simulated annual values for TSS and phosphorus loads as well as annual water balance all coincide fairly well during the entire 1987-98 period, except during 1993 and 1994. Phosphorus loads were substantially under predicted in 1993 and 1994. However, TSS loads were greatly under predicted in 1993 when the simulated load was about one third of the observed load; the difference is much greater than it was for the upstream monitoring site. One potential explanation that deserves further analysis is that the model, as currently set up, is settling too much TSS within some or all of the stream reaches. Unaccounted channel degradation occurring during the high flows in 1993 could also have been responsible for the high TSS loads in 1993. It appears unlikely that delayed planting and slow early crop growth during 1993 will be sufficient to account for the discrepancy between simulated and observed TSS loads at this location. I recommend that any further refinement of the model should utilize the 1993 data for calibration purposes at this site.

Monthly flow and loads: Simulated and observed monthly average stream flow (mm), TSS loads (metric ton) and phosphorus loads (kg) are compared in Figures 4a, 4b and 4c respectively, for the monitoring site located on Silver Creek at Koro Road. The monthly values were averaged over the 1987-96 period. Monthly precipitation (mm) is shown on the second y-axis in Figure 4, so precipitation and stream flow can be compared.

In general, monthly average simulated flows were close to observed flows except during March, when the simulated average flow was less than half the observed average, while flow during April was somewhat lower than the observed average value (Figure 4a). Except for these excursions, the model was able to reliably track seasonal changes in flow despite the less than direct correspondence between observed monthly flow and precipitation. It is understandable that the SWAT-estimated March flow (19.4 mm on an areal basis) did not match the observed flow because the average March precipitation at Ripon is 37.8 mm, which is lower than the observed flow of 42.7 mm (1987-96 averages). A number of conditions could contribute to the high flows observed in March including: a large groundwater storage component, high proportion of runoff due to frozen ground conditions, frozen surface water and snow storage in wetlands that thaws in spring, snow melt, and delayed groundwater flow over winter. In addition, some of the measured discharges in March may have been affected by ice conditions which would tend to overstate flow whenever the ice caused the water to backup. Therefore, while attempts may be made to refine the model and improve the fit between observed and simulated flows, it must be recognized that many of the conditions that cause the average March flow in Silver Creek to exceed the average March precipitation are likely to be difficult to model with SWAT.

Compared to stream flow, simulated TSS and phosphorus loads were much lower than the observed values during the month of March, suggesting that surface water contributions were also understated by the model simulations (Figure 4b, 4c). If only stream flow had been under-predicted by the simulations, while the loads had been closely estimated, it would have indicated that the groundwater contributions were not well predicted. Simulated TSS and phosphorus loads were substantially greater than observed loads during the months of April, May and June (Figure 4b, 4c). Phosphorus loads were particularly overstated. The precise cause of these over-predictions is not yet known, but further attempts may be made to refine the model and produce a better fit in the near future. Given these results, I am forced to conclude that the model as currently set up, does only a fair job of accurately representing the observed TSS and phosphorus loads during the months of April through July; while the model performs poorly in simulating loads during March. This conclusion is especially applicable to simulated phosphorus loads.

However, it is also important to note that the total simulated TSS and phosphorus loads during the March through July period were essentially identical to the observed values.

White Creek at Spring Grove Road: For reasons previously stated, TSS and phosphorus loads from the White Creek monitoring station could not be directly utilized for calibration purposes, particularly for the 1982-88 period. However, simulated and observed TSS and phosphorus loads from the 1997-98 period were compared to see how well the model could perform, given the caveat about the high observed stream flows from this drainage area. The average simulated flow for the 1997-98 calendar year period was about half of the observed average during the 1997-98 water year. The simulated average TSS load during the 1997-98 calendar year period was 102 metric tons compared to the observed average TSS load of 230 metric tons during the 1997-98 water year period. However, the simulated average phosphorus load (1997-98 calendar year) was within 10% of the observed average (1997-98 water year).

Evaluation of Alternative Management Practices

Four alternative agricultural management practices were simulated to evaluate the impact of each alternative on TSS and phosphorus loads, as routed to the outlet of each subwatershed, and routed to Big Green Lake. The four alternative management scenarios include: (A) those cropped areas practicing conventional tillage (corn - fall moldboard plow; soybeans - fall chisel plow) switched to mulch-till (corn - fall chisel plow; soybeans - field cultivator or disk in spring); (B) conventional tilled acres switched to mulch-till, plus all cropped areas practicing mulch-till switched to no-till; (C) no-till was practiced on all cropped agricultural land; and (D) no-till was practiced on all cash-crop farms, but dairy farms only switched to mulch-till, unless they were already practicing no-till. The cost to switch to mulch-till was assumed to be \$30/acre for cash-crop rotations, and \$15/acre for dairy rotations. The cost was lower for dairy rotations because the alfalfa acreage was only affected prior to planting or after the last harvest. To switch to no-till, the cost was assumed to be \$50/acre for cash-crop rotations, and \$25/acre for dairy rotations. These costs were used as examples, and different costs can be substituted in the provided spreadsheet.

Simulated subwatershed TSS and phosphorus loads for the current condition and four alternative management scenarios are shown in Tables 4 and 5, respectively (1987-98 average). The loads for subwatershed #4 do not include the discharge of the Ripon Wastewater Treatment Plant. Phosphorus loads as routed to Big Green Lake are also shown in Table 5, which includes the discharge from the Ripon Wastewater Treatment Plant. The simulated 1987-98 average annual phosphorus load from Silver Creek to Green Lake is 5,200 kg (11,500 lbs), and the total load routed to the lake is 13,700 kg (30,200 lbs).¹ The simulated 1987-98 average annual TSS load from Silver Creek to Green Lake is 750 metric tons (830 Eng. tons), and the total load routed to the lake 2,600 metric tons (2,900 Eng. tons). The simulated TSS loads are probably less reliable than phosphorus loads because of the difficulty in modeling sediment loads that are caused from stream bank erosion or severe gully erosion, which can be specific to individual subwatersheds. In addition, phosphorus is more conservative than TSS as it is routed through stream reaches, so the

¹The total simulated load to the lake is somewhat less than this total because the upper portion of the Hill Creek watershed was not routed through the lower portion of the Twin Lakes system. Subwatershed #12 should be routed through the lake system and through subwatershed #13, but this was not done at this time to simplify the modeling process. Therefore, the actual load from sub. #12 to Green Lake should be less than indicated in this report; however, the impact of sub. #12 on the Twin Lakes system is also important.

simulated TSS loads are more sensitive to possible errors in the routing process.

The subwatersheds were also ranked and sorted on the basis of total reduced phosphorus load, as routed to Big Green Lake. Based on this ranking, the phosphorus loads, reduced loads, associated cost to reduce the load, and the unit cost to reduce the loads are shown in Table 6. These rankings show how the simulated results can be used to determine where the greatest reductions might best be targeted.

The simulated phosphorus loads for each of the four alternative scenarios, as routed to Big Green Lake, are combined with the associated cumulative cost to reduce the load in Figure 5. Each dot in the figure represents the impact of adding another subwatershed to each of the alternative management scenarios (going from left to right). Thus, if the goal is to achieve an average annual total phosphorus load to Big Green Lake of 20,000 lbs (reduce load by 33%), then Alt. C and D have nearly the same cost of \$600,000/year, Alt. B will require implementation in all subwatersheds and cost \$800,000/year, and Alt. A will only be able to reach a load of 24,500 lbs. The simulated data in Figure 5 suggests that the different alternative scenarios do not differ much in cost-effectiveness until the 10 least important subwatersheds undergo the management change, at which point the lines tend to diverge. The information shown in Figure 5 is summarized in Table 7.

Model limitations

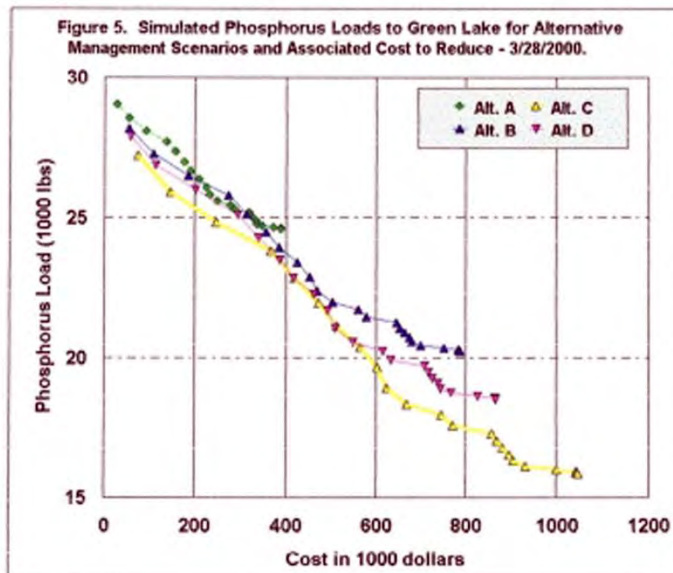
This section describes some of the limitations inherent to the simulations. These limitations should be considered when evaluating absolute and relative loads, as well as the costs to reduce these loads, within the Big Green Lake Watershed.

Until very recently, the model dealt with inorganic phosphorus fertilizer incorrectly; that is, the addition of inorganic fertilizer had no effect on the levels of phosphorus supposedly attached to sediment during erosion events. While this component of the model has been fixed, there has been insufficient time to thoroughly calibrate the response of the model to various inputs related to phosphorus. For example, the relative proportions of soluble, sediment-attached and organic phosphorus need to be adjusted to reflect results from published data for a variety of management practices. One reason the proportion of soluble phosphorus is important is that the model routes 100% of this form of phosphorus to the watershed outlets. Therefore, a pound of soluble phosphorus discharged from either an upper or a lower reach are both treated as though all of it reaches the lake.

The version of the SWAT model used in this project does not route sediment and sediment-attached phosphorus in the same manner, and this aspect of the model resulted in some undesired outcomes. In some stream reaches, sediment-attached phosphorus is settling out in the stream at a higher rate than the sediment, which is not appropriate. In addition, within every stream reach, the non-soluble portion of the phosphorus load is being trapped at a very consistent rate of about 33%. That is, on a long-term net basis, 67% of the non-soluble phosphorus that enters a stream reach passes through the reach, while 33% remains as deposited material. This is true of every reach, which seems unreasonable since there are major differences in the stream gradients. These odd results may be due to my not knowing, until recently, that the model was using a new, much more complex routine for routing phosphorus. Thus, the model results are based on the default nutrient routing parameters, without any changes to adjust these values to give better results. Without further refinement, it must be understood that the modeled results do not currently mimic the

physical world in a realistic manner, but the results are still useful for comparison purposes. If the model is showing more deposition of phosphorus when routing through Silver Creek than actually occurs, then the relative load from Silver Creek is greater than indicated by the model (compared to the rest of the watershed). The same logic holds if the relative amount of soluble phosphorus is too low, because routing through the reaches of Silver Creek would have little effect on this form of phosphorus.

For a number of reasons, the predicted load reductions associated with the alternative management scenarios are probably overly optimistic. Phosphorus reductions resulting from conservation tillage may be too high because I used management inputs which distributed manure at a deeper level than actually occurs. In addition, the amount of soluble phosphorus simulated by the model with the current settings may be too low. Soluble phosphorus is more difficult to control with conservation tillage, which cannot only increase the concentration of soluble phosphorus, but actually increase the load as well. Time did not permit testing these aspects of the model thoroughly enough to be confident that the inputs that affect manure depth or the proportion of soluble P are reasonable. Another modeling assumption that may lower the reduction potential, is the proportion of land that is assumed to be under conventional tillage. These numbers were based on the county transect survey, in which an estimate was made of the residue on the fields just after planting. If it was estimated that the residue percentage directly after planting was lower than 15% for a particular field, then "conventional tillage" was assigned to that field, even though substantial protective residue may have been present between the fall harvest and spring planting period. If the prior crop was soybeans, very little residue might remain for detection after planting the next crop because of spring tillage, even if the residue was undisturbed until the soil was tilled. Yet undisturbed residue should substantially reduce TSS and phosphorus loads until spring tillage occurs. This aspect of erosion control is important because approximately 30% of the TSS load and 42% of the phosphorus load measured at the Silver Creek-Koro Road monitoring station (1987-96) occurs between the period between typical fall harvest and spring tillage.



Cost Assumptions

	Dairy	Cash Crop
Alt. A	CT > MT	CT > MT
Alt. B	CT > MT MT > NT	CT > MT MT > NT
Alt. C	all NT	all NT
Alt. D	all MT	all NT

MT cost	\$15/acre	\$30/acre
NT cost	\$25/acre	\$50/acre

Table 7. Cumulative cost to reduce phosphorus loads, and associated load reduction to Big Green Lake.

Cumulative Cost to reduce Phosphorus				Cumulative total Phosphorus Reduction				sub#	subwatershed
Alt. A	Alt. B	Alt. C	Alt. D	Alt. A	Alt. B	Alt. C	Alt. D		
\$28,300	\$55,600	\$74,500	\$58,100	1,110	2,000	2,940	2,270	1	16 Wurchs Creek
\$55,800	\$109,400	\$146,800	\$114,400	1,620	2,900	4,250	3,300	2	15 Roy Creek
\$82,600	\$165,900	\$247,700	\$201,700	2,000	3,610	5,270	4,190	3	10 Dakin Creek
\$137,800	\$274,100	\$365,800	\$294,000	2,450	4,390	6,370	5,060	4	3 Silver Cr. 3 (routed, est.)
\$156,600	\$313,400	\$417,800	\$338,800	2,820	5,050	7,330	5,900	5	1 Silver Cr. 1
\$178,700	\$355,100	\$472,900	\$386,400	3,170	5,690	8,230	6,680	6	2 Silver Cr. 2 (routed, est.)
\$191,700	\$384,600	\$512,400	\$417,300	3,490	6,250	9,040	7,320	7	13 Hill Creek Twin Lake
\$210,300	\$423,100	\$563,300	\$461,200	3,770	6,760	9,770	7,950	8	11 White Creek
\$225,200	\$452,300	\$602,500	\$491,800	4,060	7,270	10,510	8,540	9	12 Hill Creek Lower
\$233,200	\$468,200	\$623,700	\$508,400	4,350	7,780	11,260	9,120	10	14 Spring Creek
\$249,700	\$502,400	\$668,900	\$547,400	4,580	8,190	11,840	9,620	11	4 (b) Silver Cr. 4 (routed, est.)
\$277,700	\$560,400	\$745,600	\$613,700	4,730	8,450	12,200	9,940	12	5 Silver Cr. 5 (routed, est.)
\$287,100	\$576,800	\$770,200	\$632,900	4,880	8,720	12,590	10,250	13	19 Beyers Cove
\$319,200	\$645,300	\$858,100	\$708,800	4,990	8,910	12,850	10,480	14	6 Silver Cr. 6 (routed, est.)
\$322,900	\$652,500	\$867,800	\$716,300	5,100	9,110	13,140	10,710	15	22 Green Lake North (f)
\$328,000	\$662,600	\$881,300	\$726,800	5,200	9,280	13,390	10,900	16	23 Green Lake North (f)
\$333,400	\$673,300	\$895,600	\$738,000	5,290	9,440	13,620	11,080	17	20 Norwegian Bay
\$336,400	\$679,000	\$903,300	\$744,000	5,380	9,600	13,860	11,270	18	17 South Shore
\$346,300	\$699,700	\$930,600	\$767,600	5,460	9,730	14,040	11,420	19	7 Silver Cr. 7 (routed, est.)
\$371,500	\$751,800	\$999,500	\$827,100	5,520	9,830	14,170	11,550	20	9 Silver Cr. 9 (routed, est.)
\$387,100	\$784,200	\$1,042,200	\$864,000	5,540	9,890	14,230	11,600	21	8 Silver Cr. 8 (routed, est.)
\$387,400	\$784,800	\$1,043,000	\$864,600	5,560	9,910	14,280	11,640	22	18 SW Shore
\$388,400	\$786,700	\$1,045,600	\$866,600	5,580	9,930	14,320	11,670	23	21 Pigeon Cove/Malcolm E

Table 5. Big Green Lake watershed simulated average annual phosphorus loads and yields (1987-98) - English units.
(as simulated with modified SWAT8 model: 3/28/2008)

		Current Practices CT, MT & NT(a)			Alternative A CT > MT			Alternative B CT > MT, MT > NT			Alternative C all NT (no-08)			Alternative D Dairy MT; Cash Crop (no-08)			Total Cost to reduce			Ail. A			Ail. B			Ail. C			Ail. D			Cost per lb. Total Phos. reduced								
		Area			Area			Area			Area			Area			Ag area			Ag area			Ag area			Ag area			Ag area			Ag area			Ag area					
Sub-id	Subwatershed Name	sq. mi.	lb/acre	percent reduced	sq. mi.	lb/acre	percent reduced	sq. mi.	lb/acre	percent reduced	sq. mi.	lb/acre	percent reduced	sq. mi.	lb/acre	percent reduced	sq. mi.	lb/acre	percent reduced	sq. mi.	lb/acre	percent reduced	sq. mi.	lb/acre	percent reduced	sq. mi.	lb/acre	percent reduced	sq. mi.	lb/acre	percent reduced	sq. mi.	lb/acre	percent reduced						
14(b)	Silver Cr. 1	8.01	1,840	0.481	1,840	1,840	0.000	1,178	306	36.4%	891	232	51.6%	1,016	264	45.0%	48.0%	\$18,800	\$39,200	\$51,900	\$44,800	\$52	586	\$54	\$54	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1					
	Silver Cr. 2	4.48	2,357	0.800	2,357	2,357	0.000	1,383	2,357	53.9%	1,383	2,357	53.9%	1,383	2,357	53.9%	75.3%	\$45,000	\$88,800	\$118,200	\$93,300	\$72	179	\$73	\$73	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1					
	Silver Cr. 3	9.35	3,019	0.504	3,019	3,019	0.000	1,505	3,019	32.5%	1,505	3,019	32.5%	1,505	3,019	32.5%	75.3%	\$45,000	\$88,800	\$118,200	\$93,300	\$44	338	\$45	\$45	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1					
	Silver Cr. 4	5.39	2,410	0.717	2,410	2,410	0.000	1,644	2,410	43.6%	1,644	2,410	43.6%	1,644	2,410	43.6%	81.7%	\$16,500	\$34,200	\$45,200	\$38,500	\$44	338	\$45	\$45	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1					
	Silver Cr. 5	5.70	1,481	0.409	1,481	1,481	0.000	1,086	1,481	34.4%	1,086	1,481	34.4%	1,086	1,481	34.4%	74.9%	\$16,500	\$34,200	\$45,200	\$38,500	\$44	338	\$45	\$45	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1					
	Silver Cr. 6	5.33	1,522	0.448	1,522	1,522	0.000	1,037	1,522	37.9%	1,037	1,522	37.9%	1,037	1,522	37.9%	81.7%	\$32,100	\$66,500	\$87,900	\$75,000	\$104	511	\$107	\$107	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1					
	Silver Cr. 7	2.45	968	0.617	968	968	0.000	568	968	41.3%	568	968	41.3%	568	968	41.3%	81.7%	\$10,000	\$20,700	\$27,300	\$23,600	\$47	562	\$47	\$47	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1					
	Silver Cr. 8	3.10	378	0.190	378	378	0.000	218	378	57.9%	218	378	57.9%	218	378	57.9%	71.9%	\$25,200	\$52,700	\$68,800	\$58,000	\$105	518	\$108	\$108	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1			
	Silver Cr. 9	5.22	1,162	0.348	1,162	1,162	0.000	530	1,162	38.2%	530	1,162	38.2%	530	1,162	38.2%	72.5%	\$25,200	\$52,700	\$68,800	\$58,000	\$105	518	\$108	\$108	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1			
24(c)	Dakin Creek	6.84	1,869	0.432	1,869	1,869	0.000	1,178	1,869	37.9%	1,178	1,869	37.9%	1,178	1,869	37.9%	83.3%	\$38,000	\$76,500	\$101,100	\$87,300	\$97	580	\$98	\$98	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1			
	White Creek	3.39	1,339	0.812	1,339	1,339	0.000	801	1,339	58.4%	801	1,339	58.4%	801	1,339	58.4%	83.3%	\$45,000	\$93,000	\$123,000	\$103,500	\$107	580	\$108	\$108	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1		
	Hill Creek Lower	2.96	1,353	0.714	1,353	1,353	0.000	838	1,353	61.9%	838	1,353	61.9%	838	1,353	61.9%	78.8%	\$14,800	\$32,000	\$39,300	\$33,600	\$51	557	\$52	\$52	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1		
	Hill Creek Twin Lake	3.90	1,244	0.635	1,244	1,244	0.000	893	1,244	67.1%	893	1,244	67.1%	893	1,244	67.1%	85.4%	\$15,000	\$33,000	\$39,500	\$33,800	\$47	562	\$47	\$47	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1		
	Spring Creek	2.82	1,305	0.725	1,305	1,305	0.000	749	1,305	57.4%	749	1,305	57.4%	749	1,305	57.4%	81.7%	\$14,800	\$32,000	\$39,300	\$33,600	\$51	557	\$52	\$52	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
	Roy Creek	5.58	2,598	0.660	2,598	2,598	0.000	1,458	2,598	38.2%	1,458	2,598	38.2%	1,458	2,598	38.2%	77.0%	\$27,400	\$53,800	\$72,100	\$58,500	\$104	505	\$105	\$105	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
	Wurche Creek	5.66	4,086	1.350	4,086	4,086	0.000	2,895	4,086	40.2%	2,895	4,086	40.2%	2,895	4,086	40.2%	81.7%	\$27,400	\$53,800	\$72,100	\$58,500	\$104	505	\$105	\$105	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1		
	South Shore	1.42	597	0.657	597	597	0.000	347	597	47.8%	347	597	47.8%	347	597	47.8%	70.0%	\$3,600	\$7,700	\$10,300	\$8,500	\$32	536	\$33	\$33	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
	SW Shore	0.46	180	0.640	180	180	0.000	158	180	87.8%	158	180	87.8%	158	180	87.8%	10.2%	\$300	\$590	\$790	\$620	\$18	520	\$18	\$18	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1		
	Bayers Cove	4.39	653	0.311	653	653	0.000	338	653	48.7%	338	653	48.7%	338	653	48.7%	72.0%	\$14,800	\$32,000	\$39,300	\$33,600	\$51	557	\$52	\$52	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
	Northwestern Bay	1.98	480	0.384	480	480	0.000	338	480	37.1%	338	480	37.1%	338	480	37.1%	43.1%	\$5,400	\$11,000	\$14,300	\$11,200	\$20	568	\$21	\$21	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
	Pigeon Corn/McCallum Bay	2.40	331	0.218	331	331	0.000	205	331	63.2%	205	331	63.2%	205	331	63.2%	6.4%	\$901	\$1,910	\$2,510	\$2,000	\$42	582	\$43	\$43	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
	Green Lake North (f)	1.80	817	0.709	817	817	0.000	518	817	63.4%	518	817	63.4%	518	817	63.4%	42.0%	\$27,400	\$53,800	\$72,100	\$58,500	\$104	505	\$105	\$105	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
	Green Lake North (f)	1.74	772	0.883	772	772	0.000	526	772	47.1%	526	772	47.1%	526	772	47.1%	46.2%	\$5,100	\$10,500	\$13,500	\$10,500	\$33	536	\$33	\$33	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
	Puchyan	1.21	557	0.721	557	557	0.000	336	557	59.2%	336	557	59.2%	336	557	59.2%	42.0%	\$3,600	\$7,700	\$10,300	\$8,500	\$32	536	\$33	\$33	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
	For R. Berlin - UF68	0.96	379	0.801	379	379	0.000	230	379	59.9%	230	379	59.9%	230	379	59.9%	77.4%	\$4,400	\$9,300	\$12,300	\$10,300	\$24	561	\$25	\$25	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
	25(c)	Silver Cr. 1	8.01	1,840	0.481	1,840	1,840	0.000	1,178	306	36.4%	891	232	51.6%	1,016	264	45.0%	48.0%	\$18,800	\$39,200	\$51,900	\$44,800	\$52	586	\$54	\$54	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1
Silver Cr. 2 (rotated, est.)		4.48	2,357	0.800	2,357	2,357	0.000	1,221	2,357	53.9%	1,221	2,357	53.9%	1,221	2,357	53.9%	75.3%	\$20,100	\$41,700	\$55,100	\$47,800	\$66	565	\$61	\$61	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
Silver Cr. 3 (rotated, est.)		9.35	2,115	0.353	2,115	2,115	0.000	1,348	2,115	36.2%	1,012	1,691	52.2%	1,240	2,037	41.3%	75.3%	\$45,000	\$88,800	\$118,200	\$93,300	\$101	511	\$111	\$111	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
Silver Cr. 4 (rotated, est.)		5.39	1,765	0.508	1,765	1,765	0.000	1,343	1,765	23.4%	1,177	1,341	32.9%	1,240	1,682	28.7%	81.7%	\$16,500	\$34,200	\$45,200	\$38,500	\$71				\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
Silver Cr. 5 (rotated, est.)		5.70	1,653	0.290	1,653	1,653	0.000	1,086	1,653	35.2%	1,086	1,653	35.2%	1,086	1,653	35.2%	74.9%	\$16,500	\$34,200	\$45,200	\$38,500	\$71				\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
Silver Cr. 6 (rotated, est.)		5.33	1,573	0.148	1,573	1,573	0.000	386	1,573	32.8%	386	1,573	32.8%	386	1,573	32.8%	81.7%	\$32,100	\$66,500	\$87,900	\$75,000	\$120	554	\$124	\$124	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
Silver Cr. 7 (rotated, est.)		2.45	1,037	0.420	1,037	1,037	0.000	568	1,037	48.9%	568	1,037	48.9%	568	1,037	48.9%	81.9%	\$10,000	\$20,700	\$27,300	\$23,600	\$47	562	\$47	\$47	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1
Silver Cr. 8 (rotated, est.)		3.10	378	0.190	378	378	0.000	218	378	57.9%	218	378	57.9%	218	378	57.9%	71.9%	\$25,200	\$52,700	\$68,800	\$58,000	\$104	518	\$107	\$107	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1
Silver Cr. 9 (rotated, est.)		5.22	337	0.161	337	337	0.000	236	337	30.0%	236	337	30.0%	236	337	30.0%	73.5%	\$25,200	\$52,700	\$68,800	\$58,000	\$104	518	\$107	\$107	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
subwatersheds 1-23 (drain to lake)	47.06	15,108	0.504	12,241	601	10.3%	6,780	207	35.0%	7,424	1,578	51.1%	8,578	1,823	43.4%		\$211,300	\$433,100	\$573,600	\$485,800	\$73	580	\$74	\$74	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1		
	subwatershed 24 (rotated to lake) (f)	10.97	19,195	0.828	9,965	0.086	14.3%	6,813	1,830	25.2%	7,393	1,571	49.8%	8,061	1,704	30.4%		\$211,300	\$433,100	\$573,600	\$485,800	\$128	514	\$138	\$138	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	
	subwatersheds 25 (drain to lake) (f)	19.19	30,567	0.812	24,716	2.02	17.9%	20,644	2,233	35.1%	16,264	1,768	48.1%	18,892	2,044	37.4%		\$388,400	\$768,700	\$1,043,000	\$869,800	\$72	582	\$77	\$77	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1	\$1		

Table 6. Phosphorus Loads Routed to Big Green Lake, and Cost to Reduce Loads: Ranked by Total Reduction.

Table 6. Phosphorus Loads Round to Big Green Lake, and Cost to Reduce Loads: Ranked by Total Reductions.																														
Area	Current Practices CT, MT and NT (lb)	Total Phosphorus (lb/yr)	Area (sq. mi.)	Alt. A CT, MT and NT (lb)	Alt. B CT, MT and NT (lb)	Alt. C CT, MT and NT (lb)	Alt. D Dairy MT Cash crop all NT	Alternative A CT + MT				Alternative B CT + MT MT + NT				Alternative C all NT (no-DB)				Alternative D Dairy MT; Cash Crop (no-DB)				Total Cost to reduce Phosphorus Load						
								Cost/lb Phos. reduced				Total				Total				Total				Total						
								Phosphorus	percent	total	lb/acre reduced	Phosphorus	percent	total	lb/acre reduced	Phosphorus	percent	total	lb/acre reduced	Phosphorus	percent	total	lb/acre reduced	Alt. A	Alt. B	Alt. C	Alt. D			
								Phosphorus	percent	total	lb/acre reduced	Phosphorus	percent	total	lb/acre reduced	Phosphorus	percent	total	lb/acre reduced	Phosphorus	percent	total	lb/acre reduced	Alt. A	Alt. B	Alt. C	Alt. D			
								Phosphorus	percent	total	lb/acre reduced	Phosphorus	percent	total	lb/acre reduced	Phosphorus	percent	total	lb/acre reduced	Phosphorus	percent	total	lb/acre reduced	Alt. A	Alt. B	Alt. C	Alt. D			
18	Wurcho Creek	5,981	4,860	356	\$25	\$28	\$26		3,784	1,044	22.7%	1,112	2,895	0.760	40.9%	2,001	1,954	0.530	66.1%	2,842	2,821	0.720	46.5%	2,278	\$28,325	\$55,585	\$74,468	\$58,141	1	18
10	Roy Creek	5,317	3,336	1,980	\$54	\$60	\$56		1,648	0.517	21.7%	511	1,458	0.408	38.2%	905	1,551	0.264	55.4%	1,308	1,334	0.374	43.4%	1,023	\$27,426	\$53,821	\$72,105	\$56,290	2	10
10	Dakin Creek	6,861	1,888	7,433	\$97	\$108	\$96		1,509	0.345	20.1%	381	1,178	0.260	37.6%	711	870	0.199	54.0%	1,019	899	0.228	47.1%	880	\$38,688	\$76,495	\$101,094	\$87,265	4	10
3	Silver Cr. 3 (rounded, est.)	9,531	2,116	3,353	\$101	\$115	\$107		1,689	0.279	21.1%	448	1,348	0.225	36.3%	787	1,012	0.169	52.2%	1,103	1,240	0.207	41.3%	874	\$44,952	\$68,313	\$116,101	\$92,269	3	3
1	Silver Cr. 1	6,017	1,840	4,451	\$52	\$58	\$54		1,433	0.365	19.8%	368	1,179	0.306	36.4%	673	881	0.232	51.8%	968	1,018	0.264	46.0%	833	\$18,949	\$39,283	\$55,918	\$44,631	5	1
2	Silver Cr. 2 (rounded, est.)	4,491	1,582	3,550	\$56	\$65	\$61		1,228	0.428	22.6%	357	645	0.328	40.3%	636	681	0.237	57.0%	902	798	0.277	46.7%	787	\$20,110	\$41,891	\$55,068	\$47,577	6	2
13	Hill Creek Twin Lake	3,891	1,484	3,639	\$47	\$53	\$49		1,144	0.496	21.9%	330	603	0.362	38.3%	580	651	0.283	55.5%	812	825	0.358	43.7%	639	\$15,034	\$29,595	\$36,527	\$30,881	7	13
11	White Creek	3,352	1,330	3,812	\$68	\$76	\$70		1,055	0.468	20.6%	274	822	0.376	38.2%	507	608	0.279	54.5%	724	688	0.322	47.5%	631	\$18,568	\$38,494	\$50,872	\$43,929	10	11
12	Hill Creek Lower	2,991	1,353	3,714	\$51	\$57	\$52		1,061	0.560	21.6%	292	839	0.443	38.0%	514	606	0.320	55.2%	747	787	0.405	43.3%	586	\$14,910	\$29,260	\$38,169	\$30,605	8	12
10	Spring Creek	2,827	1,305	3,728	\$28	\$31	\$28		1,017	0.565	22.1%	288	784	0.441	39.2%	511	559	0.310	57.2%	746	720	0.403	44.4%	579	\$8,076	\$15,850	\$21,224	\$16,579	9	14
4	Silver Cr. 4 (rounded, est.)	5,391	1,163	3,068	\$71	\$83	\$78		1,521	0.441	13.2%	222	1,343	0.588	23.4%	410	1,177	0.241	32.6%	676	1,248	0.362	36.7%	904	\$18,465	\$34,198	\$45,185	\$38,020	11	4
5	Silver Cr. 5 (rounded, est.)	5,701	763	3,006	\$187	\$223	\$211		614	0.168	19.6%	149	503	0.138	34.1%	280	400	0.119	47.6%	363	440	0.121	42.3%	323	\$28,002	\$68,052	\$76,720	\$66,247	13	5
6	Beyers Cove	4,291	853	3,311	\$62	\$69	\$63		702	0.258	17.7%	181	585	0.213	31.4%	268	485	0.189	45.5%	368	549	0.200	35.6%	304	\$9,362	\$18,371	\$24,613	\$19,216	12	6
6	Silver Cr. 6 (rounded, est.)	5,331	875	3,188	\$383	\$344	\$342		482	0.135	19.4%	111	389	0.113	32.8%	188	318	0.093	44.9%	257	341	0.100	40.4%	232	\$32,075	\$69,498	\$87,881	\$75,885	15	6
22	Green Lake North (f)	1,861	617	2,708	\$32	\$35	\$32		702	0.609	14.0%	115	613	0.632	24.9%	204	519	0.450	36.5%	268	687	0.509	28.2%	230	\$3,677	\$7,215	\$9,666	\$7,547	14	22
23	Green Lake North (f)	1,741	772	2,693	\$53	\$56	\$55		676	0.607	12.4%	96	602	0.541	22.0%	170	528	0.472	31.9%	240	580	0.521	24.8%	192	\$5,130	\$10,086	\$13,488	\$10,529	16	23
20	Nowegian Bay	1,811	649	2,364	\$59	\$68	\$61		407	0.321	18.4%	82	338	0.267	32.3%	181	285	0.209	46.9%	234	317	0.250	36.5%	182	\$5,444	\$10,682	\$14,311	\$11,173	17	20
17	South Shore	1,432	567	3,907	\$33	\$38	\$33		507	0.568	15.1%	90	438	0.482	26.7%	159	366	0.403	38.7%	231	416	0.457	30.4%	162	\$2,939	\$5,789	\$7,729	\$6,034	18	17
7	Silver Cr. 7 (rounded, est.)	2,451	329	3,210	\$131	\$159	\$151		253	0.181	23.1%	78	169	0.127	30.5%	130	149	0.095	54.9%	181	170	0.108	46.5%	180	\$9,957	\$20,642	\$27,260	\$23,596	19	7
8	Silver Cr. 8 (rounded, est.)	3,221	337	3,101	\$48	\$51	\$48		278	0.083	18.3%	82	236	0.071	30.0%	101	202	0.060	40.2%	136	214	0.084	36.6%	124	\$25,150	\$52,141	\$68,897	\$58,801	20	8
8	Silver Cr. 8 (rounded, est.)	3,301	171	3,568	\$621	\$742	\$721		148	0.074	14.7%	25	128	0.064	25.5%	44	112	0.059	31.8%	55	\$15,805	\$32,352	\$42,755	\$36,919	21	8				
8	SW Shore	0.491	189	\$840	\$18	\$20	\$18		172	0.584	8.8%	17	159	0.538	15.9%	30	145	0.489	23.5%	44	155	0.525	18.0%	34	\$300	\$589	\$790	\$617	22	18
21	Pigeon Cove/Malcolm Bay	2,491	331	3,216	\$62	\$69	\$64		315	0.206	4.8%	18	303	0.168	6.3%	28	291	0.190	12.1%	40	300	0.189	9.4%	31	\$975	\$1,913	\$2,563	\$2,001	23	21