# **Green Lake Buffer Assessment Project**

# **Green Lake County Land Conservation Department**

December 31, 2014



An assessment of current riparian buffer and stream conditions in select sub-watersheds of Green Lake.

# Green Lake County, WI

Derek R Kavanaugh



Land Conservation Department Green Lake County 571 County Rd A PO Box 3188 Green Lake, WI 54941 (920) 294-4051



# **TABLE OF CONTENTS**

Description of Project Area	8
Description of Problem to be Addressed	8
Survey Goals and Objects	10
Description of Methods and Activities	10
Overview of USDA_NRCS Stream Visual Assessment Protocol (SVAP)	10
Wisconsin Buffer Initiative	12
Green Lake Watershed	13
NRCS Stream Visual Assessment Protocol (SVAP)	
Channel Condition	16
Riparian Zone	18
Bank Stability	22
Water Appearance	23
Nutrient Enrichment	24
Instream Fish Cover	25
Bank Height	26
Sub Watershed of Green Lake	
Roy Creek	27
Wuerches Creek	35
Hill Creek	41
White Creek	48
Dakin Creek	55
Mandatory Stream Buffers	63
Appendix	70

# Figures

# Green Lake Watershed

# Roy Creek

Watershed	27
Restoration Photographs	29
SVAP Scores	30
Riparian Buffers	30
Bank Stability	31
Bank Stability Photographs	31
Bank Height	30
Ground Cover	33
Shrub Cover	33
Tree Canopy	33
Land Use	32

# **Wuerches Creek**

Watershed	35
SVAP Scores	37
Riparian Buffers	37
Bank Stability	38
Bank Stability Photographs	38
Bank Height	36
Ground Cover	40
Shrub Cover	40
Tree Canopy	40
Land Use	39

# Hill Creek

Watershed	42
SVAP Scores	44
Riparian Buffers	44
Bank Stability	45
Bank Stability Photographs	45
Bank Height	44
Ground Cover	47
Shrub Cover	47
Tree Canopy	47
Land Use	46

# White Creek

Watershed	49
SVAP Scores	51
Riparian Buffers	51
Bank Stability	<i>52</i>
Bank Stability Photographs	52
Bank Height	51
Ground Cover	54
Shrub Cover	54
Tree Canopy	54
Land Use	53

# Figures (continued)

# Dakin Creek

Watershed	56
SVAP Scores	<b>58</b>
Riparian Buffers	<b>58</b>
Bank Stability	59
Bank Stability Photographs	59
Bank Height	<b>58</b>
Ground Cover	61
Shrub Cover	61
Tree Canopy	61
Land Use	60

# **Appendices**

# **WUERCHES CREEK**

**Aerial** 

Landuse

**Soil Hydrologic Groups** 

Wetlands

**Buffer Width** 

**Ground Cover** 

**Shrub Cover** 

**Canopy Cover** 

**Contours** 

**Bank Height** 

**Bank Stability** 

**Bank Slope** 

**Hydrologic Features** 

**Water Clarity** 

**Nutrient Enrichment** 

**Habitat Features** 

**Survey Points** 

# **ROY CREEK**

Aerial

Landuse

Soil Hydrologic Groups

Wetlands

**Buffer Width** 

**Ground Cover** 

**Shrub Cover** 

**Canopy Cover** 

**Contours** 

**Bank Height** 

**Bank Stability** 

**Bank Slope** 

**Hydrologic Features** 

Water Clarity

**Nutrient Enrichment** 

**Habitat Features** 

**Survey Points** 

# **Appendices (continued)**

# HILL CREEK

**Aerial** 

Landuse

**Soil Hydrologic Groups** 

Wetlands

**Buffer Width** 

**Ground Cover** 

**Shrub Cover** 

**Canopy Cover** 

Contours

**Bank Height** 

**Bank Stability** 

**Bank Slope** 

Hydrologic Features

**Water Clarity** 

**Nutrient Enrichment** 

**Habitat Features** 

**Survey Points** 

# WHITE CREEK

Aerial

Landuse

Soil Hydrologic Groups

Wetlands

**Buffer Width** 

**Ground Cover** 

**Shrub Cover** 

**Canopy Cover** 

**Contours** 

**Bank Height** 

**Bank Stability** 

**Bank Slope** 

**Hydrologic Features** 

**Water Clarity** 

**Nutrient Enrichment** 

**Habitat Features** 

**Survey Points** 

# **Appendices (continued)**

# **DAKIN CREEK**

**Aerial** Landuse

**Soil Hydrologic Groups** 

Wetlands

**Buffer Width** 

**Ground Cover** 

**Shrub Cover** 

**Canopy Cover** 

Contours

**Bank Height** 

**Bank Stability** 

**Bank Slope** 

Hydrologic Features

Water Clarity

**Nutrient Enrichment** 

**Habitat Features** 

**Survey Points** 

# 2014 VOLUNTEER PHOSPHORUS MONITORING

ASSESSMENT OF HILL, ROY, SILVER, AND WUERCHES CREEKS

WISCONSIN BUFFER INITIATIVE FINAL REPORT

# **Description of the Project Area**

The project area entailed several sub-watersheds of Green Lake. These sub-watersheds include three 'impaired waters' (aka 303(d)), one listed trout waters, and one exceptional water resources water (aka EWR). The included waters are listed below

- (1) Hill Creek (Impaired Water)
- (2) Roy Creek (Impaired Water)
- (3) Wuerches Creek (Impaired Water)
- (4) Dakin Creek. (Class II Trout)
- (5) White Creek (Exceptional Water Resource)

The original project also included Assemble Creek and two Tributaries to Silver Creek. Silver Creek survey was initiated late in the field season, and much of the tributaries were dry, and could not be legally accessed. They will be part of a project to assess all of Silver Creek during a single field season. Assemble Creek was removed due to the lack of data it would provide. Assemble Creek is less than 500 feet long, and under one land ownership.

# Description of the problems to be addressed

Modern agricultural practices have significant impacts on water quality in many areas of the county due to non-point runoff. Past and current farming practices have increased the amount of sediment and nutrients leaving the uplands, and delivered to the streams and lakes throughout the County. In order to control and mitigate negative impacts, conservation practices have been developed, studied and implemented over the past decades. There are dozens individual of conservation practices that may be applied to an agricultural operation to reduce sediment movement and minimize impact on nearby waters.

One of these practices that have been gaining public attention is vegetative buffers, and in particular, riparian vegetative buffers. In a more general application, buffers are utilized along streams, drainage ditches, field borders, along steep slopes, and within fields to dissipate runoff and encourage sedimentation. A buffer can provide a wide variety of environmental services, dependent upon design. Buffers may be designed to trap sediment, nutrients, and pesticides contained in runoff. They also aid in stabilizing stream banks, storing storm water, providing habitat, and provide for recreational and aesthetics opportunities.

The efficacy of riparian buffers for improving water quality could be enhanced by distinguishing differences in buffer capability and capacity across the watersheds, and accounting for these differences in conservation planning.

There are two approaches for designing and implementing buffers as a conservation practice. The first is a purely arbitrary approach, by installing a standardized fixed width and type of buffer along all applicable areas, usually adjacent to perennial streams. The second approach is a precision conservation buffer, or variable width buffer. Within the theory of precision conservation buffers, each buffer is designed and installed according to the purpose for which it is intended, and in implemented with a compliment of conservation practices, known as a conservation system to achieve overall water quality goals. A systemic approach is very critical, due to the fact the closer to the source a pollutant can be controlled, the more effective the system is as a whole. The most effective approach, both from an economical and environmental sense is to prevent detachment of soil particle at the point of origin. Secondly, in a conservation system an inefficient conveyance system that tends to diffuse, not concentrate flow will result in less transport of pollutants. The historical approach to stormwater management tends to be highly efficient in its transport, which entails quickly collecting and channelizing runoff to an efficient transport system, such as a ditch or pipe. The methods aids in transporting sediment and other pollutants in a very effective manner, which is contrary to the goals of a conservation system. The third component of the conservation system is the last line of defense before a pollutant reaches surface water. While a critical component, it can be largely ineffective if not properly designed and installed on the landscape.

#### Description of the problems to be addressed (continued)

A precision riparian buffer is spatially variably buffer designed to achieve a specific conservation goal, such as entrapment of non-point source sediment. The design accounts for the contributing runoff areas, slopes, soil types, land use, and vegetative cover. A site-specific, designed riparian buffer is an integrated part of a watershed, rather than a standalone component.

The primary physical properties that affect sediment entrapment are vegetation density, vegetation spacing, vegetation type, initial abstraction, saturated hydraulic conductivity, and soil characteristics.

Large host of buffer studies has shown a corresponding wide variability in trapping efficacies for both sediment and nutrients. This variability can be attributed to the variety in landforms and flow patterns through the buffers, as well as the upland control of pollutants. It needs to be recognized that not all sections of a buffer receive equal portion of runoff in a diffuse sheet flow. Rather, field studies have found that over 80% of the buffer areas receive almost no significant runoff, while 9-18% (Dosskey et al 2002), known as the 'active buffer' receive the majority of the upland runoff, and are often inundated, and quickly are rendering ineffective for trapping pollutants when improperly designed and installed. These active buffer areas can be identified by determining specific contributing drainage area and slope index. The ratio of the total drainage area to the active buffer area is known as the convergence factor. By determining the drainage area (DA), watershed length, and watershed slop index, and property designed precision buffer can be installed, with the corresponding peak flow (Qp). The presence of active and non-active buffers areas, when applied to uniform, standardized buffers result 20% of the buffer receive the vast majority of the runoff, and are undersized, rendering them largely ineffective during storm events. Other studies have shown that utilizing a variable-width, conservation system approach can increase buffer efficacies by up to 400%, when compared to standalone, fixed width buffers.

Other factors that were found to be influential in buffer effectiveness are the slope and soil characteristics within the buffer itself. Both the slope and soil type factors control infiltration rates within the designed buffer area. Soil types also play a critical role, not only in infiltration rates, and water holding capacity, but also nutrient retention in relation the its cation exchange capacity (CEC).

Precision conservation, explicitly precision buffers should start by defining potential pollutants, and take a goal-oriented, rather than a standardized approach. While research is on-going, there are several pieces of information that can be collected to aid in directing conservation efforts to ensure the greatest return per effort.

A large watershed, such as the one associated with Green Lake, in combination with limited funding sources for conservation work, makes it critical to identity target areas for conservation work. This work entails both identifying specific sources areas contributing to impaired waters, but also threats or potential threats to exceptional water resources, and trout waters.

One of the principal objectives of the Green Lake Buffer Assessment Project is to provide locally gathered, scientifically sound data on the current condition of conservation efforts within the watershed to municipal officials, planners, and local residents so that planning policy decisions can be effectively explored and employed.

The proposed project will lay the groundwork for future conservation efforts directed at impaired and exceptional water resources surrounding Green Lake, and ultimately Green Lake itself.

# **Description of Goals and Objectives**

The primary goals of the project are to inventory vegetative buffer condition & spatial coverage, areas of convergence, soil types, watersheds, land use, bank condition, and physical stream status.

Through this data gathering process we will be able to assess the location and spatial extent of riparian buffer, identify impaired buffers, recognize floodplain areas, locate areas bank instability, classify target areas for precision buffers, and locate target areas for upland conservation practices.

# **Description of Methods and Activities**



# **Establish Sampling Stations**

Utilizing AutoCAD, each streamline was digitized from aerial photography. All visible streams and tributaries were included in the mapping. ArcMap 10 was then employed to place a GPS waypoint every 150 feet along the stream corridor. The GPS points converted to WGS84 Latitude-Longitude format and uploaded into a Garmin 76CSX hand held GPS.

# USDA-NRCS Stream Visual Assessment Protocol (SVAP)

SVAP is an easy-to-use assessment protocol to evaluate the condition of aquatic ecosystems associated with streams. The protocol does not require expertise in aquatic biology or extensive training.

The SVAP assessment protocol provides a basic level of stream health evaluation. It can be successfully applied by conservationists with little biological or hydrological training. This protocol is the first level in a four-part hierarchy of assessment protocols which can be utilized, depending on local needs.

This method provides an assessment based primarily on physical conditions within the assessment area. It may not detect some resource problems caused by stressors located beyond the immediate area being assessed. The use of higher tier methods is required to more fully assess the ecological condition and to detect problems originating elsewhere in the watershed. However, most landowners are mainly interested in evaluating conditions on their land, and this protocol is well suited to supporting that objective.

Assessment data was recorded utilizing a modified version of the USDA National Water and Climate Center Technical Note 99-1 (Stream Visual Assessment Protocol).

The SVAP tool was originally developed by the US Department of Agriculture and modified for Green Lake County streams. SVAP has been used extensively by the water resources programs to document stream conditions and relate these conditions to land use, water quality, and habitat. The SVAP tool is used to assign a numeric value to stream conditions such as channel stability, riparian health, water appearance, nutrient enrichment, and access to the floodplain. This numeric value allow for the mapping, comparison, and interpretation of varying conditions across a watershed.

With the data that is collected, and stream reaches photo-documented they can be monitored for changes over the long-term. Additionally, the information can be utilized to prioritize streams for restoration, educational and awareness campaigns, and revegetation efforts.

Organization and access of data has been made easier recently with a data format that can be imported to ArcGIS. This enables a spatial display of all SVAP data points, and allows for environmental data analysis.

#### Description of Methods and Activities (continued)

A Limited Term Employee (LTE) was hired for the 2013 field season to conduct the survey. Jordan Resop, a college student at the University of Wisconsin, Stevens Point was selected from a pool of applicants. Mr. Resop was field trained by Green Lake County Land Conservation Department staff for 2 days in the field. This training included GPS navigation and survey techniques. After the initial 2 training days, Mr. Resop conducted all the surveys included in this assessment. By utilizing a single person during one field season, any introduced basis (i.e. judgment of severe vs. moderate) should remain constant throughout the study.

Each survey point was field located with the use of a Garmin 76CSX GPS. A recreational grade GPS was selected for this project due to the large amount of canopy cover encountered during the survey. Recreational grade GPS, while less accurate than survey grade, accepts a larger range of satellite signals, allowing it to gain reception in areas that more accurate GPS units would not. The Garmin 76 series is WAAS enabled, making it capable of 3-meter accuracy, which was well with the tolerances required for the survey.

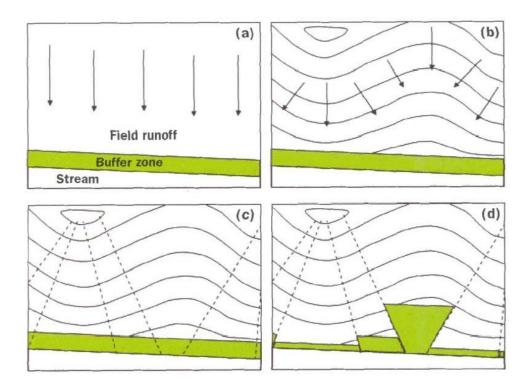
At each survey point a datasheet was completed, assessing each bank individually, as well as the stream channel proper.



#### **Wisconsin Buffer Initiative**

The Wisconsin Buffer Initiative was a collaborative effort between a diverse group of Wisconsin citizens and UW-Madison scientists to develop recommendations for the Wisconsin DNR on how riparian buffers can be part of a larger conservation system to address agricultural nonpoint source pollution.

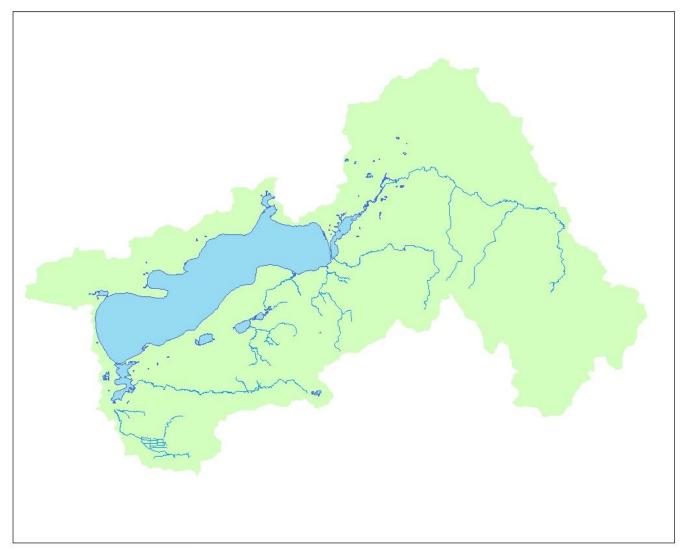
Rather than simply coming up with a fixed standard that would be uniformly applied across the diversity of Wisconsin's agricultural landscapes, the collaboration developed an innovative approach that identified site-specific areas where buffers, as part of a larger conservation system, would have the greatest likelihood of reducing pollution in waters that would benefit the most from this reduction. The WBI collaboration developed this science-based approach as it promises to be more efficient (i.e., water quality gains per dollar invested of public funds) and effective (i.e., the goal is to enhance water quality and not just install practices) than the approaches used in other water quality programs. How these goals were developed and the final recommendations on this science-based approach are contained in the report to the Wisconsin Natural Resources Board. It emphasized an adaptive management approach based on sound and applicable science to guide the policy process.



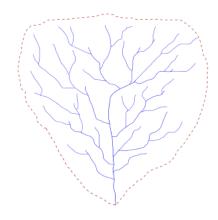
Above is an illustration from the Wisconsin Buffer Initiative Final Report (Dosskey et al 2005) depicting the concept how uniform buffers do not receive uniform amounts of runoff, and a more adaptive approach would be more efficient.

In lieu of reiterating portions of the report, a copy has been included as an appendix to this study.

# **Big Green Lake Watershed**



65,902 acres (103 square miles) drain to Green Lake



watershed, or drainage basin is an extent or an area of land where surface water from rain, melting snow, or ice converges to a single point at a lower elevation where the waters join another waterbody, such as a river, lake, reservoir, estuary, or wetland,. For example, a tributary stream of a brook that joins a small river is tributary of a larger river, which is thus part of a series of successively smaller area but higher elevation drainage basins.

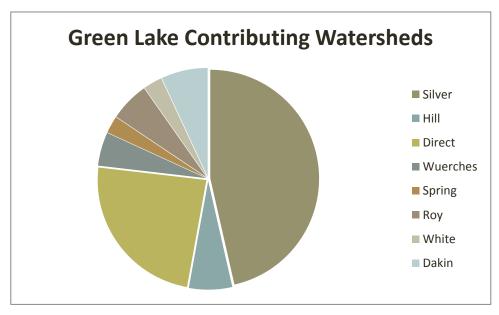
Other terms that are used to describe drainage basins are catchment, catchment area, catchment basin, drainage area, river basin, and water basin.

#### Watersheds (continued)

The drainage basin includes both the streams and rivers that convey the water as well as the land surfaces from which water drains into those channels.

The drainage basin acts as a funnel by collecting all the water within the area covered by the basin and channeling it to a single point. Each drainage basin is separated topographically from adjacent basins by a perimeter, the drainage divide making up a succession of higher geographical features (such as a ridge, hill or mountains) forming a barrier.

Drainage basins are similar but not identical to hydrologic units, which are drainage areas delineated so as to nest into a multi-level hierarchical drainage system. Hydrologic units are designed to allow multiple inlets, outlets, or sinks. In a strict sense, all drainage basins are hydrologic units but not all hydrologic units are drainage basins.



Contributing Watershed by Land Area (%)

# **Hydrology**

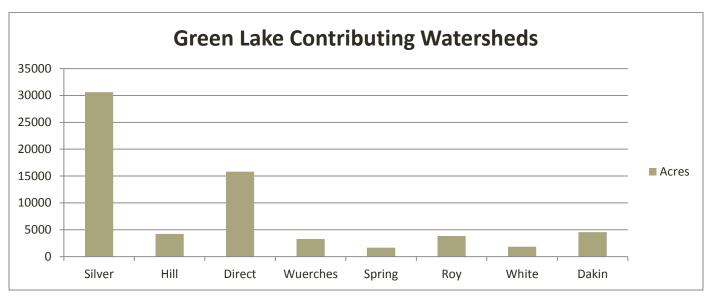
In hydrology, the drainage basin is a analytical unit of focus for studying the movement of water within the hydrological cycle, because the preponderance of water that discharges from the basin outlet originated as precipitation falling on the basin. A portion of the water that enters the groundwater system underneath the drainage basin may flow towards the outlet of another drainage basin because groundwater flow directions do not always match those of their overlying drainage network.

Watersheds are the principal hydrologic unit considered in fluvial geomorphology. A drainage basin is the source for water and sediment that moves through the river system and reshapes the channel.

Watersheds are important elements to consider also in ecology. As water flows over the ground and along rivers it can pick up nutrients, sediment, and pollutants. Like the water, they get transported towards the outlet of the basin, and can affect the ecological processes along the way as well as in the receiving water source.

Modern practice of using artificial fertilizers containing nitrogen, phosphorus, and potassium, has affected the receiving waters within the watershed. The pollutants will be carried by the drainage basin to the outlet and accumulate, disturbing the natural chemical balance. This can cause eutrophication where plant growth is accelerated by the additional nutrients.

Because watersheds are primary units in a hydrological sense, it has become common to manage water resources on the basis of individual basins.



Contributing watershed by land use (Acres)

A stream is a complex ecosystem in which several biological, physical, and chemical processes interact. Changes in any one characteristic or process have cascading effects throughout the system and result in changes to many aspects of the system.

Many factors can influence the integrity and character of streams. Several factors can combine to cause intense changes. Increased nutrient loads might not cause a change to a forested stream individually, but when combined with other stressors such as channel straightening and tree removal, the resulting shift in energy dynamics within the aquatic biological community could be profound. The resulting chemical and physical changes may further contribute to a completely different biological community.

Stream processes are in a subtle equilibrium. Stream power, sediment load, and channel roughness must be in equilibrium for a channel to be stable. Hydrologic changes that increase stream power (i.e. channel straightening), if not balanced by greater channel complexity and roughness (manning's n), result in "sediment starved" water that erodes banks or the stream bottom. Inversely, increases in sediment load beyond the transport capacity of the stream leads to downstream deposition.

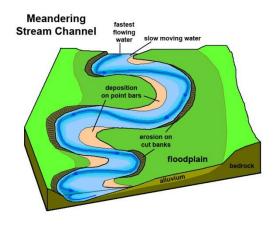
# **Channel Condition**



Stream meandering generally increases as the gradient of the surrounding valley decreases. Land development changes may affect the way a stream naturally does its work, such as the transport of sediment. Some modifications to stream channels have more impact on stream health than others. Channelization and damming affect a stream more than bridges and crossings.

Active downcutting and excessive lateral cutting are serious impairments to stream function. Both conditions are indicative of an unstable stream channel. Usually, this instability must be addressed before committing time and money toward improving other stream problems. For example, restoring the woody vegetation within the riparian zone becomes increasingly difficult when a channel is downcutting because banks continue to be undermined and the water table drops below the root zone of the plants during their growing season. In this situation or when a channel is fairly stable, but already

incised from previous downcutting or mechanical dredging, it is usually necessary to plant upland species. Extensive bank-armoring of channels to stop lateral cutting usually leads to more problems downstream. Often stability can be obtained by using a series of structures that reduce water velocity, deflect currents, or act as gradient controls. These structures are used in conjunction with large woody debris and woody vegetation plantings.



#### Channel condition was recorded in one of three categories, Natural, Altered, or Stressed.

#### Natural

No structures, dikes, evidence of down cutting, or excessive lateral cutting.

#### Altered

Evidence of past channel alteration. The channel is altered by riprap, channelization, or the floodplain is reduced by dikes or levys.

#### Stressed

The channel is actively down cutting or widening. The majority of the stream reach is riprapped or channelized.

# **Channel Condition (continued)**

Signs of channelization or straightening of the stream may include an unnaturally straight section of the stream, high banks, dikes or berms, lack of flow diversity (e.g., few point bars and deep pools), and uniform-sized bed materials (e.g., all cobbles where there should be mixes of gravel and cobble). In newly channelized reaches, vegetation may be missing or appear very different (different species, not as well developed) from the bank vegetation of areas that were not channelized. Older channelized reaches may also have little or no vegetation or have grasses instead of woody vegetation. Drop structures (such as check dams), irrigation diversions, culverts, bridge abutments, and riprap also indicate changes to the stream channel.

Indicators of downcutting in the stream channel include nick points associated with head cuts in the stream bottom and exposure of cultural features, such as pipelines that were initially buried under the stream. Exposed footings in bridges and culvert outlets that are higher than the water surface during low flows are other examples. A lack of sediment depositional features, such as regularly-spaced point bars, is normally an indicator of incision. A low vertical scarp at the toe of the streambank may indicate down cutting, especially if the scarp occurs on the inside of a meander. Another visual indicator of current or past down cutting is high streambanks with woody vegetation growing well below the top of the bank (as a channel incises the bank full flow line moves downward within the former bank full channel). Excessive bank erosion is indicated by raw banks in areas of the stream where they are not normally found, such as straight sections between meanders or on the inside of curves.



# **Riparian Buffer Zone**



plain, vegetation composition, and slope.

iparian buffers can be an effective protection for our water resources. These strips of grass, shrubs, and/or trees along the banks of rivers and streams filter polluted runoff and provide a transition zone between water and human land use. Buffers are also complex ecosystems that provide habitat and improve the stream communities they shelter.

Natural riparian buffers have been lost in many places over the years. Restoring the buffers and their associated functions will be an important step forward for water quality, stream bank stability, wildlife, and aesthetics. There are several components to a riparian buffer, including width of the natural vegetation zone from the edge of the active channel out onto the flood

A healthy riparian vegetation zone is one of the most important elements for a healthy stream ecosystem. The quality of the riparian zone increases with the width and the complexity of the woody vegetation within it.

Riparian buffers provides an assortment of interconnected ecological amenities, including:

- Reduces the amount of pollutants that reach the stream in surface runoff.
- Helps control erosion.
- Provides a microclimate that is cooler during the summer providing cooler water for aquatic organisms.
- Provides large woody debris from fallen trees and limbs that form instream cover, create pools, stabilize the streambed, and provide habitat for stream biota.
- Provides fish habitat in the form of undercut banks with the "ceiling" held together by roots of woody vegetation.
- Provides organic material for stream biota that, among other functions, is the base of the food chain in lower order streams.
- Provides habitat for terrestrial insects that drop in the stream and become food for fish, and habitat and travel corridors for terrestrial animals.
- Dissipates energy during flood events.
- Often provides the only refuge areas for fish during out-of-bank flows (behind trees, stumps, and logs).

The type, timing, intensity, and extent of activity in riparian zones are critical in determining the impact on these areas. Riparian zones that have roads, agricultural activities, residential or commercial uses, or significant areas of bare soils have reduced functional value for the stream. The filtering function of riparian zones can be compromised by concentrated flows. No evidence of concentrated flows through the zone should occur or, if concentrated flows are evident, they should be from land areas appropriately buffered with vegetated strips.

The width of the buffer can have significant impacts on the buffer's ability to attenuate pollutants, such as sediment and nutrients. It also impacts the type and amount of wildlife that will utilize the riparian corridor. Identifying and isolating a site specific restoration and management plan is essential in restoring the functionality of a riparian buffer. Just as critical is the vegetation composition of the riparian buffer. The physical dimension of the buffer has little value if the vegetation within the buffer is not able to adequately filter and attenuate pollutants at a capacity and degree proportional to the contributing area. Many buffer areas observed during the survey did not have adequate quantity or species composition of ground cover to stabilize banks nor filter upland pollutants. In some cases, the buffer area may be contributing more pollutants per unit area than the adjacent cropland.

#### Riparian Buffer Zone (continued)

The Riparian Zone was recorded in one of 5 categories (<10 feet, 10-35 feet, 35-75 feet, 75-100 feet, or greater than 100 feet). This dataset was then converted into one of three SVAP categories for the purpose of calculating a SVAP score.

The three categories included 1) Natural vegetation extends at least two channel widths on each side of the channel. 2) Natural vegetation extends at least one channel width on each side of the channel. 3) Natural vegetation extends less than one channel length on each side of the channel.

A common problem is lack of shrubs and understory trees. Another common problem is lack of regeneration. The presence of only mature vegetation and few seedlings indicates lack of regeneration. Healthy riparian zones on both sides of the stream are important for the health of the entire system. If one side is lacking the protective vegetative cover, the entire reach of the stream will be affected. There are many areas with evidence of concentrated flows through the riparian zone that are not adequately buffered before entering the riparian zone. These concentrated flows can reduce the effectiveness of the remainder of the buffer. Customized conservation plans need to be developed to adequately treat the water from concentrated flows.

Riparian buffers help catch and filter out sediment and debris from surface runoff. Depending upon many factors, including the width and complexity of the buffer, 20–100% of the sediments and the nutrients can settle out and be absorbed as buffer plants slow sediment-laden runoff waters. Wider, forested buffers are even more effective than narrow, grassy buffers.

The riparian buffer traps pollutants that could otherwise wash into surface and groundwater. Phosphorus and nitrogen from fertilizer and animal waste can become pollutants if more is applied to the land than plants can use. Because excess phosphorus bonds to soil particles, 80–85% can be captured when sediment is filtered out of surface water runoff by passing through the buffer. Chemical and biological activity in the soil, particularly of streamside forests, can capture and transform nitrogen and other pollutants into less harmful forms. These buffers also act as a sink when nutrients and excess water are taken up by root systems and stored in the biomass of trees.

By slowing the velocity of runoff, the riparian buffer allows water to infiltrate the soil and recharge the groundwater supply. Groundwater will reach a stream or river at a much slower rate, and over a longer period of time, than if it had entered the river as surface runoff. This helps control flooding and maintain stream flow during the driest time of the year.

Riparian buffer vegetation helps to stabilize streambanks and reduce erosion. Roots hold bank soil together, and stems protect banks by deflecting stream flows.

Riparian buffers can also reduce the amount of streambed scour by absorbing surface water runoff and slowing water velocity. When plant cover is removed, more surface water reaches the stream, causing the water to crest higher during storms or snowmelt. Stronger flow can scour streambeds, and can disturb aquatic life.

The distinctive habitat offered by riparian buffers is home to a multitude of plant and animal species, including those rarely found outside this narrow band of land influenced by the river. Continuous stretches of riparian buffer also serve as wildlife travel corridors.

Forested riparian buffers benefit aquatic habitat by improving the quality of nearby waters through shading, filtering, and moderating stream flow. Shade in summer maintains cooler, more even temperatures, especially on small streams. Cooler water holds more oxygen and reduces stress on fish and other aquatic creatures. A few degrees difference in temperature can have a major effect on their survival. Woody debris feeds the aquatic food web. It also can create stepped pools, providing cover for fish and their food supply while reducing erosion by slowing flow.

For every buffer there is a reason. Whether it is pollution filtration, erosion control, wildlife habitat, or visual screening, the size and vegetation of the buffer should match the land use and topography of the site.

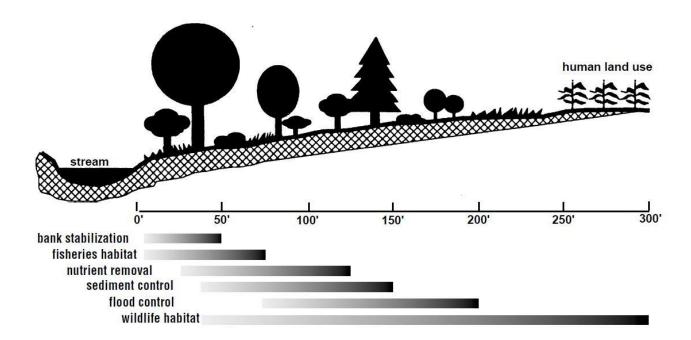
#### Riparian Buffer Zone (continued)

A buffer is more important for water quality in areas that collect runoff and deliver it to streams, and less critical on land that tips away from the water. Steeper slopes call for a wider riparian buffer below them to allow more opportunity for the buffer to capture pollutants from faster moving runoff. This is also true at both ends of a flood chute, or the path a river takes across a meander at high water.

The ability of the soil to remove pollutants and nutrients from surface and ground water also depends upon the type of soil, its depth, and relation to the water table. On a wetted soil, a wider buffer is needed to get the same effect.

The purpose(s) of the buffer will influence the kind of vegetation to plant or encourage. In urban and residential areas, trees and shrubs do a better job at capturing pollutants from parking lots and lawn runoff and providing visual screening and wildlife habitat. Between cropland and waterways, a buffer of shrubs and grasses can provide many of the benefits of a forested buffer without shading crops, and trees can be used on the north side of fields. Trees have several advantages over other plants in improving water quality and offering habitat. Trees are not easily smothered by sediment and have greater root mass to resist erosion. Above ground, they provide better cover for birds and other wildlife using waterways as migratory routes. Trees can especially benefit aquatic habitat on smaller streams. Native vegetation is preferable to non-native plants.

How big should a buffer be? One size doesn't fit all. It depends on what you want the buffer to do. There isn't one generic buffer which will keep the water clean, stabilize the bank, protect fish and wildlife, and satisfy human demands on the land. The minimum acceptable width is one that provides acceptable levels of all needed benefits at an acceptable cost.



A combination of trees, shrubs, and grasses provide good erosion control on most streams. Severe bank erosion on larger streams requires engineering to stabilize and protect the bank - but this engineering can be done with plants. For better stabilization, put more of the buffer in shrubs and trees.

#### Riparian Buffer Zone (continued)

For slopes gentler than 8%, most sediment settling occurs within a 50-75" wide buffer of grass, depending on variables. Greater width is needed on steeper slopes, for shrubs and trees, or where sediment loads are particularly high. As much sediment should be controlled in the upland as possible to increase the efficacy of the buffer, and prevent it from becoming overloaded with sediment or runoff.

A width greater than 100' or more may be necessary on steeper slopes and less permeable soils to allow runoff to soak in sufficiently, Vegetation and microbes to need sufficient time to attenuate nutrients and pesticides. Many pollutants can be captured with 100' on flatter slopes with permeable soils, although in less permeable soils, such as clay soils, this may not happen within 500'.

Buffer width impacts the type of fisheries a stream may contain. For cold water fisheries, the stream channel should be



Photo by Derek Kavanaugh

shaded completely. Unless there are problems with algae blooms, warm water fisheries do not require as wide a buffer or as much shade, but they still benefit from water cleaned by a buffer's filtering action. Studies show that buffers wider than 100' provide a healthier aquatic food web.

For many wildlife species a 300' buffer is a generally the accepted minimum. Much larger streamside forest buffer widths are needed for wildlife habitat purposes than for water quality purposes. The larger the buffer zone, the more valuable it is. Larger animals and interior forest species generally require more room. Some use so many habitats that would be nearly impossible to protect the size buffers they require. A narrow width may be acceptable for a travel corridor to connect larger areas of habitat. Continuity is important — even small patches of trees are better than none at all when it comes to migrating birds.

# **Bank Stability**



Treambank erosion is a natural process, but acceleration of this natural process leads to a disproportionate sediment supply, stream channel instability, land loss, habitat loss and other adverse effects. Streambank erosion processes, although complex, are driven by two major components: stream bank characteristics (erodibility) and hydraulic/gravitational forces. Many land use activities can affect both of these components and lead to accelerated bank erosion. The vegetation rooting characteristics can protect banks from fluvial erosion and collapse, and also provide internal bank strength. When riparian vegetation is changed from woody species to annual grasses and/or forbs, the internal strength is weakened, causing acceleration of mass wasting processes. Streambank aggradation or degradation is often a response to stream channel instability. Since bank erosion is often a symptom of a larger, more complex problem, the long-term solutions often involve much more than just bank stabilization.

Numerous studies have demonstrated that streambank erosion contributes a large portion of the annual sediment yield. In some studies, up to 90% of annual TSS loads were associated with streambank erosion during spring snow melt.

This component is the existence of or the potential for detachment of soil from the upper and lower stream banks and its movement into the stream. Some bank erosion is normal in a healthy stream. Excessive bank erosion occurs where riparian zones are degraded or where the stream is unstable because of changes in hydrology, sediment load, or isolation from the floodplain. High and steep banks are more susceptible to erosion or collapse. All outside bends of streams erode, so even a stable stream may have 50 percent of its banks bare and eroding. A healthy riparian corridor with a vegetated flood plain contributes to bank stability. The roots of perennial grasses or woody vegetation typically extend to the baseflow elevation of water in streams that have bank heights of 6 feet or less. The root masses help hold the bank soils together and physically protect the bank from scour during bankfull and flooding events. Vegetation seldom becomes established below the elevation of the bankfull surface because of the frequency of inundation and the unstable bottom conditions as the stream moves its bed load.

The type of vegetation is important. For example, trees, shrubs, sedges, and rushes have the type of root masses capable of withstanding high streamflow events, while Kentucky bluegrass does not. Soil type at the surface and below the surface also influences bank stability. For example, banks with a thin soil cover over gravel or sand are more prone to collapse than are banks with a deep soil layer.

#### Bank Stability was recorded in one of three categories (Stable, Moderately Stable, or Unstable)

**Stable Bank** – Banks are stabile with vegetation and/or roots protecting the bank from erosion. Vegetation is actively growing on the banks.

**Moderately Stable** – Banks are generally stable. Small areas of erosion may be present. Banks that are not actively eroding have little or no vegetation and/or roots protecting the banks.

**Unstable Banks** – Active erosion is present. Vertical, denuding banks indicate recent erosion. Little or no vegetation and/or roots are protecting the banks.

# **Water Appearance**



Stream appearance can be an indicator to a nearby or systemic pollution source. A recent event, such an upstream disturbance, can create localized turbidity. Recent runoff events from snowmelt or a rainfall can create temporary turbid conditions. Under some circumstances, the water may be turbid most of the time.

The depth to which an object can be clearly seen is a measure of turbidity. Turbidity is caused mostly by particles of soil and organic matter suspended in the water column. Water often shows some turbidity after a storm event because of soil and organic particles carried by runoff into the stream or suspended by turbulence. The water in some streams may be naturally tea-colored. This is particularly true in watersheds with extensive bog and wetland areas. Water that has slight nutrient enrichment may support communities of algae, which provide a greenish color to the water. Streams with heavy loads of nutrients have thick coatings of algae

attached to the rocks and other submerged objects. In degraded streams, floating algal mats, surface scum, or pollutants, such as dyes and oil, may be visible. Clarity of the water is an obvious and easy feature to assess. The deeper an object in the water can be seen, the lower the amount of turbidity. A pea-green color indicates nutrient enrichment beyond what the stream can naturally absorb.

#### Water appearance was recorded in one of three categories (Clear, Cloudy, or Turbid)

**Clear** – Water is very clear, or tea-colored but objects are visible in 3-6 feet of water.

**Cloudy** – Water is considerably cloudy most of the time, or after storm events. Objects are visible in 0.5 feet of water.

**Turbid** – Water is very turbid or muddy most of the time. Objects are not visible in 0.5 feet of water. Water may have obvious pollutants on the surface. Odors may be present.



#### **Nutrient Enrichment**



utrient enrichment is often reflected by the types and amounts of aquatic vegetation in the water. High levels of nutrients (especially phosphorus and nitrogen) promote an overabundance of algae, floating and rooted macrophytes. The presence of some aquatic vegetation is normal in streams. Algae and macrophytes provide habitat and food for all stream animals. However, an excessive amount of aquatic vegetation is not beneficial to most stream life. Plant respiration and decomposition of dead vegetation consume dissolved oxygen in the water. Lack of dissolved oxygen creates stress for all aquatic organisms and can cause fish kills. A landowner may have seen fish gulping for air at the water surface during warm weather, indicating lack of dissolved oxygen.

#### Nutrient enrichment was recorded in one of four categories (None, Light, Moderate, or Heavy)

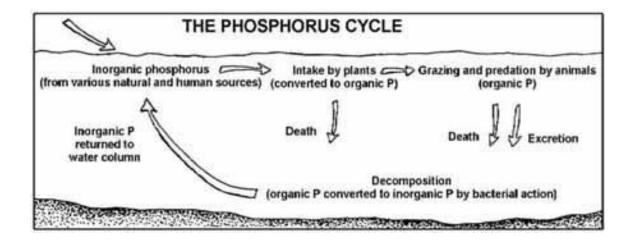
None - Water is clear along the entire section. Little or no algae growth is present. Plant species are in low abundance.

**Light** – Water is fairly clear or slightly colored green. Moderate algae growth on stream substrates.

**Moderate** – Greenish water along the entire reach. Overabundance of macrophytes. Abundant algae growth, especially during warmer months.

**Heavy** – Pea green, gray, or brown colored water. Dense stands of macrophytes which clog the stream and impede flow. Severe algae blooms creating mats at the surface.

What to look for: Some aquatic vegetation (rooted macrophytes, floating plants, and algae attached to substrates) is normal and indicates a healthy stream. Excess nutrients cause excess growth of algae and macrophytes, which can create greenish color to the water. As nutrient loads increase the green becomes more intense and macrophytes become lusher and deep green. Intense algal blooms, thick mats of algae, or dense stands of macrophytes degrade water quality and habitat. Clear water and a diverse aquatic plant community without dense plant populations are optimal for this characteristic.



#### **Instream Fish Cover**



tream systems often benefit from increased complexity and diversity in physical structure. Structural complexity is provided by trees fallen into the channel, overhanging banks, roots extending into the flow, pools and riffles, overhanging vegetation, and a variety of bottom materials. This complexity enhances habitat for organisms and also restores hydrologic properties that often have been lost.

This assessment component measures availability of physical habitat for fish. The potential for the maintenance of a healthy fish community and its ability to recover from disturbance is dependent on the variety and abundance of suitable habitat and cover available.

Several types of stream habitat and/or fish cover was recorded at each station (woody debris, pools, overhanging vegetation, cobbles, riffles, undercut banks, and macrophytes beds).

The presence or absence of each individual habitat features were recorded and summarized to create an average SVAP fish cover score.

Feature types are described below.

**Logs/large woody debris**—Fallen trees or parts of trees that provide structure and attachment for aquatic macroinvertebrates and hiding places for fish.

**Deep pools**—Areas characterized by a smooth undisturbed surface, generally slow current, and deep enough to provide protective cover for fish (75 to 100% deeper than the prevailing stream depth).

**Overhanging vegetation**—Trees, shrubs, vines, or perennial herbaceous vegetation that hangs immediately over the stream surface, providing shade and cover.

**Boulders/cobble**—Boulders are rounded stones more than 10 inches in diameter or large slabs more than 10 inches in length; cobbles are stones between 2.5 and 10 inches in diameter.

**Undercut banks**—Eroded areas extending horizontally beneath the surface of the bank forming underwater pockets used by fish for hiding and protection.

**Dense macrophyte beds**—Beds of emergent (e.g., water willow), floating leaf (e.g., water lily), or submerged (e.g., riverweed) aquatic vegetation thick enough to provide invertebrate attachment and fish cover.

**Riffles**—Area characterized by broken water surface, rocky or firm substrate, moderate or swift current, and relatively shallow depth (usually less than 18 inches).



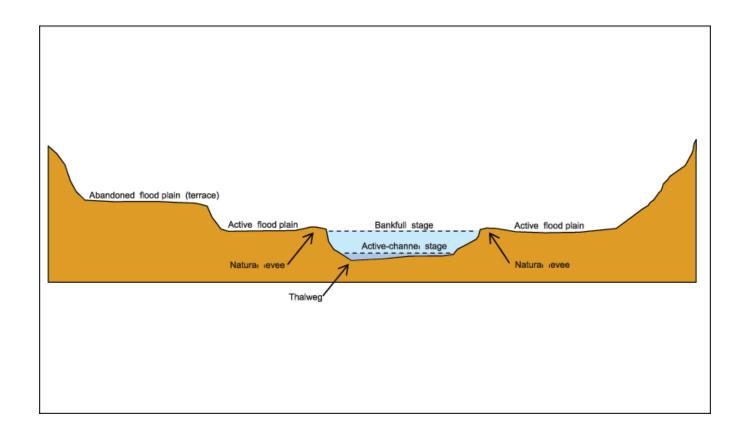
Bank Height can influence erosion potential, as well as floodplain connectedness. It is important to recognize that streams and flood plains need to operate as a connected system. Flooding is necessary to maintain the flood plain biological community and to relieve the erosive force of flood discharges by reducing the velocity of the water.

Flooding and bankfull flows are also essential for maintaining the instream physical structure. These events scour out pools, clean coarser substrates (gravel, cobbles, and boulders) of fine sediment, and redistribute or introduce woody debris.

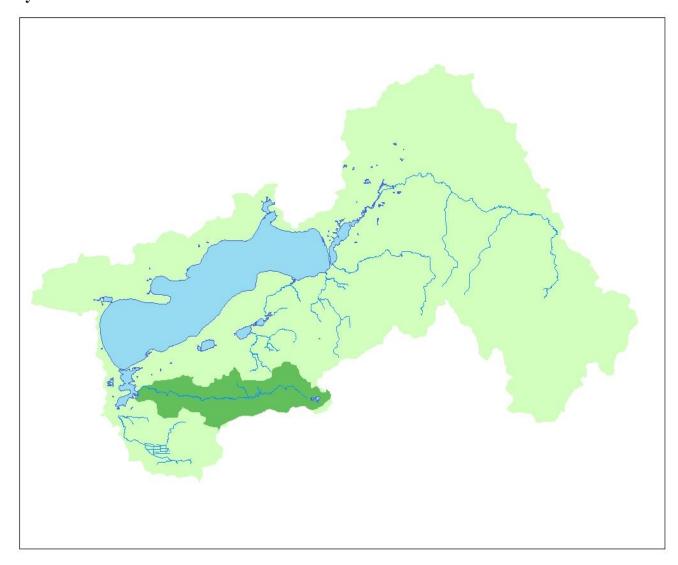
Bank height data can be used to identify floodplain connectedness, as well as assess potential risk areas when overlaid with actively eroding

banks. Streams with low bank heights will be more connected to adjacent floodplains, and not present a great erosion risk during high water events. In contrast, streams that are deeply entrenched will experience increased stream power during high water events, and will not have a connected floodplain to reduce the power, which will translate into bank erosion potential.

In addition, during the survey, bottom substrate (sand, silt, gravel) and hydrologic features (run, riffle, pool) data was recorded. These features can provide evidence as to the stream activity and gradient. Areas of deposition and/or erosion can be isolated from the statistics.



# **Roy Creek**



Roy Creek is a 6.8 mile long stream which flow west to County K Marsh on the west end of Green Lake. The drainage area is primarily agricultural cropland, making up 76% of the watershed, followed by woodland and wildlife land, jointly comprising 20% of the watershed. The entire watershed encompasses just under than 6% of the total drainage area for Green Lake.

Roy Creek is made of a main stem that starts 3.5 miles south of Green Lake and flows 35,924 feet (6.8 miles) west to Green Lake, where it empties into the east side of the backwaters of Green Lake, locally known as County Marsh. In addition, 11,556 feet (2.2 miles) of tributary streams flow into the Creek. Most of the tributaries are relatively short and steep, and may be a source of sediment loading to the Creek.

Roy Creek drops approximately 228 feet from beginning to end, with a mean slope of 0.63%. In contrast, the Fox River drops 36 feet from Portage, Wisconsin to Lake Winnebago, a distance of 110 miles. These equates to an average stream gradient of 0.006% along the Fox River. This means Roy Creek is 100 times (10,000%) steeper than the Fox River. In Wisconsin, high gradient streams typically fall in the 1-3% range, however specific ranges are not typically dictated. Roy Creek is a pool-riffle system, falling on the lower end of high gradient streams. Although the mean gradient is 0.63% (33 feet of drop per stream mile), the lower reaches of the stream are low gradient, denoting the upper reaches exceed the mean in several areas. A high gradient indicates a steep slope and rapid flow of water (i.e. more ability to erode); whereas a low gradient indicates a more nearly level stream bed and sluggishly moving water, that may be able to carry only small amounts of very fine sediment. High gradient streams tend to have steep, narrow V-shaped valleys, and are referred to as young streams. Low gradient streams have wider and less rugged valleys, with a tendency for the stream to meander.

#### Field Notes by Field Technician (Jordan Resop, Green Lake County LTE)

The stream was survey in June 2013 from the mouth of the stream to County Road N, where the stream ran dry.

The beginning stretch of Roy Creek is extremely marshy, dominated by Cattail and Reed Canary Grass with a few willow trees. It begins quite wide and silty with a good amount of downed trees crossing the stream. For the most part, this lowland area is all natural with only a couple of crossing bridges and ATV paths. The stream is very much the same throughout until it emerges from the forested area into an agricultural field on both sides. Here there is on average a 10ft – 20ft buffer of Reed Canary Grass, Cottonwood, and Buckthorn. The banks are fairly steep, but stable with no undercuts and grass roots growing almost down to water level. There still is a good amount of downed wood in or around the stream. The substrate is predominantly sand with some gravel and cobble mixed in. There were two separate areas locations that had PVC piping draining water into the stream from the left (north) bank. These were located at GPS points 59 and 62. The stream takes a 90 degree turn to the north, then another back to the east for a stretch. There is a farm road that crosses through the creek where there has been cobble laid in. After this location, the Green Lake County restored portion (GPS 64) begins and continues to County Rd O. The restored portion looks stable. The planted grasses are beginning to thicken more now and the banks are looking to be in good shape as well. There is mainly Box Elder here with a couple Cottonwoods mixed about.

Starting from the east side of the culvert of County Road O, the substrate is mainly sand with a lot of woody cover and a few cobbles and gravel mixed in along the way. The banks are steep for the most part at 1 to 3 feet high. There are a few elbows that have higher (~5') banks, which have considerable high water erosion (Note: In November 2013, 1400 feet of stream bank were restored. See photos on page 29). The beginning portion has a narrower buffer which eventually becomes wider, as the stream meanders into thicker woods, out of agricultural field. The canopy is made up largely of Box Elder.

Continuing upstream to the east, the stream stays much the same. The majority of the area is wooded, with many downed trees on the banks and in the water. There are multiple up rooted trees also. The wooded areas are dominated mostly by Basswood, with Maple appearing closer to County Road N. These shaded areas have a ground cover of various broadleaf plants. Opens stretches in the canopy lead to heavy Reed Canary Grass growth and some Buckthorn and Box Elder. Along the way, there are a few small, trickling tributary streams as well. Most of these flow from steep, higher elevation – they have definite evidence of fast flowing and high water erosion most likely from runoff. There are cobbles and sandstone exposed. There have been a few cobble farm road crossings that go through the stream as well as a few ATV and private bridges which cross.

The stream width is fairly narrow and cobble filled in the open, marshy areas and becomes much wider and sandy/silty in wooded areas. Patches of thick Garlic Mustard have been found on long woody stretches as well, controlling the entire undergrowth. The stream eventually breaks the wood line and comes into a ditched corn field. Here, there is thick Reed Canary Grass and no shrub or canopy cover. The stream slows and becomes basically sitting water with grass growing in the stream bed, all the way to Co Rd N.

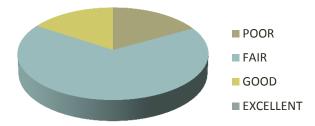
Once the stream crosses County Road N, it dries up as it flows through an agricultural field, with little to no buffer along the channel.

Roy Creek has been the subject of several stream restoration projects, and a catalyst for imploring a watershed stream survey. Several areas of Roy Creek were found to have denuded and severely eroded banks. Since the entire stream runs through private property, regular inspections of these corridors were not feasible. After discovering the severe degree of deterioration of the banks of Roy Creek, further exploration was conducted on area streams, and it was quickly illustrated that many tributaries to Green Lake were in poor condition, and likely contributing significant amounts of sediment and nutrients to the lake. It was also noted that many of the buffer along the creek, while intact, where wholly ineffective at reducing sediment loading to the stream from upland sources. This was an amalgamation of lack of ground cover (vegetation), steep slopes, and expansive monocultures of invasive plants.

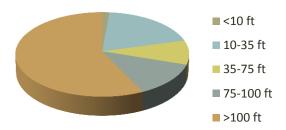
After consulting several resources professionals, it was concluded that a comprehensive, systematic approach to catalog the condition and extent of streambank and riparian buffers was essential.

# Roy Creek Restoration Project #001-12 (2012)

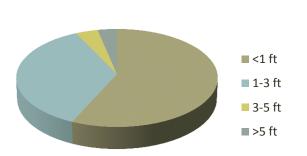




The stream visual assessment (SVAP) scores for Roy Creek placed over 84% of stream reaches in the Fair to Poor rating category. A weighted average of stream characteristics was calculated to determine a final score (poor, fair, good, or excellent).



The majority (79%) of Roy Creek has riparian buffers greater than 35feet. Buffer width was measured from top of bank to outside edge. No distinction was made between native and nonnative plants species for the purposes of scoring the buffer. Cropfields (hay) or maintained lawns (mowed) were not scored as buffers.



Ninety-two percent (92%) of Roy Creek's banks are less than 3 feet high.



A grassed waterway conveys concentrated flow through an agricultural field.

SVAP scores, Roy Creek

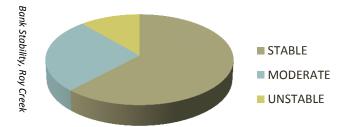
SVAP scores, Roy Creek



Example of a stable bank along Roy Creek. Trees and perennial vegetation provides erosion protection against streamflow.



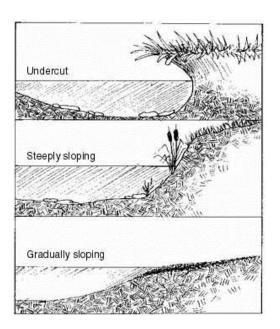
Example of a moderately stable bank along Roy Creek. No active erosion faces. Shows signs of past erosion with exposed tree roots. Bank vegetation consisting mainly of annuals. Presence of plants and tree roots to provide erosion resistance.

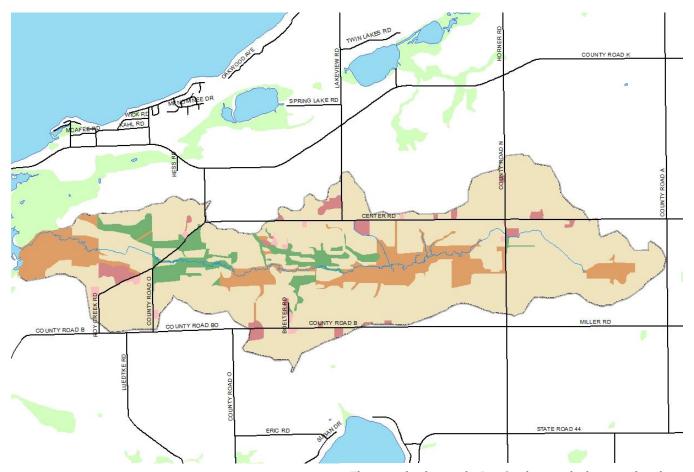


The vast majority of Roy Creek's streambanks are either stable or moderately stable (89%), leaving 11% of the streambank in an unstable, actively eroding state. While 11% is in the minority, its contribution to sediment and phosphorus delivery may be significant.

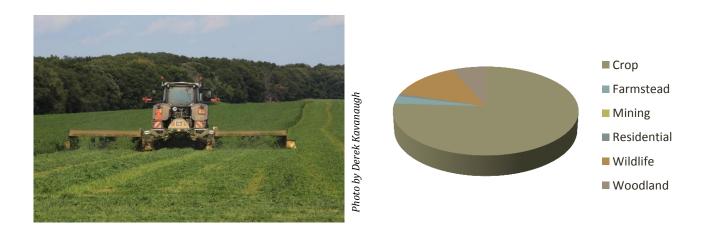


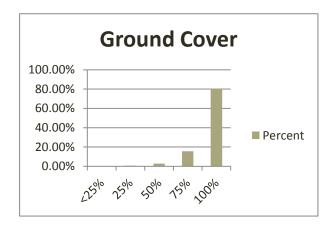
Example of an unstable bank along Roy Creek. Lack of vegetation and an active erosion face are signs of an unstable bank, which is contributing sediment and phosphorus to the stream.

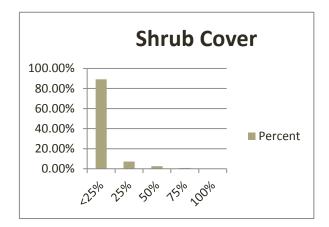


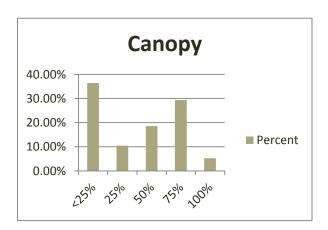


The major landuse in the Roy Creek watershed is agricultural.









Roy Creek's buffer area has a high percentage of ground cover, and a mixture of mature canopy trees. It most areas the buffer is lacking an understory of shrubs and young trees. This may be due to overgrazing by wildlife or the existence of a climax forest. **Forest** activities could be management implemented to create a more complex age structure within the buffer areas. A lack of shrubs to provide roots systems to anchor the soil, make bank vulnerable to future erosion, and mass wasting, especially in areas with high banks, and steep slopes.

# Wisconsin DNR Information on Roy Creek:

Watershed: UF07

Waterbody ID: 148200

Natural Community: Coldwater, Cool-Cold Headwater

Condition: Poor

**Trout Water: No** 

Outstanding Water Resource: No.

Impaired Water: Yes

Pollutants: Degraded Habitat, Impairment Unknown, Total Phosphorus, Sediment/Total Suspended Solids

Last Monitored: 2014

**Current Conditions:** 



Fish and Aquatic Life - Poor

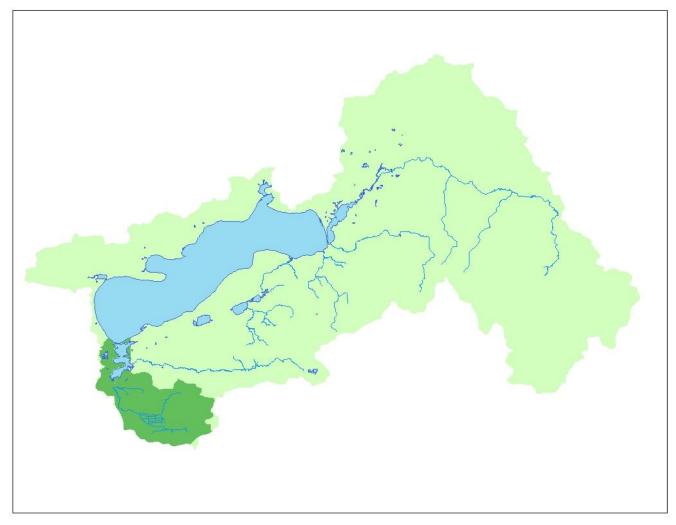


Fish Consumption - Unknown



Fish Consumption - Good

# WUERCHES CREEK



3288 acre watershed (5.0% of Green Lake's Watershed)

Werches Creek, including its tributaries, measures 10.4 miles. Tributaries include several agricultural drainage ditches, which flows to County K Marsh on the west end of Green Lake. The drainage area is primarily agricultural cropland, making up 71% of the watershed, followed by woodland and wildlife land, jointly comprising 21% of the watershed. The entire watershed comprised just fewer than 5% of the total drainage area for Green Lake.

Wuerches Creek is consists of a main stem that starts near Little Green Lake and flows 18,300 feet (3.5 miles) northwest to Green Lake. In addition 36,937 feet (7.0 miles) of farm ditches and tributary streams flow into the Creek. Wuerches Creek drops approximately 124 feet from beginning to end, with a mean slope of 0.67%.

#### **Wuerches Creek (continued)**

#### Field Notes by Field Technician (Jordan Resop, Green Lake County LTE)

#### Big Green to Co Rd B (July 2, 2013)

This area started in a marshy lowland thick with Reed Canary Grass and Dogwood. The Dogwood grows over the stream in places and is near 100% coverage in others. The stream meanders past a food plot, a couple deer stands, and under a snowmobile bridge. It this enters a wooded area dominated mainly by Box Elder. It gains elevation, proceeding into crop land, with high banks. The stream also widens out and contains cobble and downed trees. The banks for the most part are quite stable with little erosion. It travels under two more small wooden bridges and an area where a farm road crosses through. The stream stays similar for the remainder of the distance toward Co Rd B culverts.

#### Co Rd B to Luedtke Rd (July 3 - July 12, 2013)

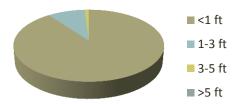
This entire stretch has very thick Reed Canary Grass buffers. It has been ditched, so is straight and fairly deep with a silty substrate and a high amount of nutrient enrichment (submerged grasses, duck weed, algae). There is little to no shrub or canopy cover throughout. Near GPS 69, the buffer lessens to about 10' and ag. field begins on the left (north) side. Further upstream the nutrient enrichment appears to less as well as the stream width. The various tributary channels, or ditches, are all surrounded by corn with a Reed Canary Grass buffer largely of 10'-15'.

In the area that has been made into drainage ditches, there appears to be heavy nutrient enrichment, with very dense Duckweed growth and certain areas of Wild Celery. The channels are roughly 4' deep and approximately 5'-10' wide with a silty substrate. There is little to no shrub or canopy growth here either. It is all ag. field, mostly corn. The buffers range from 10'-15' wide and are sloped or steep. This area has definitely been altered but is overall stable. The portion that runs to the south of Phil Marshall's land becomes more natural with gravel and cobble substrate. There are higher and less stable buffers nearer to Luedtke Rd. The stretch that goes runs from the north out of Brooks' corn field toward Luedtke becomes a bit more natural along the way. It is dry in most areas, just having puddles and silty mud from the previous heavy rain. There are a couple of areas which have some severe erosion and are quite unstable.

#### Luedtke Rd and East (July 15)

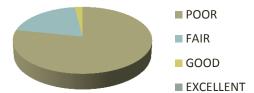
The northern stretch runs through ag. field but maintains a fairly healthy buffer of Box Elder and Walnut. There are areas here which also have very steep banks and are unstable. The stream dries out eventually, leaving a cobble and silt flood track.

The southern portion runs through an open area dominated completely by Reed Canary Grass. This grows within the stream channel which hold little to no water. This stretch will carry water from the dammed water body to the east, during times of high water.

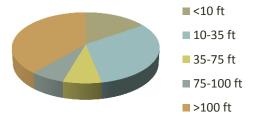


Over nine-eight (98%) of Wuerches Creek's banks are less than 3 feet high.

## **Wuerches Creek (continued)**



The stream visual assessment (SVAP) scores for Wuerches Creek placed over 98 % of stream reaches in the Fair to Poor rating category. A weighted average of stream characteristics was calculated to determine a final score (poor, fair, good, or excellent).



Fifty three (53%) percent of Wuerches Creek has riparian buffers greater than 35 feet. Buffer width was measured from top of bank to outside edge. No distinction was made between native and non-native plants species for the purposes of scoring the buffer. Cropfields (hay) or maintained lawns (mowed) were not scored as buffers.



Lower reach of Wuerches Creek. Photo by Jordan Resop

## **Wuerches Creek (continued)**



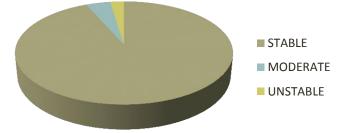
Example of a stabile streambank along Wuerches Creek. Many of the banks are covered in heavy Reed Canary Grass, while monotypic stands provide reduce ecology functions compared to a diverse mixture of native plant species, it does provide excellent erosion control and bank stability.



Example of a moderately stable streambank along Wuerches Creek. Annual vegetation does not provide full protection in the spring.

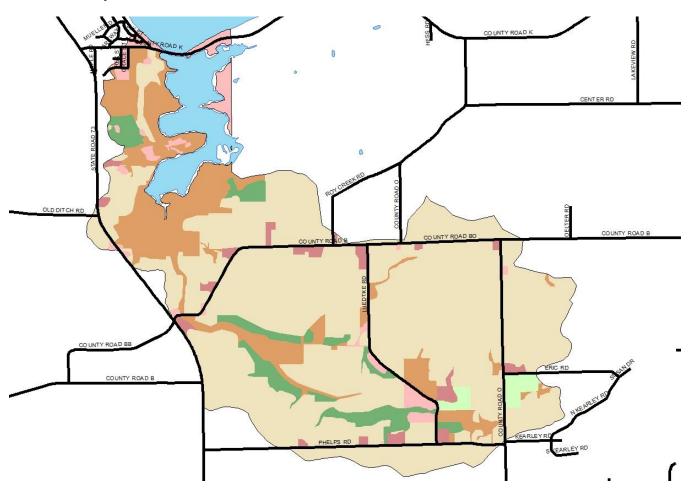


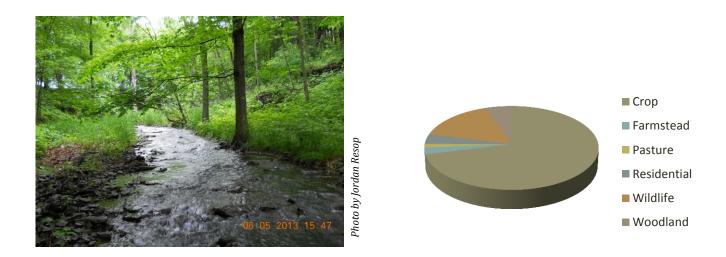
Example of an unstable streambank along Wuerches Creek. Actively eroding banks are a source of sediment and nutrients.



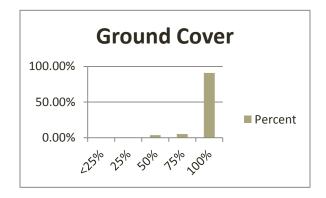
The vast majority of Wuerches Creek's streambanks are either stable of moderately stabile (97%), leaving less than 3% of the streambank length in an unstable, actively eroding state. It needs to be recognized that small, isolated bank stability concerns may not be fully characterized by the survey methods implored.

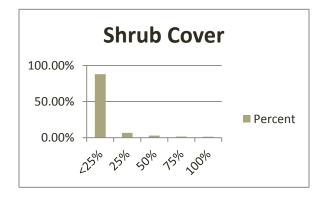
# **Wuerches Creek (continued**

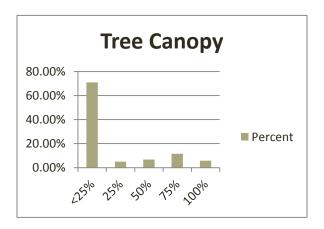




## **Wuerches Creek (continued)**







Wuerches Creek's buffer area has a high percentage of ground cover, relatively low occurrence of trees and shrubs. A significant footage of the tributaries are agricultural drainage ditches, which are typically maintained without tree and shrubs to reduce crop shading. A lack of trees and shrubs to provide roots systems to anchor the soil, make bank vulnerable to future erosion, and mass wasting, especially in areas with high banks, and steep slopes.



Example of a grass-only buffer along a Wuerches Creek tributary.

## **Wuerches Creek (continued)**

# Wisconsin DNR Information on Wuerches Creek:

Watershed: UF07

Waterbody ID: 148300

Natural Community: Coldwater

Condition: Poor

Trout Water: No.

Outstanding Water Resource: No.

Impaired Water: Yes

Impairment: Low DO, Elevated Water Temperature, Degraded Habitat

Pollutants: Total Phosphorus, Sediment/Total Suspended Solids.

Last Monitored: 2014

**Current Conditions:** 



Fish and Aquatic Life - Poor

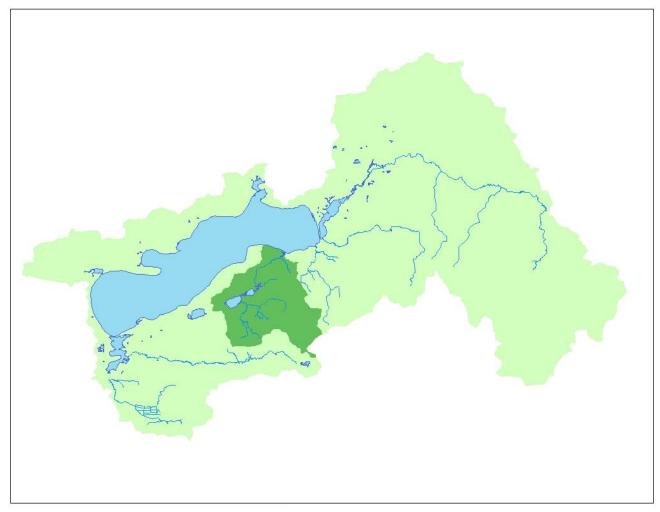


Fish Consumption - Unknown



Fish Consumption - Good

### **HILL CREEK**



4230 acre watershed (6.4% of Green Lake's watershed)

Hill Creek is a 1.81 mile long stream, with 3.91 miles of tributaries, which flows into Green Lake along the south shore. The drainage area is primarily agricultural cropland, making up 66% of the watershed, followed by woodland and wildlife land, jointly comprising 23% of the watershed.

From the upper tributaries to the outflow at Green Lake (17,662 feet), the stream drops 200 feet in elevation, with a mean gradient of 1,13%. This high gradient stream displays several areas of exposed bedrock as it's streambed as it makes it's decent to Green Lake. During heavy rains, and spring runoff, the amount of water and sediment transported through the steep channel can be significant.

#### Field Notes by Field Technician (Jordan Resop, Green Lake County LTE)

#### Lake View Rd to Little Twin Lake (May 29, 2013)

Working towards the south from Lake View, this portion has a very consistent substrate of silt, gravel, and cobble. It starts out with a fairly thick wooded canopy then begins to open up and yields thick canary grass on both sides for a turn and a stretch. At this point there begins to be algal growth on some of the submerged grasses of the substrate. The stream continues just on the inside of the wood line with a very stable bank (flat buffer)on the left and a steeper, less stable bank (and long, sloped buffer) on the right which has undercuts and exposed tree roots. The canopy reopens on the left side where there is a manmade pond and driveway. The assumed buffer was pushed out past the pond to a fence line where a crop field begins.

At this point, the stream appears to become altered. The stream flows under a culverted driveway and a couple other manmade crossings. It also loses its meander and emulates a man-made ditch. There are a few large tree stumps that have been cut with a saw. The remainder of the stream (to the dam on the north side of Little Twin) is very similar in depth, width, and substrate. On the left side of the last sampling point there is an inlet drainage from a wet lowland area.

### Horner Rd to Lake View Rd (May 30, 2013)

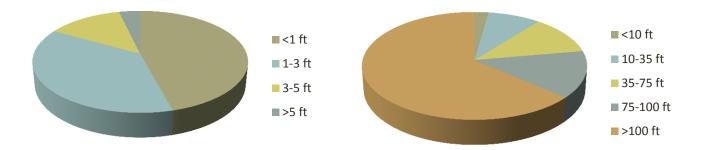
This stretch is surrounded by agriculture field with quite large buffers consisting primarily of reed canary grass throughout. There is little to no canopy covering the stream, however the few scattered trees throughout are Box Elder and Willow. About halfway through on the right side, there is an Oak finger that extends from the west on the far end of the buffer, next to the field line. The stream continues south, away from the oak finger, where there is a stand of Phragmites on the right side. Lake View road is approximately 600 feet upstream from this point. Also important to note, there were three small inlet drainages, all generously spaced from one another.

The substrate is pretty consistent, having a silt base with gravel and cobbles. Certain areas (mainly the second 2/3) have intermittent macrophyte beds (much of which being Elodea). Overall, the banks are fairly stable, with a few areas with moderate to severe washout, appearing above the current water line. For various reasons, the stream depth and width was recorded for this stretch. The depth varied from 8" to 2'5". The width varied from 4' to 10'.

### Spring Grove Rd to Horner Rd (May 31, 2013)

This is the bottom stretch of Hill Creek that dumps into Big Green. I started from the culvert on the south side of Spring Grove Road. Here, it was somewhat altered, ditch-like. It was deeper near the culvert but flattened out to about two FOW. This portion was quite congested throughout with fallen timber and log jams. For roughly the first 500 feet, there was a very small left side buffer. There is a pond and grass mowed to about 8'-15' of the stream bank until the stream turns into the wooded area. There were certain areas I came across that had large cobbles which created riffles, but the majority of the stretch was runs. At times, the banks would be higher (up to 5'-6') and areas of high stress with quite unstable banks, high water erosion, and undercuts.

The forested area contained primarily thick Reed Canary Grass and a few Garlic Mustard patches. There were also a handful of areas with various Forbes, which were shaded out by a thicker canopy of Willow or Box Elder. From here, the streams venture out of the forested area into an open area dominated by Reed Canary Grass. The stream narrows to about 6' and becomes faster moving. There are large cobbles scattered throughout this portion and the bank is low and thick with grass. The water appears to be high here, covering the roots of grass on the bank. There are a couple of areas that have small pools/eddies which have evolved. The final right turn of the stream has a high left bank with considerable erosion from higher water. This bank is steep and unstable.



Eighty-three (83%) percent of Hill banks area less than 3 feet high.

The majority (89%) of Hill Creek's riparian buffer are wider than 35 feet. Two percent (2%) are less than 10 feet.



Photo by Jordan Resop

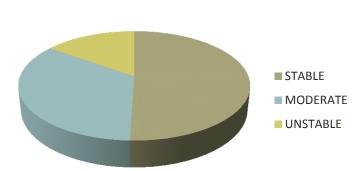
Photo by Jordan Resop



Example of a stable streambank along Hill Creek.



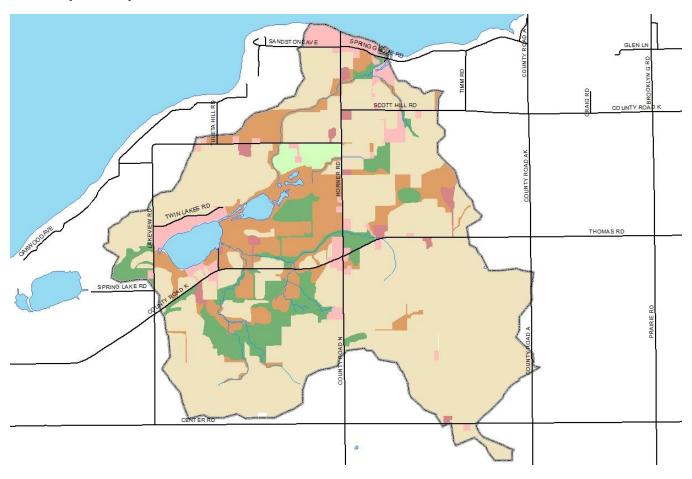
Example of a moderately stable streambank along Hill Creek.

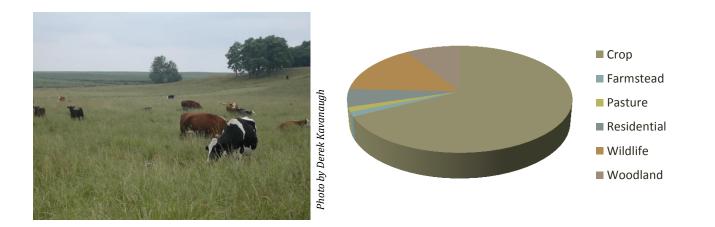


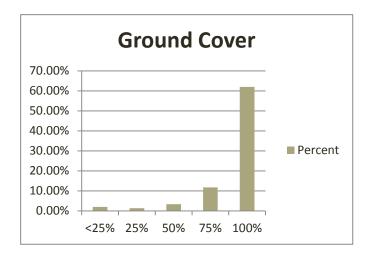
The vast majority of Hill Creek's streambanks are either stable or moderately stable (85%), leaving 15% of banks in an actively eroding state.

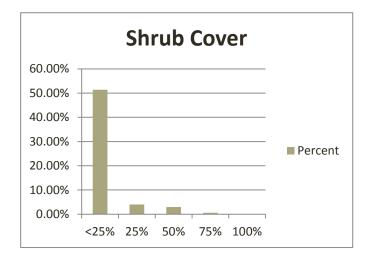


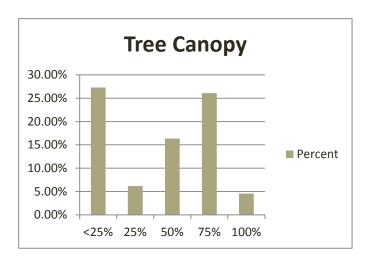
Example of an unstable streambank along Hill Creek.











Hill Creek's buffer area has a high percentage of ground cover, and a mixture of mature canopy trees. It most areas the buffer is lacking an understory of shrubs and young trees. This may be due to overgrazing by wildlife or the existence of a climax forest. Forest management activities be implemented to create a more complex age structure within the buffer areas. A lack of shrubs to provide roots systems to anchor the soil, make bank vulnerable to future erosion, and mass wasting, especially in areas with high banks, and steep slopes.



Example of a buffer with grasses and trees, lacking a shrub layer.

Photo by Jordan Resop

# Wisconsin DNR Information on Hill Creek:

Watershed: UF07

Waterbody ID: 146200

Natural Community: Cool-Warm Headwater

Condition: Poor

Trout Water: No.

Outstanding Water Resource: No.

Impaired Water: Yes.

Impairment: Degraded Habitat

Pollutants: Sediment, Total Suspended Solids

Last Monitored: 2014

### **Current Conditions:**



Fish and Aquatic Life - Poor



Fish Consumption - Unknown

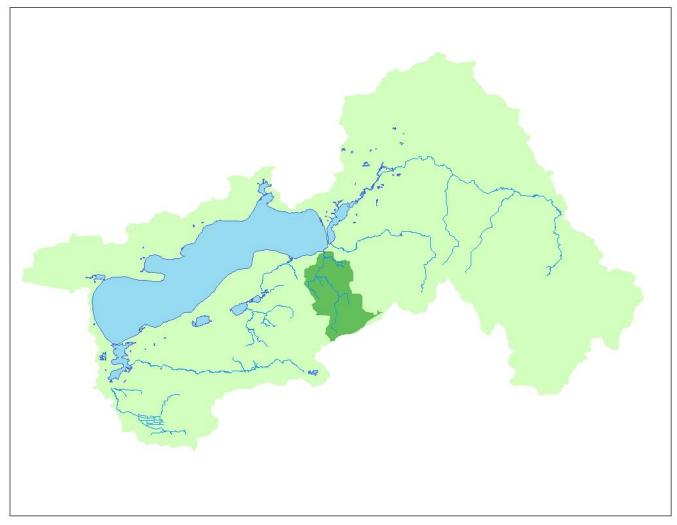


Fish Consumption - Good

### Phosphorus Monitoring\*:

Year	Mean (ug/L)	Min (ug/L)	Max (ug/L)	Threshold (ug/L)		
2014	147	97	200	75		
2014	59	16	84	<i>75</i>		
•	Monitoring locations are not the same.					

## WHITE CREEK



1857 acre watershed (2.8% of Green Lake's watershed)

The main stream only cover 6,914 feet (1.3 miles), while its tributaries add another 30,272 feet (5.7 miles), for a total length of 7.0 miles. The main channel drops 124 feet over its length, with a mean gradient of 1.8%. The longest tributary (17,157 feet), flowing north towards it's junction with the main creek, drops 240 feet over its path, with a mean bed slope of 1.4%. This steep slope gradient means this creek has a lot of potential to move sediment once introduced into the flow. Consequently, much of the streambed consists of bedrock and cobble.

### Field Notes by Field Technician (Jordan Resop, Green Lake County LTE)

Jordan Resop

Big Green to Scott Hill Rd (June 10, 2013)

The mouth of White Creek starts out in Big Green Lake. The area is dredged out to allow for boat access and docking. The banks are lined with rock and it is fairly straight up until the Spring Grove Rd culvert. After the culvert, it becomes much shallower with a cobble bottom. This portion begins forested with flat, stable banks dominated by forbes, Buckthorn, and Box Elder. There is a considerate amount of downed trees and a mixture of fast and slow waters. The stream forks at GPS 8, to the right (south), and will be discussed in the following paragraph. Continuing to County Rd K, the area's canopy opens up for the most part and there is thick Reed Canary Grass, with some Black Walnut. The substrate is majorly silt with 5+ foot banks.

The beginning point of the first branch (GPS 47) is a quite stagnant pool. The banks are very unstable and at least 5 feet high. The area is dominated by Reed Canary Grass and Black Walnut. The following portion has very steep, high banks as it proceeds out of the wooded area and into the open, towards a pond. The RCG is so thick the stream cannot be seen without physically pushing the grass away. The stream is also very narrow and only inches deep here. Past the pond, the stream flows down a limestone ridge and has extremely unstable and steep banks. This area begins to have a Box Elder canopy and cobble substrate. Further up the ridge, the stream bed becomes terraced bedrock with some algal growth on the cobble and slabs. The area is forested for the remained of the stretch and continues to gain elevation.

#### Scott Hill Rd to County Rd K (June 11, 2013)

This entire stretch is made up of cobbles and riffles and exhibits signage of high water flow (high banks, erosion, undercuts). The first point on this stretch is just north of the Scott Hill Rd culvert. It begins shaded but quickly moves out into an open area dominated by thick RCG. There is a run that enters from the right (south) but is dry and must only flow during wet periods. The main channel of hits portion continues in a forested area where it splits again, to the south. This short leg flows from an elevated area with a good amount of limestone substrate and banks up until it dries out at GPS 208. There are areas of this stretch that have close to 100% canopy cover, making for little or no undergrowth. Most of the canopy in the lower portion is Sugar Maple. Basswood and White Oak was found at the higher elevations near fence lines.

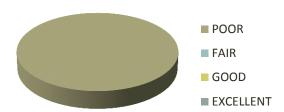
Back at the main channel, it continues to stay under the canopy and has a generous amount of fallen trees and wood in the channel. Mainly gravel and cobble substrate. It gains elevation where, nearer to Cnty Rd A, steep limestone walls are found on the stream banks. There is more signage of heavy erosion from high water and fast flow. Once again, the dominant species are Sugar Maple and Basswood with few Buckthorn shrubs and mainly RCG or small dicot ground cover. The stream crosses Cnty Rd A, into a bean field where it is has a thick RCG buffer of roughly 20' – 30' on each side. This area holds only sitting water at the time, up GPS 101 where it finally dries up.

### County Rd K to Craig Rd (June 12, 2013)

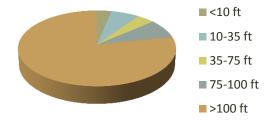
This stretch begins from the culvert on the east end of County Rd K. It is predominantly sand and cobble with low banks and thick Reed Canary Grass. There are some Willow trees scattered at the beginning until it opens into marshland with a large population of Dogwood. This marsh portion has a fairly wide and sloped drainage area. The stream meanders into a forest, widens, and becomes shallower. This entire forested area has mainly a rocky substrate, RCG and forbes on banks, large amounts of Buckthorn, and a thick canopy of Basswood, Box Elder, Willow, and Sugar Maple.

Near GPS point 33, the stream has another inlet stream from the right (south) which will be discussed here after the main run. The main run stays much the same, with stronger running riffles and some undercuts. It also runs past a few sheds and cabins as well as a few crossing bridges and rock driveways which run through the water. There are a few points of major erosion, showing signage of high water flow. The last point on the GPS, as of now, is directly downstream of a log home, before Mitchell's Glen.

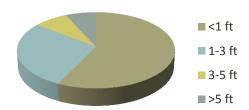
The stretch of the previously mentioned inlet stream is very similar to the other portions of this creek. With strong Buckthorn growth and similar canopy cover, there are also areas of widened riffles and cut out banks. Closer to the top of the stream, there is a small glen that exhibits very strong water flow during high water.



The stream visual assessment (SVAP) scores for Wuerches Creek placed 100 % of stream reaches in the Poor rating category. A weighted average of stream characteristics was calculated to determine a final score (poor, fair, good, or excellent).



Fifty three (90%) percent of Wuerches Creek has riparian buffers greater than 35 feet. Buffer width was measured from top of bank to outside edge. No distinction was made between native and non-native plants species for the purposes of scoring the buffer. Cropfields (hay) or maintained lawns (mowed) were not scored as buffers.

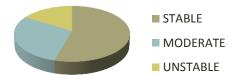


Eighty-five (85%) percent of White Creek's banks are less than 3 feet high.



Photo by Jordan Resop

Example of stable stream bank in lower reaches of White Creek.



Just over 84% of White Creek's banks are either stable or moderately stable.

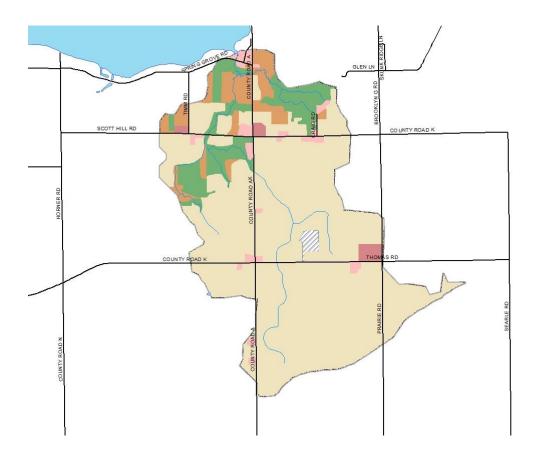


Photo by Jordan Resop

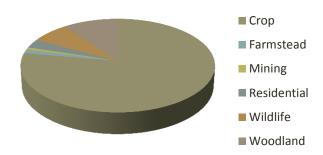
Example of moderately stable stream bank in lower reaches of White Creek.

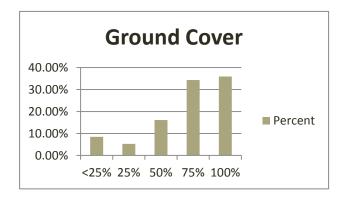


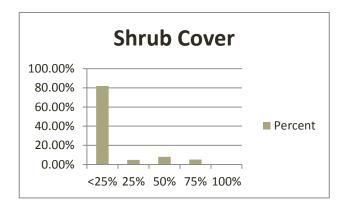
Example of an unstable stream bank in along White Creek. This area has since been restored and stabilized.

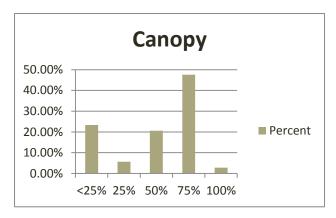












White Creek's buffer area has a high percentage of ground cover, and a mixture of mature canopy trees. It most areas the buffer is lacking an understory of shrubs and young trees. This may be due to overgrazing by wildlife or the existence of a climax forest. Forest management activities could be implemented to create a more complex age structure within the buffer areas. A lack of shrubs to provide roots systems to anchor the soil, make bank vulnerable to future erosion, and mass wasting, especially in areas with high banks, and steep slopes.

# Wisconsin DNR Information on White Creek:

Watershed: UF07

Waterbody ID: 146600

Natural Community: Coldwater

Condition: Unknown

Trout Water: Yes. (Class I)

Outstanding Water Resource: Yes.

Impaired Water: No.

Last Monitored: 2014

**Current Conditions:** 



Fish and Aquatic Life - Unknown



Fish Consumption - Unknown



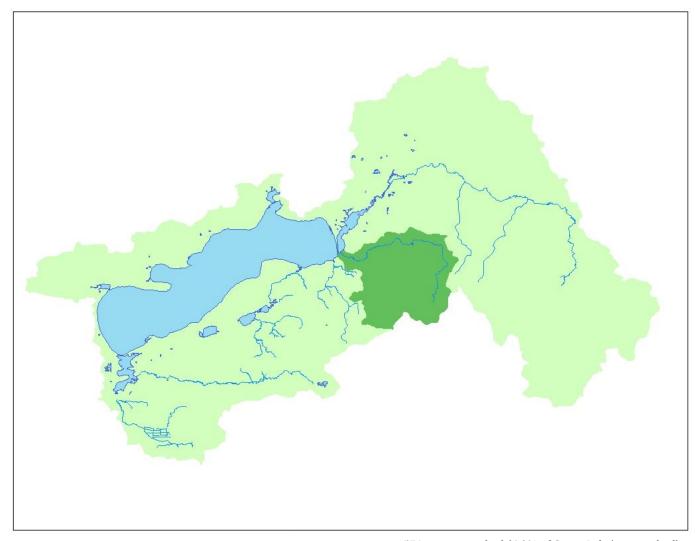
Fish Consumption - Good

## Phosphorus Monitoring\*:

Year	Mean (ug/L)	Min (ug/L)	Max (ug/L)	Threshold (ug/L)		
2011	<i>58.5</i>	<i>37</i>	836	<i>75</i>		
2014	80.5	<i>58</i>	163	<i>75</i>		
•	Monitoring locations are not the same.					

Long term water quality monitoring data is available from USGS  $\,$ 

## **DAKIN CREEK**



4571 acre watershed (6.9% of Green Lake's watershed)

akin Creek is a 3.4 mile long stream, flows to Silver Creek Marsh on the east end of Green Lake. The drainage area is primarily agricultural cropland, making up 83% of the watershed, followed by woodland and wildlife land, jointly comprising 13% of the watershed.

Dakin Creek starts in Fond du Lac County, following west to Green Lake. Over its 18,195 foot journey, it drops 152 feet in elevation with a mean gradient of 0.83%.

### Big Green Lake to Maug Rd (June 4, 2013)

This first stretch of Dakin Creek today I parked on Skunk Hollow's land with permission from Kent. I walked the creek down to the mouth on the east side of the Big Green inlet. This entire stretch displayed a much defined change of landscape and habitats -beginning marsh-like then moving to forested/sandy to forested/cobbled. The stream started out wide and deep, becoming narrower in the forested areas. When the cobbles began, the stream slowly became shallower, while the landscape became much steeper. In the steepest areas, the stream transformed from runs to riffles while along having dispersed pools.

While the landscape changed, as did the flora. In the open areas, there was once again a dominating population of Reed Canary Grass. The downstream flatter portion of the creek was populated with mainly Willow and Ash while the upstream steeper portion was made up of primarily Sugar Maple, and some Basswood. There were also areas of some Buckthorn and Honeysuckle shrubs.

### Maug Rd to Brooklyn G Rd (June 5, 2013)

The first point of this stretch was the corner between the two culverts on Maug and Skunk Hollow. This area was fairly altered from the culverts and being brushed out to open up the corner. The water coming out of the culvert has opened up the area and made a bit of a pool. Once across Skunk Hollow Rd, the substrate was mainly sandy and cobbled with many downed trees, woody cover, pools, and prevalent algal growth on cobbles. This stretch provided a varying canopy and directly related ground cover. The majority of the canopy was made up of Sugar Maple, Basswood, and Ash while the latter end held areas of White and Red Pine (sandier soils). The areas of more open canopy had dominant ground species of Reed Canary Grass and Bracken Fern while areas of denser canopy had less ground coverage, dominated by various dicots and forbes.

This portion's buffer was quite sloped and steep throughout, with an overall wider and shallower riffle than the previous portion. Near GPS point 151, the south (right) bank began to show signs of sandstone and became very steep. The sandstone has been eroded over the years and is overhanging the creek. The entire stretch exhibits signage of high water erosion and major flooding at times (garbage and human objects washed up on banks intermittently). At GPS 143, there is an inlet tributary stream on right bank. Nearer to Brooklyn G, there is left bank stress from human activity (mowed, flat bank; ATV trails).

### Brooklyn G Rd to Cty KK (June 6, 2013)

This stretch stared from the Brooklyn G culvert in a Basswood area. It was a shallow ripple and continued into an area with steeper buffers and evidence of human interaction. It remained fairly consistent until an area that was dammed (3 culverts present), also dug out upstream. This created a deeper section for roughly 600 to 1,000 feet. The sediment which was removed from the streambed was piled along the banks. This area had no canopy and was dominated by

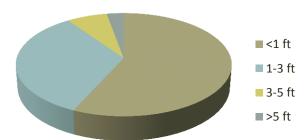
Reed Canary Grass. There was also macrophytes growth on the southern side. Following this area, the stream appeared to be ditched at some point, as it ran through ag. fields and was quite straight. The buffers in these areas were very small and little canopy was present. The substrate was consistently silt, with sporadic cobble appearing.

There stream ran through a couple different small forested areas. The substrate changed quickly here from silt to gravel, also transitioning into a riffle. In these areas, there also seemed to be a good deal of inlet drainages or ground water. The stream ventured out of these spots and became much more narrow and shallow. The canopy opened completely; Reed Canary Grass dominated the entire stretch. The stream split a few different times and transitioned into puddles, then dried up completely around GPS point 89, before Reed Rd. The stream remains dry grass runs until County Rd KK, where this is a channel (GPS 26) that holds water behind the farm. This is the only portion of the remainder of Dakin that holds water.



The stream visual assessment (SVAP) scores for Dakin Creek placed over 99 % of stream reaches in the Fair to Poor rating category.

Eighty-seven three (87%) percent of Dakin Creek has riparian buffers greater than 35 feet. Buffer width was measured from top of bank to outside edge. No distinction was made between native and non-native plants species for the purposes of scoring the buffer. Cropfields (hay) or maintained lawns (mowed) were not scored as buffers.



Over 90% of Dakin Creek's banks are less than 3 feet high.





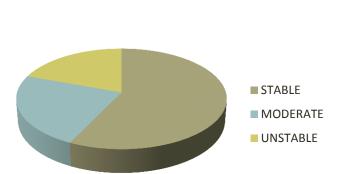
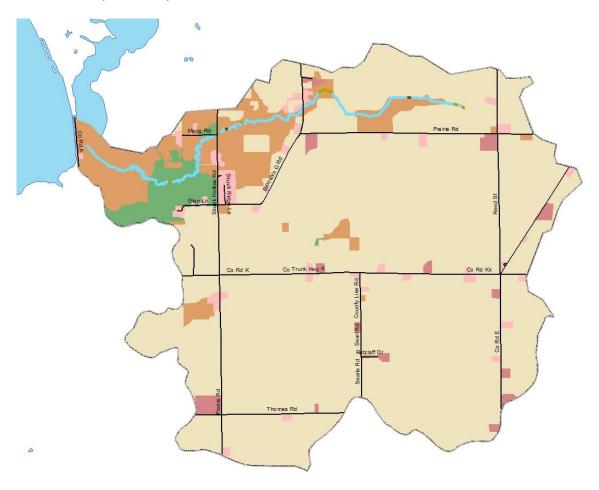


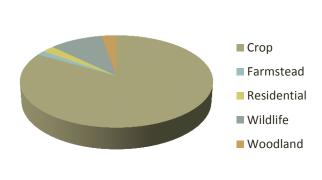


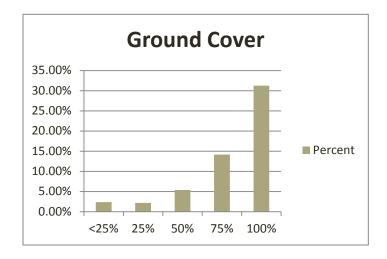
Photo by Jordan Resop

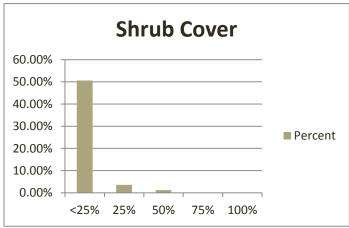
The majority of Dakin Creek's streambanks are either stable or moderately stable (80%), however nearly 20% were ranked as unstable, which could be a large source sediment and phosphorus delivery to the stream, and ultimately to Green Lake. Neither the amount of sediment delivery, nor the severity of the active erosion was detailed in this rapid assessment survey. Photographic records can provide additional details about the active erosion areas. Also, but combining bank stability data with other datasets, such as bank height and slope, may offer additional information about the sites. Detailed site inspections should be completed at these locations.











Dakin Creek's buffer area has a high percentage of ground cover, and a mixture of mature canopy trees. It most areas the buffer is lacking an understory of shrubs and young trees. This may be due to overgrazing by wildlife or the existence of a climax forest. Forest activities management could be implemented to create a more complex age structure within the buffer areas. A lack of shrubs to provide roots systems to anchor the soil, make bank vulnerable to future erosion, and mass wasting, especially in areas with high banks, and steep slopes.



Photo by Derek Kavanaugh

# Wisconsin DNR Information on Dakin Creek:

Watershed: UF07

Waterbody ID: 146700

Natural Community: Cool-Cold Headwater, Cool-Warm Headwater

Condition: Unknown

**Trout Water: No** 

Outstanding Water Resource: No.

Impaired Water: No.

Last Monitored: 2014

**Current Conditions:** 



Fish and Aquatic Life - Unknown



Fish Consumption - Unknown



Recreation - Unknown

## Phosphorus Monitoring\*:

Year	Mean (ug/L)	Min (ug/L)	Max (ug/L)	Threshold (ug/L)	
2011	53.5	<i>37</i>	203	75	
2014	46.5	35	108	75	
_	Monitoring locations are not the same				

Monitoring locations are not the same.

### MANDITORY STATE STREAM BUFFER

As part of this project the State of Wisconsin (WDNR), asked Green Lake County to provide "recommendations on the creation of a buffer ordinance". A number of states have implemented mandatory stream buffers. Some of the laws are very basic and do not address the actual extent of the issues, while others are very detailed and through. Before adopting any law, the State should consider how implementation and enforcement would occur. Below is a brief analysis how these laws would impact the tributaries of Green Lake if adopted.

### **PENNSYLVANIA**

Pennsylvania's law is very comprehensive, and well thought albeit, limited in applicability. The buffers width are significantly wider than the "traditionally" accepted width of 35, with mandatory widths ranging from 100-150 feet depending on the receiving waters. The coverage is limited in scope, because it is triggered by a permitted activity, meaning areas with buffers not meeting standards are not required to come into compliance until such an activity is undertaken. It does recognize the importance of forested riparian buffers, going as far as require the conversion or creation of new forested riparian buffers. It also specifically identifies native and noxious plants, and requires control and/or remove of invasive species within the buffers. The ordinance dictates the composition of the buffer, and requires an inspection and management plan, which is critical to the success of a quality effective buffer zone. A very important element, unique to Pennsylvania's law, is the requirement to control stormwater into the buffer. Not only does it require the upland control of sediment, but addressed sheet flow into the buffer.

As with most comprehensive laws, it is followed by a long list of exemptions and waivers including roads, single family homes, forest harvesting, disturbance under an acres, etc.

Once buffers are established are this ordinance, they are protected by deed restriction in perpetuity.

If Wisconsin was to pursue a buffer ordinance, many elements contained in the Pennsylvania law would be important to protecting water quality, however, as written it would have very limited applicability to increase the quantity or quality of buffers within the Green Lake watershed.

#### Pennsylvania's law

- § 102.14. Riparian buffer requirements.
- (a) General requirements for mandatory riparian buffers.
- (1) Except as in accordance with subsection (d), persons proposing or conducting earth disturbance activities when the activity requires a permit under this chapter may not conduct earth disturbance activities within 150 feet of a perennial or intermittent river, stream, or creek, or lake, pond or reservoir when the project site is located in an exceptional value or high quality watershed attaining its designated use as listed by the Department at the time of application and shall protect any existing riparian buffer in accordance with this section.
- (2) Except as in accordance with subsection (d), persons proposing or conducting earth disturbance activities when the activity requires a permit under this chapter where the project site is located in an Exceptional Value or High Quality watershed where there are waters failing to attain one or more designated uses as listed in Category 4 or 5 on Pennsylvania's Integrated Water Quality Monitoring and Assessment report, as amended and updated, at the time of the application, and the project site contains, is along or within 150 feet of a perennial or intermittent river, stream, or creek, lake, pond or reservoir shall, in accordance with the requirements of this section do one of the following:
  - (i) Protect an existing riparian forest buffer.
  - (ii) Convert an existing riparian buffer to a riparian forest buffer.

- (iii) Establish a new riparian forest buffer.
- (b) Riparian forest buffer criteria. To qualify as a riparian forest buffer under this chapter, an existing, converted or newly established riparian forest buffer, whether mandatory or voluntary, must meet the following requirements related to composition, width and management:
- (1) *Composition.* A riparian forest buffer is a riparian buffer that consists predominantly of native trees, shrubs and forbs that provide at least 60% uniform canopy cover. An existing riparian forest buffer does not have to be altered to establish individual Zones 1 and 2 under subparagraph (iii). At a minimum, it must have a total aggregate width of the combined zones under paragraph (2).
- (i) Existing riparian buffer conversion to a riparian forest buffer. Riparian buffers that consist predominantly of native woody vegetation that do not satisfy the composition of this paragraph or the width requirements in paragraph (2) shall be enhanced or widened, or both, by additional plantings in open spaces around existing native trees and shrubs that provide at least 60% uniform canopy cover. An existing riparian forest buffer does not have to be altered to establish individual Zones 1 and 2 under subparagraph (iii). At a minimum, it must be a total aggregate width of the combined zones under paragraph (2). Noxious weeds and invasive species shall be removed or controlled to the extent possible.
- (ii) Riparian forest buffer establishment. On sites without native woody vegetation, a riparian forest buffer shall be established and be composed of zones in accordance with subparagraph (iii), and meet the width requirements in paragraph (2). Noxious weeds and invasive species shall be removed or controlled to the extent possible.
  - (iii) Zones.
- (A) *Zone 1.* Undisturbed native trees must begin at the top of the streambank or normal pool elevation of a lake, pond or reservoir and occupy a strip of land measured horizontally on a line perpendicular from the top of streambank or normal pool elevation of a lake, pond or reservoir. Predominant vegetation must be composed of a variety of native riparian tree species.
- (B) Zone 2. Managed native trees and shrubs must begin at the landward edge of Zone 1 and occupy an additional strip of land measured horizontally on a line perpendicular from the top of streambank or normal pool elevation of a lake, pond or reservoir. Predominant vegetation must be composed of a variety of native riparian tree and shrub species.
- (2) Average minimum widths.
- (i) Waters other than special protection. A total of 100 feet (30.5 meters), comprised of 50 feet (15.2 meters) in Zone 1 and 50 feet (15.2 meters) in Zone 2 for newly established riparian forest buffers established under subsection (e)(3) along all rivers, perennial or intermittent streams, lakes, ponds or reservoirs.
- (ii) *Special protection waters.* A total of 150 feet (45.7 meters), comprised of 50 feet (15.2 meters) in Zone 1 and 100 feet (30.5 meters) in Zone 2 on newly established riparian forest buffers along all rivers, perennial or intermittent streams, lakes, ponds or reservoirs in special protection waters (high quality and exceptional value designations).
- (iii) Average riparian forest buffer width. The average riparian forest buffer width shall be calculated based upon the entire length of streambank or shoreline that is located within or along the boundaries of the project site. When calculating the buffer length the natural streambank or shoreline shall be followed.
- (3) Management requirements. Existing, converted and newly established riparian forest buffers shall be managed in accordance with a riparian forest buffer management plan in paragraph (4) and will be protected in accordance with subsection (g).
- (4) Management plan. The riparian forest buffer management plan shall be a part of the PCSM Plan and include, at a minimum, the following:
- (i) A planting plan for converted or newly established riparian forest buffers that identifies the number, density and species of native trees and shrubs appropriate to geographic location that will achieve 60% uniform canopy cover.

- (ii) A maintenance schedule and measures for converted or newly established riparian forest buffers to ensure survival and growth of plantings and protection from competing plants and animals including noxious weeds and invasive species over a 5-year establishment period including activities or practices used to maintain the riparian forest buffer including the disturbance of existing vegetation, tree removal, shrub removal, clearing, mowing, burning or spraying in accordance with long-term operation and maintenance.
- (iii) An inspection schedule and measures to ensure long-term maintenance and proper functioning of riparian forest buffers meeting the requirements in paragraph (1), including measures to repair damage to the buffer from storm events greater than the 2-year/24-hour storm.
- (c) Mandatory requirements for all riparian buffers.
- (1) Management of stormwater into the riparian buffer. Stormwater and accelerated erosion and sedimentation shall be managed in accordance with § § 102.4(b)—(e) and 102.8 (relating to erosion and sediment control requirements; and PCSM requirements) to ensure that stormwater enters the area upgrade and along the riparian buffer as sheet flow or shallow concentrated flow during storm events up to and including the 2 year/24 hour storm.
- (2) Wetlands. Wetlands located in the riparian buffer shall be protected and maintained consistent with Chapter 105 (relating to dam safety and waterway management).
- (3) *Measurements*. Riparian buffers must be measured horizontally and perpendicularly to the bank with no more than a 10% variation below the minimum width from the normal pool elevation for lake, pond or reservoir and from top of streambank.
- (d) Exceptions.
- (1) Subsection (a) does not apply for earth disturbance activities associated with the following:
- (i) A project site located greater than 150 feet (45.7 meters) from a river, stream, creek, lake, pond or reservoir.
- (ii) Activities involving less than 1 acre (0.4 hectare) of earth disturbance.
- (iii) Activities when permit coverage is not required under this chapter.
- (iv) Activities when a permit or authorization for the earth disturbance activity required under this chapter was obtained, or application submitted prior to November 19, 2010.
  - (v) Road maintenance activities so long as any existing riparian buffer is undisturbed to the extent practicable.
- (vi) The repair and maintenance of existing pipelines and utilities so long as any existing riparian buffer is undisturbed to the extent practicable.
- (vii) Oil and gas, timber harvesting, or mining activities for which site reclamation or restoration is part of the permit authorization in Chapters 78 and 86—90 and this chapter so long as any existing riparian buffer is undisturbed to the extent practicable.
- (viii) A single family home that is not part of a larger common plan of development or sale and the parcel was acquired by the applicant prior to November 19, 2010.
- (ix) Activities authorized by a Department permit under another chapter of this title which contains setback requirements, and the activity complies with those setback requirements.

- (2) For earth disturbance activities associated with the following, the Department, or the conservation district after consultation with the Department, may grant a waiver from any of the requirements of subsections (a) and (b) upon a demonstration by the applicant that there are reasonable alternatives for compliance with this section, so long as any existing riparian buffer is undisturbed to the extent practicable and that the activity will otherwise meet the requirements of this chapter:
  - (i) The project is necessary to abate a substantial threat to the public health or safety.
  - (ii) Linear projects which may include pipelines, public roadways, rail lines or utility lines.
  - (iii) Abandoned mine reclamation activities that are conducted under Department authorization or permit.
- (iv) Projects of a temporary nature where the site will be fully restored to its preexisting condition during the term of the permit under this chapter.
- (v) Redevelopment projects which may include brownfields or use of other vacant land and property within a developed area for further construction or development.
- (vi) Projects for which compliance with subsection (a) or (b) is not appropriate or feasible due to site characteristics, or existing structures at the project site.
- (3) The applicant shall submit a written request for a waiver to the Department or the conservation district as part of the application for a permit under this chapter.
- (4) An applicant requesting a waiver may propose and the Department may allow offsite protection, conversion or establishment of riparian forest buffers or provide compensation to fund riparian forest buffer protection, enhancement or establishment.
- (5) Projects qualifying for an exception under this subsection are not relieved from compliance with other applicable requirements of this chapter or other laws administered by the Department.
- (e) Utilization of riparian forest buffers.
- (1) Antidegradation presumption. Except for riparian buffers protected under subsection (a)(1) or (d), a riparian forest buffer meeting the requirements of this section will prevent thermal impacts and is a nondischarge alternative. When included in an E&S Plan or PCSM Plan meeting the requirements of this chapter, the proposed earth disturbance activity will satisfy § § 102.4(b)(6) and 102.8(h), unless data or information provided or available to the Department during the permit application or authorization review process shows that the proposed earth disturbance activity will degrade water quality.
- (2) Trading or offsetting credits. Except for riparian buffers protected under subsection (a)(1) or (d) when protection of existing, or conversion, or the establishment of a riparian forest buffer which meets the requirements of this section and is above baseline regulatory requirements, credits may be available for trading or offsets in accordance with any procedures established by the Department or any regulations related to trading or offsetting developed under this title.
- (3) *Voluntary riparian forest buffer.* Persons that protect, convert or establish a new riparian forest buffer meeting the requirements of this section, may qualify for benefits under paragraph (1) or (2).
- (f) Activities within a riparian buffer.
- (1) The following practices and activities are prohibited within the riparian buffer:
- (i) Soil disturbance by grading, stripping of topsoil, plowing, cultivating or other practices except as allowed in paragraph (3)(i).

- (ii) Draining by ditching, underdrains or other drainage systems.
- (iii) Housing, grazing or otherwise maintaining animals for agricultural or commercial purposes.
- (iv) Storing or stockpiling materials.
- (v) Off-road vehicular travel.
- (2) The following practices and activities are allowable in the riparian buffer when authorized by the Department:
- (i) Construction or placement of roads, bridges, trails, storm drainage, utilities or other structures.
- (ii) Water obstructions or encroachments.
- (iii) Restoration projects.
- (3) The following practices and activities are allowed within the riparian buffer:
- (i) Activities or practices used to maintain the riparian buffer including the disturbance of existing vegetation, and tree and shrub removal, as needed to allow for natural succession of native vegetation and protection of public health and safety.
  - (ii) Timber harvesting activities in accordance with the riparian forest buffer management plan as part of the PCSM Plan.
  - (iii) Passive or low impact recreational activities so long as the functioning of the riparian buffer is maintained.
  - (iv) Emergency response and other similar activities.
  - (v) Research and data collection activities, which may include water quality monitoring and stream gauging.
- (g) Permanent protection of riparian buffers.
- (1) Existing, converted and newly established riparian buffers including access easements must be protected in perpetuity through deed restriction, conservation easement, local ordinance, permit conditions or any other mechanisms that ensure the long-term functioning and integrity of the riparian buffer.
- (2) For any existing or newly established riparian buffer, the boundary limits of the riparian buffer must be identified and clearly marked.
- (h) *Reporting.* Persons who protect an existing riparian buffer or convert or establish a riparian buffer in accordance with this section shall complete data forms provided by the Department and submit the forms to the Department or conservation district within 1 year of establishment or protection.

### **Authority**

The provisions of this § 102.14 issued under sections 5 and 402 of The Clean Streams Law (35 P. S. § § 691.5 and 691.402); sections 1917-A and 1920-A of The Administrative Code of 1929 (71 P. S. § § 510-17 and 510-20); and section 11(2) of the Conservation District Law (3 P. S. § 859(2)).

### **MINNESOTA**

The Minnesota Buffer rule can be admired for its simplicity. All lakes over 25 acres, and all perennial streams are required to have a 50 foot buffer. It is an attractive model for multiple reasons, it is standard across the board and it is easy to measure compliance without special knowledge or skills. It falls short in a couple of areas. There was no compliance and enforcement mechanisms put into place by the State, and an recent report by EWG found that only 18% of streams in southern Minnesota were meeting standards. Another shortfall is it does not address the composition or quality of the buffer areas. The law also fails to address stormwater in any capacity. It does allow for harvesting/farming/grazing of the buffer areas if maintained in permanent vegetation. This can be viewed as a positive or negative, depending on the situation. There have been various interpretations of the law's language, which has been interpreted that vegetated buffers are not required under law in many circumstances, leading to recent efforts by Minnesota's Governor to pass new legislation to strengthen the law.

While simplistic in design may be ideal for implementation and enforcement, it would be recommended that more sited specific buffer plans be implemented to impact water quality.

This law, as written with the current ambiguities would likely have little impact on the tributaries to Green Lake. With some language modifications and clarifications, it could be improved.

### MN chapter 6120

#### Subp. 7.

### Agricultural use standards.

The agricultural use standards for shoreland areas are contained in items A, B, C, and D.

- A. The shore impact zone for parcels with permitted agricultural land uses is equal to a line parallel to and 50 feet from the ordinary high water level.
- B. General cultivation farming, grazing, nurseries, horticulture, truck farming, sod farming, and wild crop harvesting are permitted uses if steep slopes and shore and bluff impact zones are maintained in permanent vegetation or operated under an approved conservation plan (Resource Management Systems) consistent with the field office technical guides of the local soil and water conservation districts or the United States Soil Conservation Service.
- **C.** Animal feedlots as defined by the Minnesota Pollution Control Agency, where allowed by zoning district designations, must be reviewed as conditional uses and must meet the following standards:
- (1) New feedlots must not be located in the shoreland of watercourses or in bluff impact zones and must meet a minimum setback of 300 feet from the ordinary high water level of all public waters basins.
- (2) Modifications or expansions to existing feedlots that are located within 300 feet of the ordinary high water level or within a bluff impact zone are allowed if they do not further encroach into the existing ordinary high water level setback or encroach on bluff impact zones.
- (3) A certificate of compliance, interim permit, or animal feedlot permit, when required by parts <u>7020.0100</u> to <u>7020.1800</u>, must be obtained by the owner or operator of an animal feedlot.
- D. Use of fertilizer, pesticides, or animal wastes within shorelands must be done in such a way as to minimize impact on the shore impact zone or public water by proper application or use of earth or vegetation.

#### Minnesota's law (continued)

Subp. 8.

#### Forest management standards.

The harvesting of timber and associated reforestation or conversion of forested use to a nonforested use must be conducted consistent with the following standards:

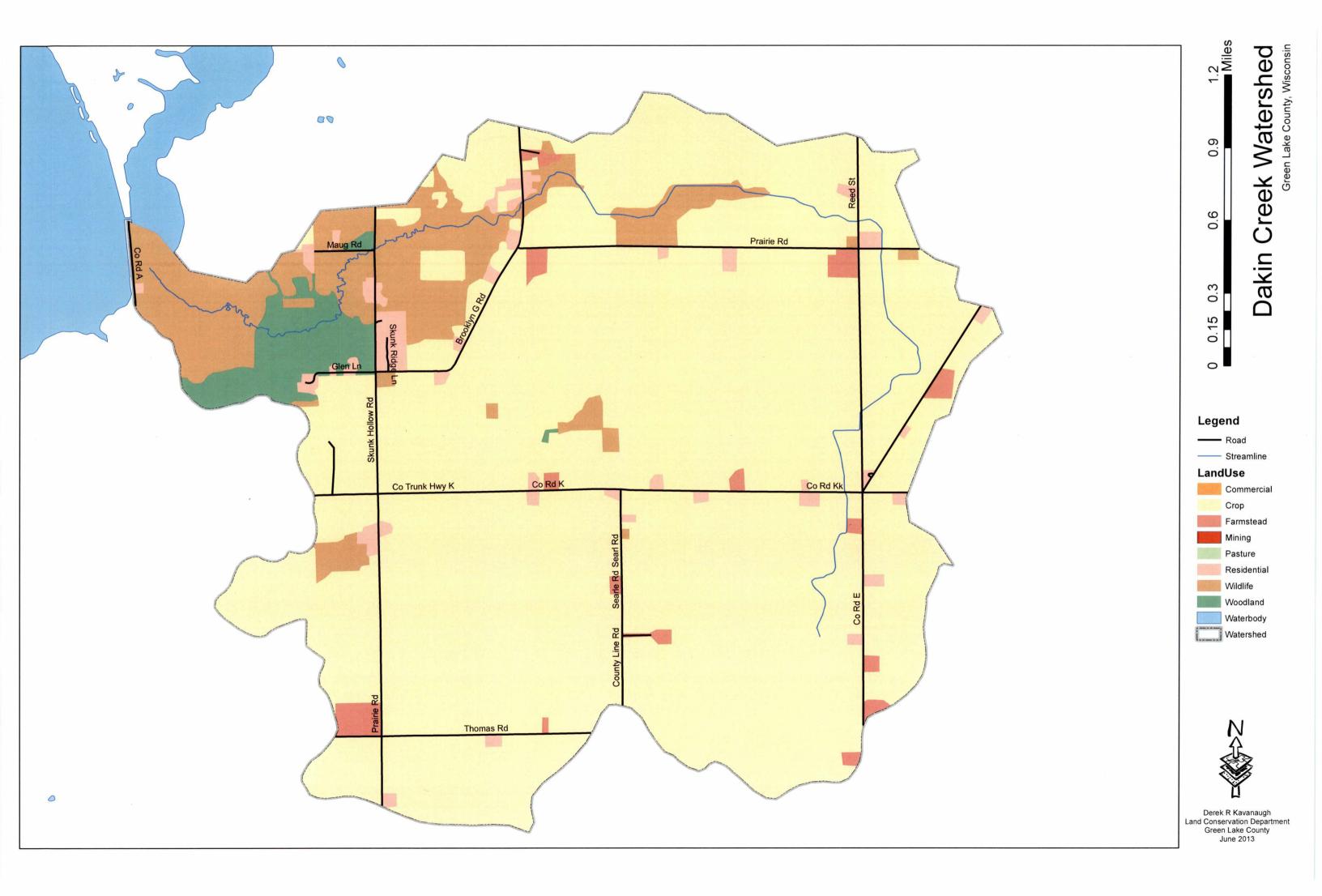
- A. Timber harvesting and associated reforestation must be conducted consistent with the provisions of the Minnesota Nonpoint Source Pollution Assessment-Forestry and the provisions of Water Quality in Forest Management "Best Management Practices in Minnesota."
- B. If allowed by local governments, forest land conversion to another use requires issuance of a conditional use permit and adherence to the following standards:
- (1) shore and bluff impact zones must not be intensively cleared of vegetation; and
- (2) an erosion and sediment control plan is developed and approved by the local soil and water conservation district before issuance of a conditional use permit for the conversion.
- C. Use of fertilizer, pesticides, or animal wastes within shorelands must be done in such a way as to minimize impact on the shore impact zone or public water by proper application or use of earth or vegetation.

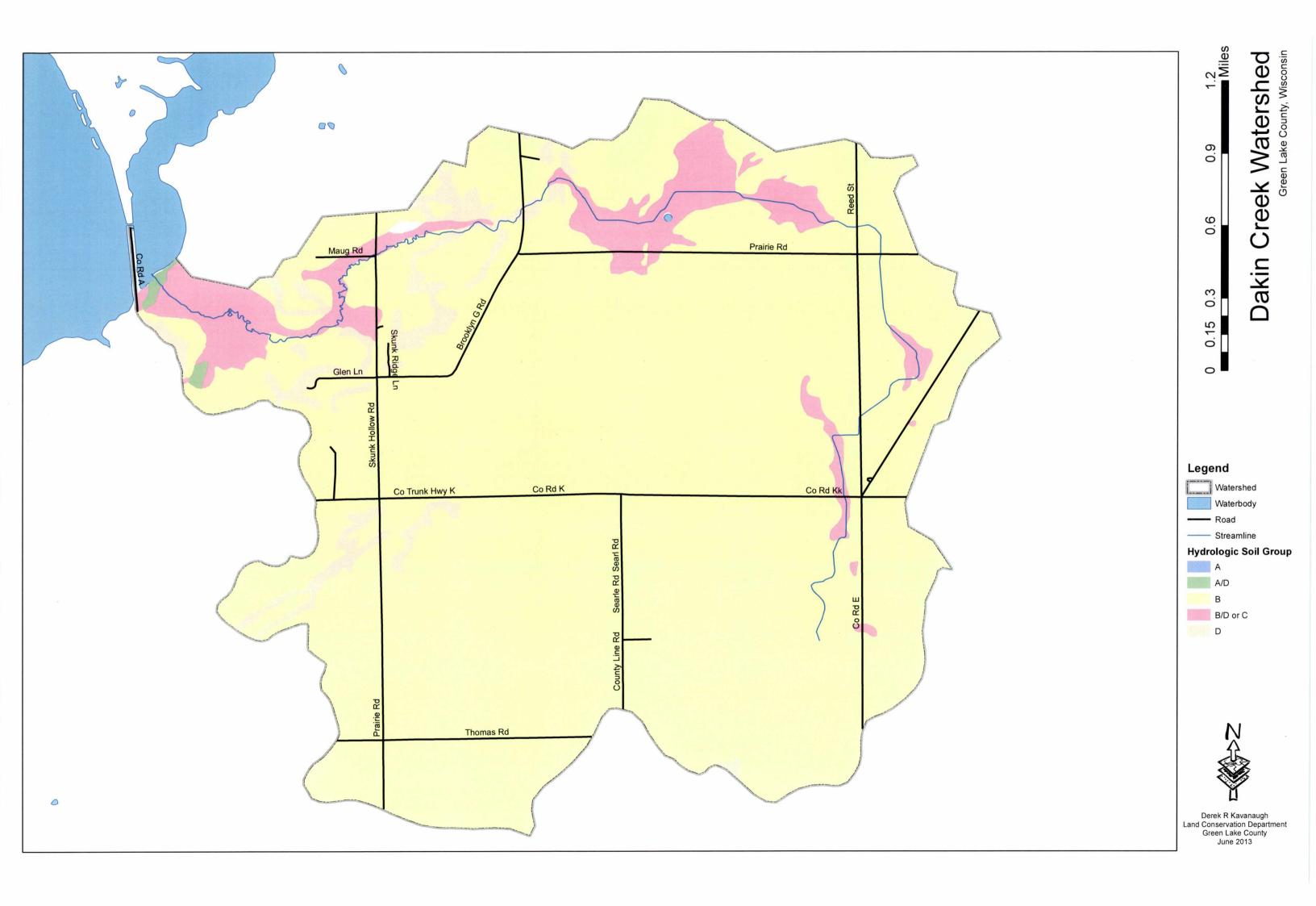
These are two examples on either end of the spectrum of possible legislative approaches. Either approach will require the commitment of thousands of hours and manpower for education, implementation, and enforcement. These factors should be factored into any designed program. Other approaches that have been implemented in various regions include voluntary programs, incentive programs, tax-incentive programs, easement programs, amongst others. Each has its merits and limitations that should be identified and recognized prior to implementation.

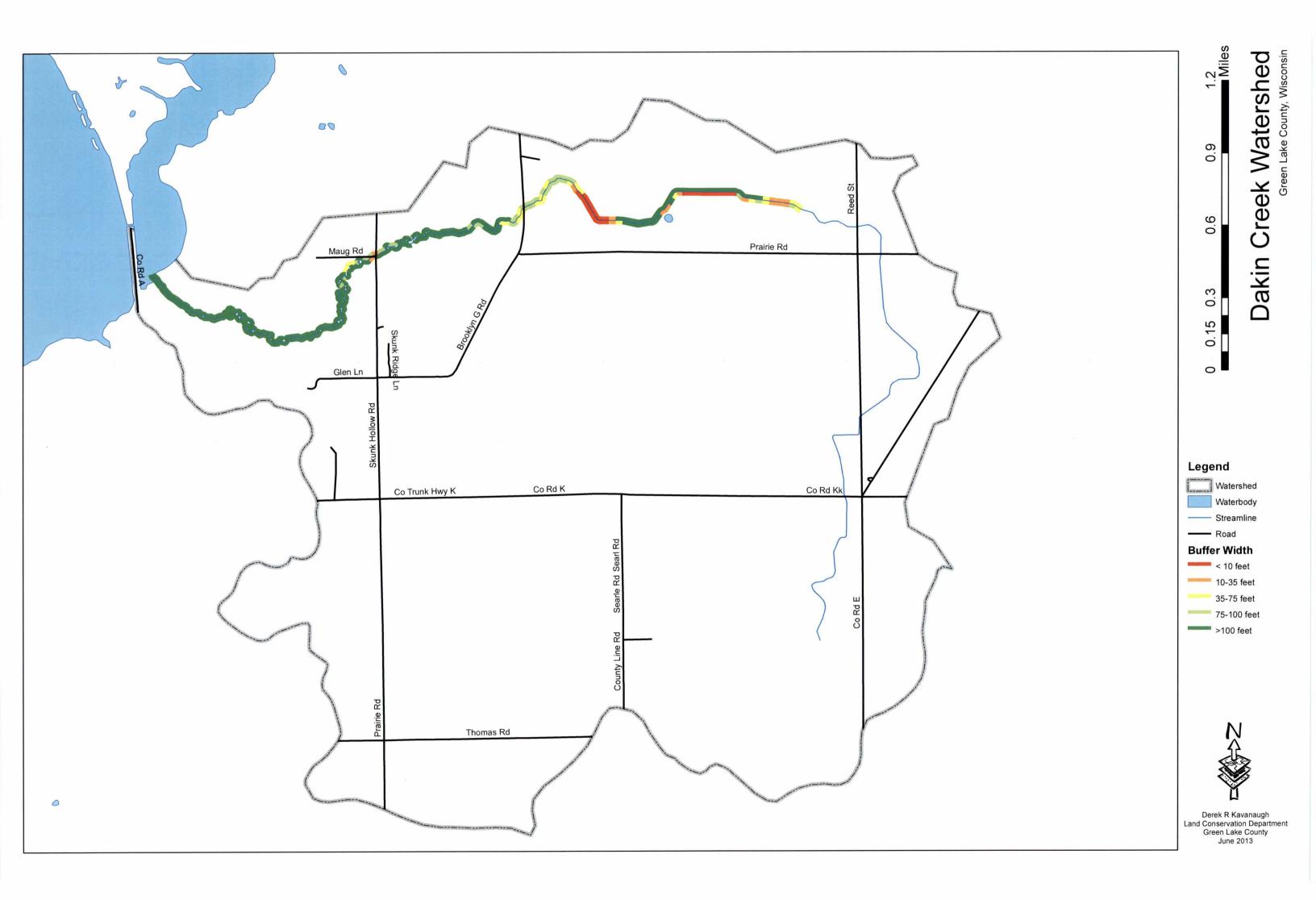
### **SUMMARY**

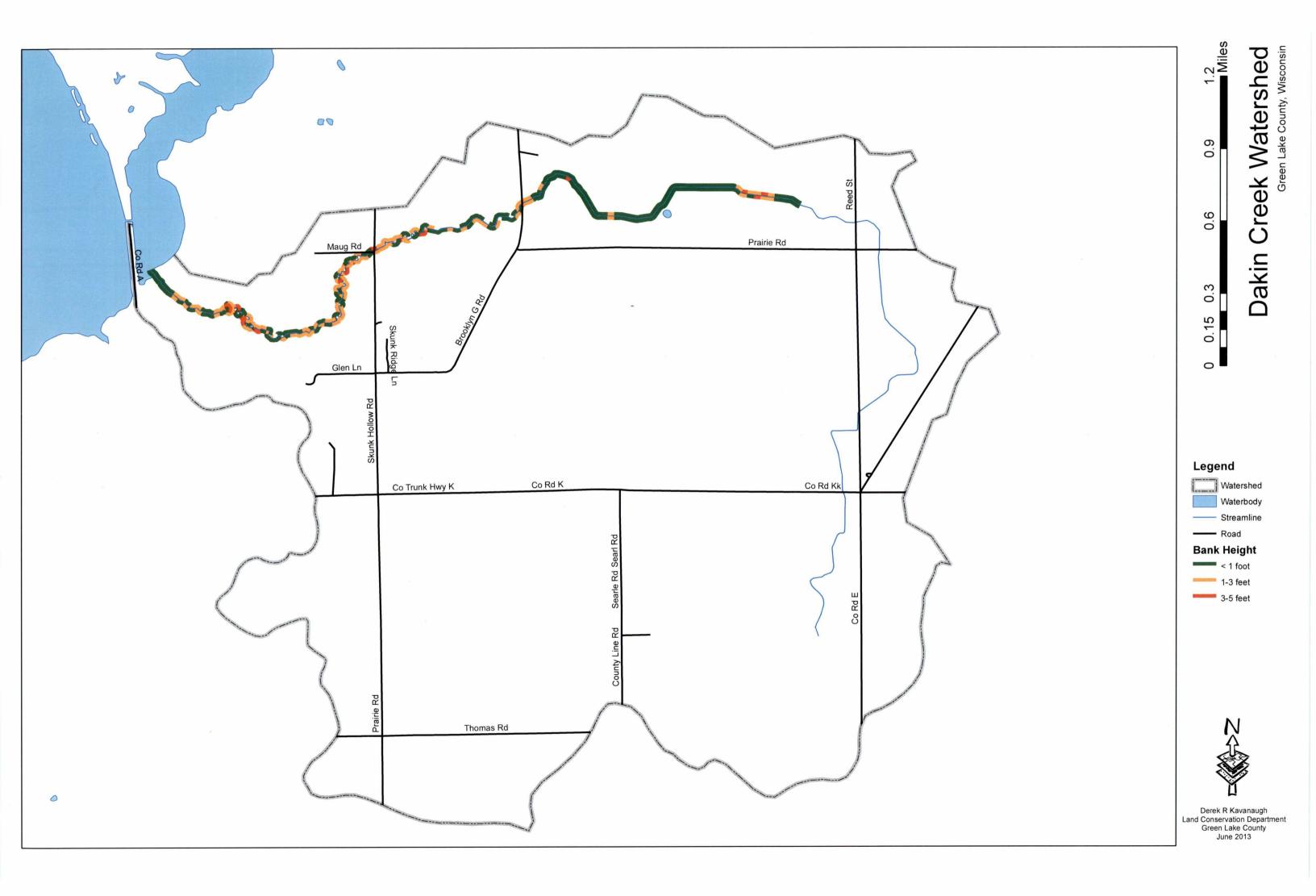
This information is not static, but ever changing, and regular monitoring and updates will be required to maintain its current status. This dataset is not intended to be used as a holistic solution to the water quality and runoff problems for Green Lake, but it intended to be used as a tool to aid in planning the future management of the watershed and riparian areas. With the power of GIS, information can be combined and compared to identify priority areas to target for conservation work. While some of the printed maps may provide some insight to what exists on the land, the true utility is contained within the ability to work with the raw data within GIS. ArcGIS shapefiles (\*.shp) are available from the Green Lake County Land Conservation Department for the data contained within this report.

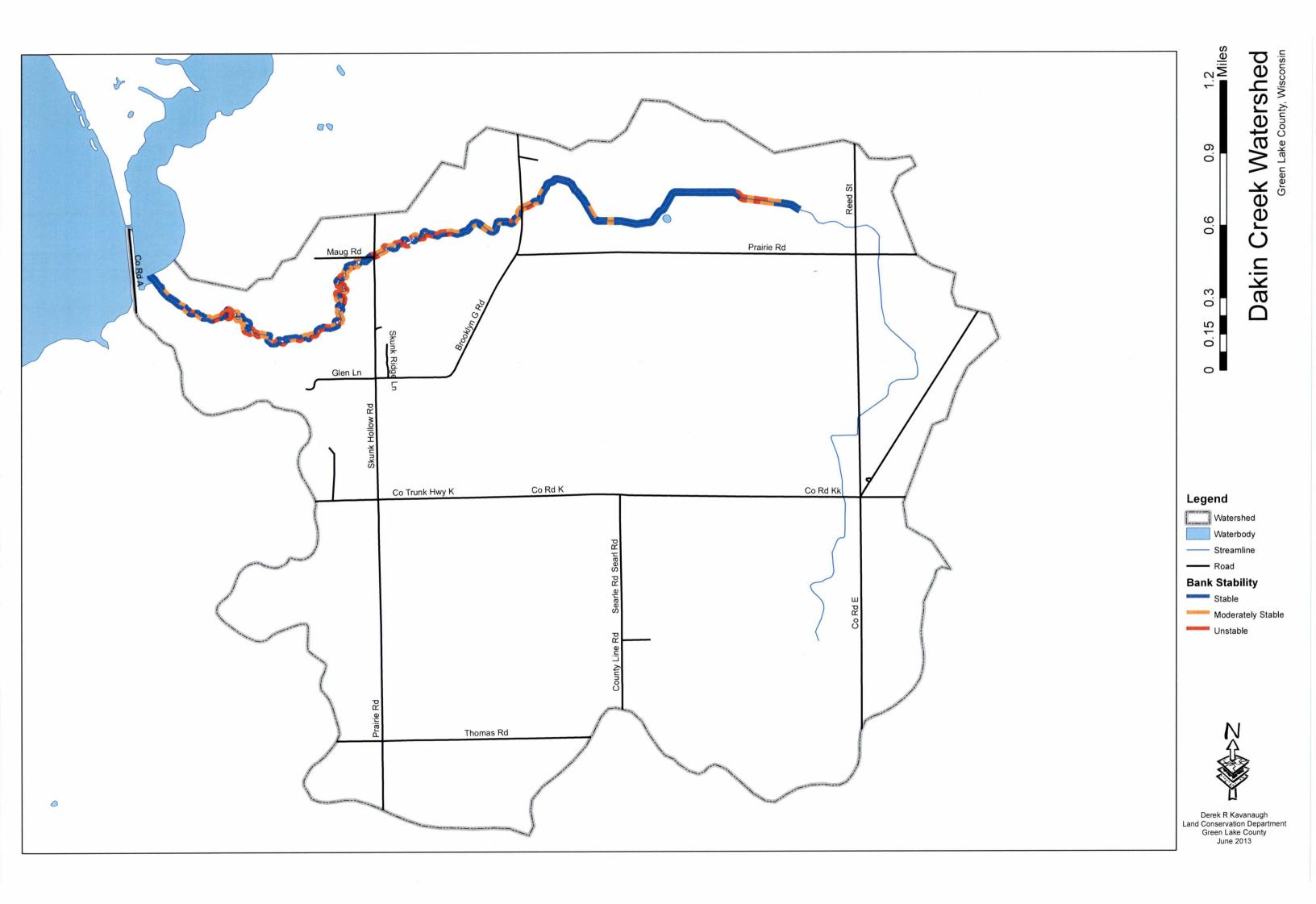


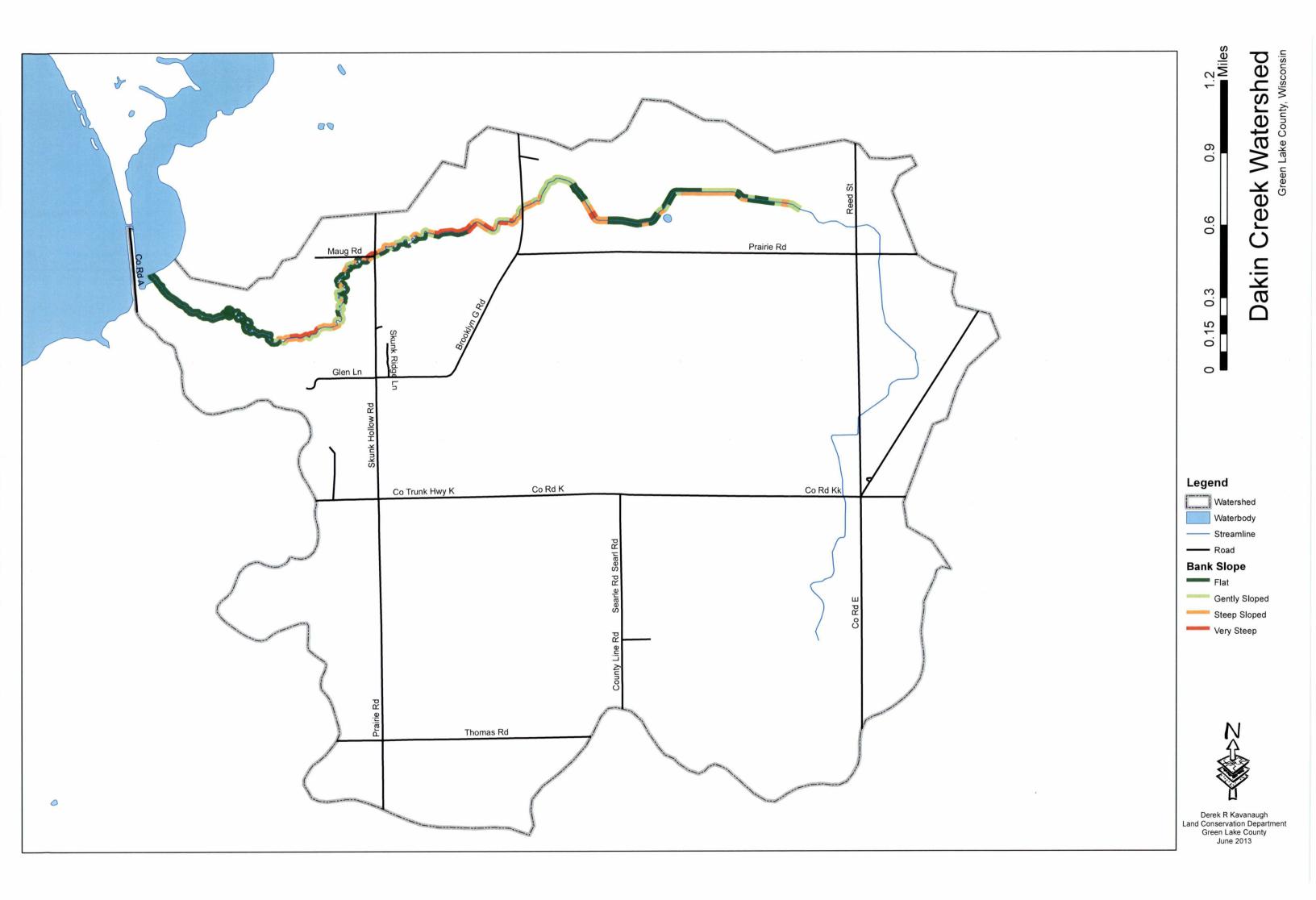


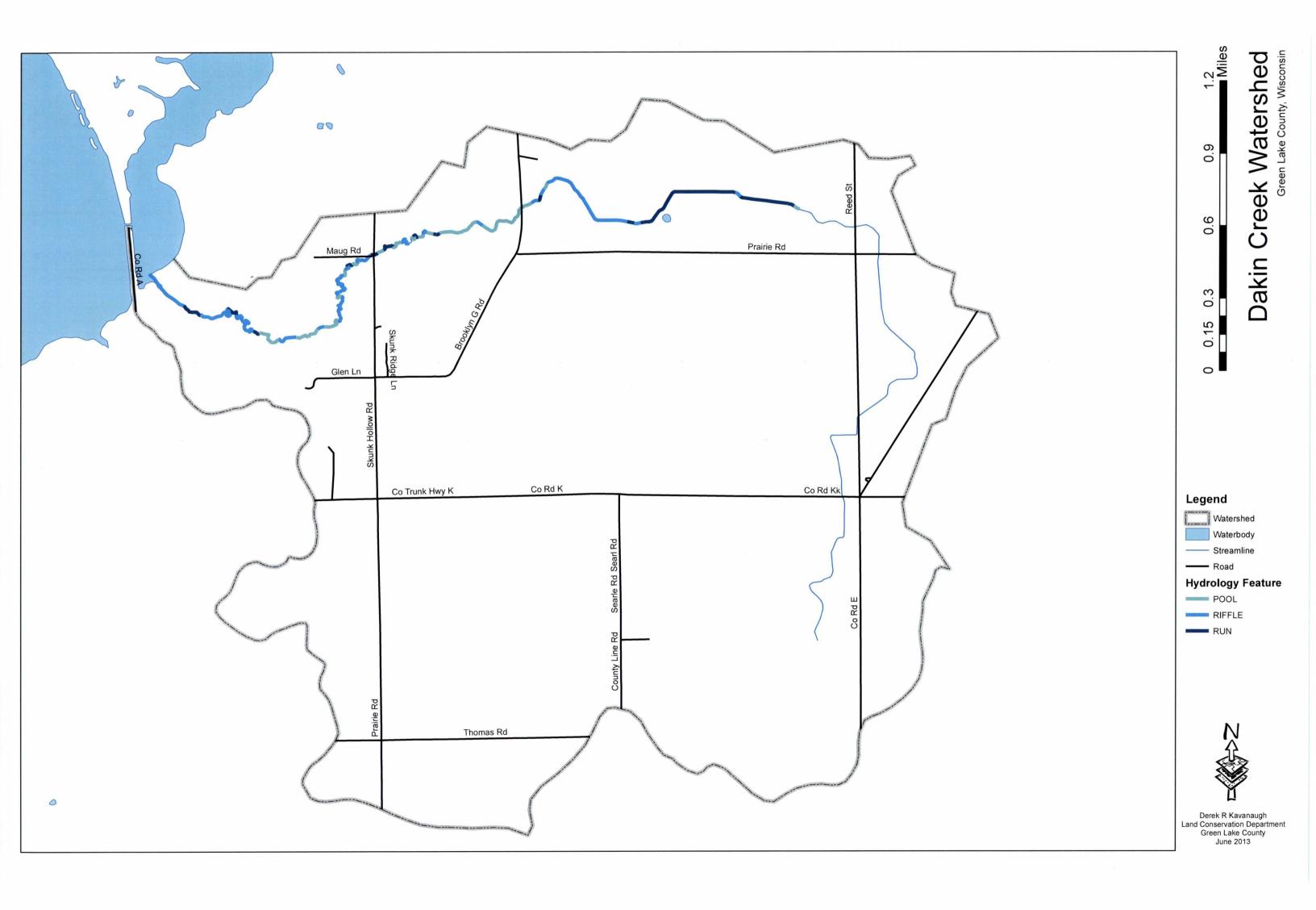


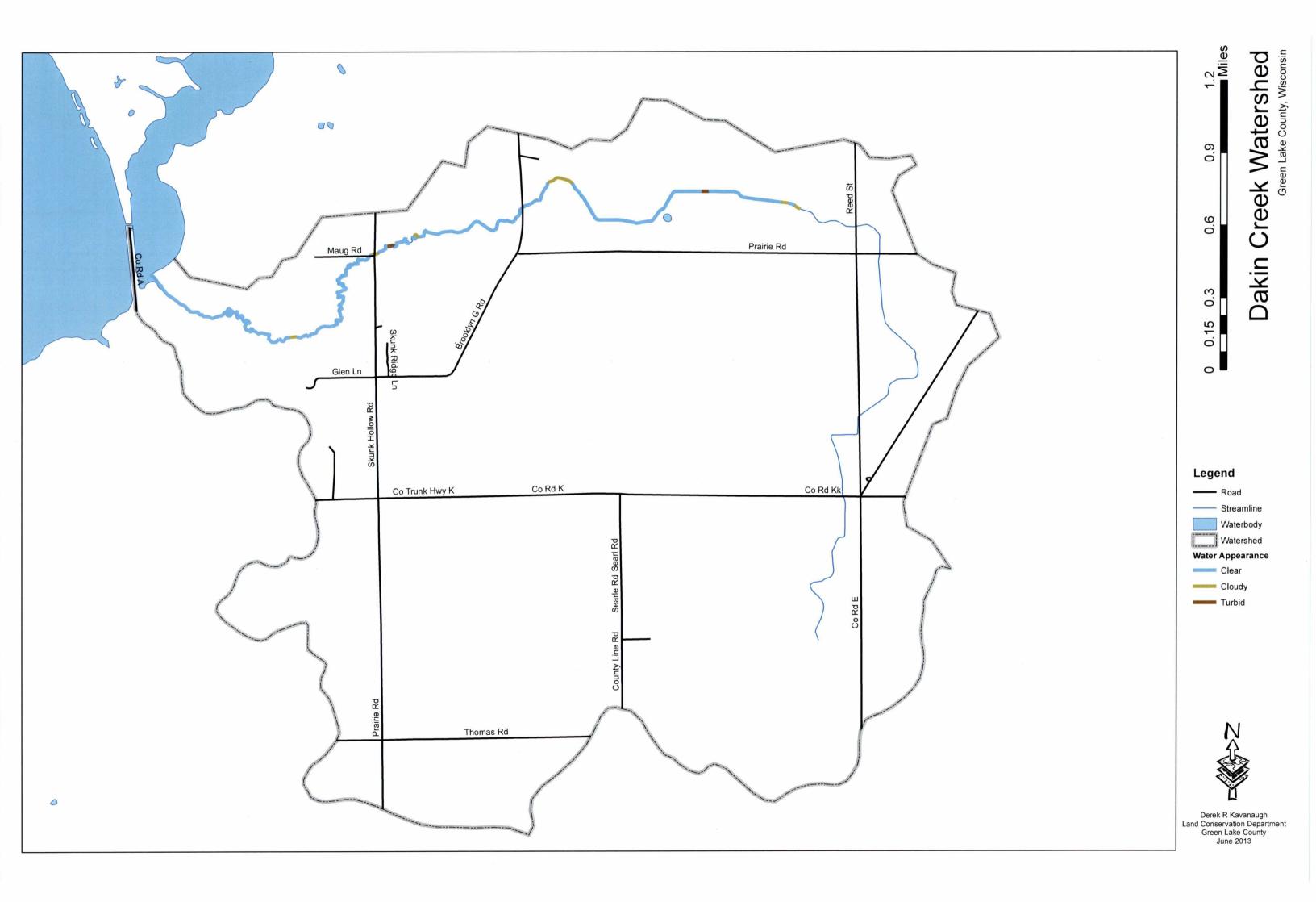


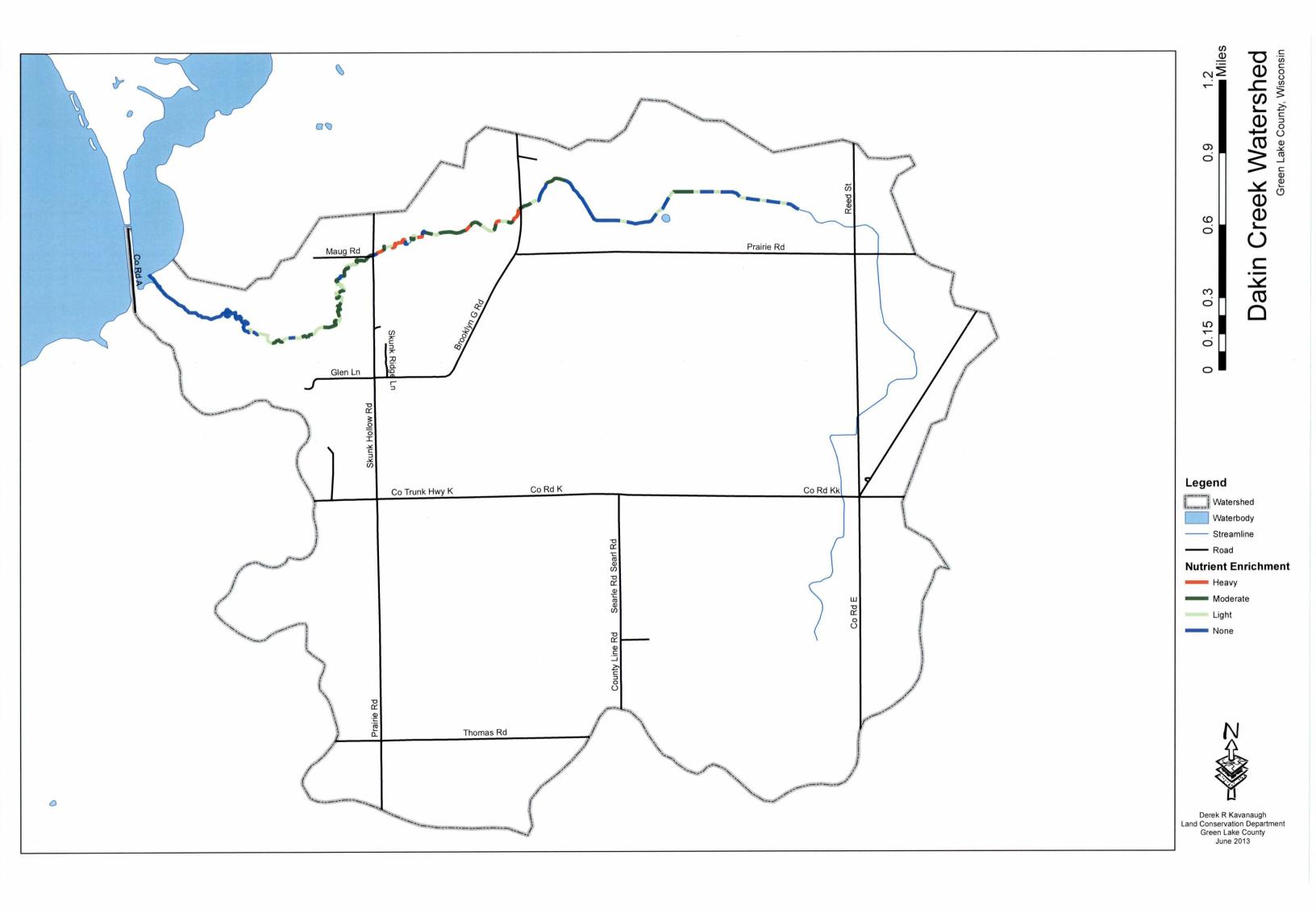


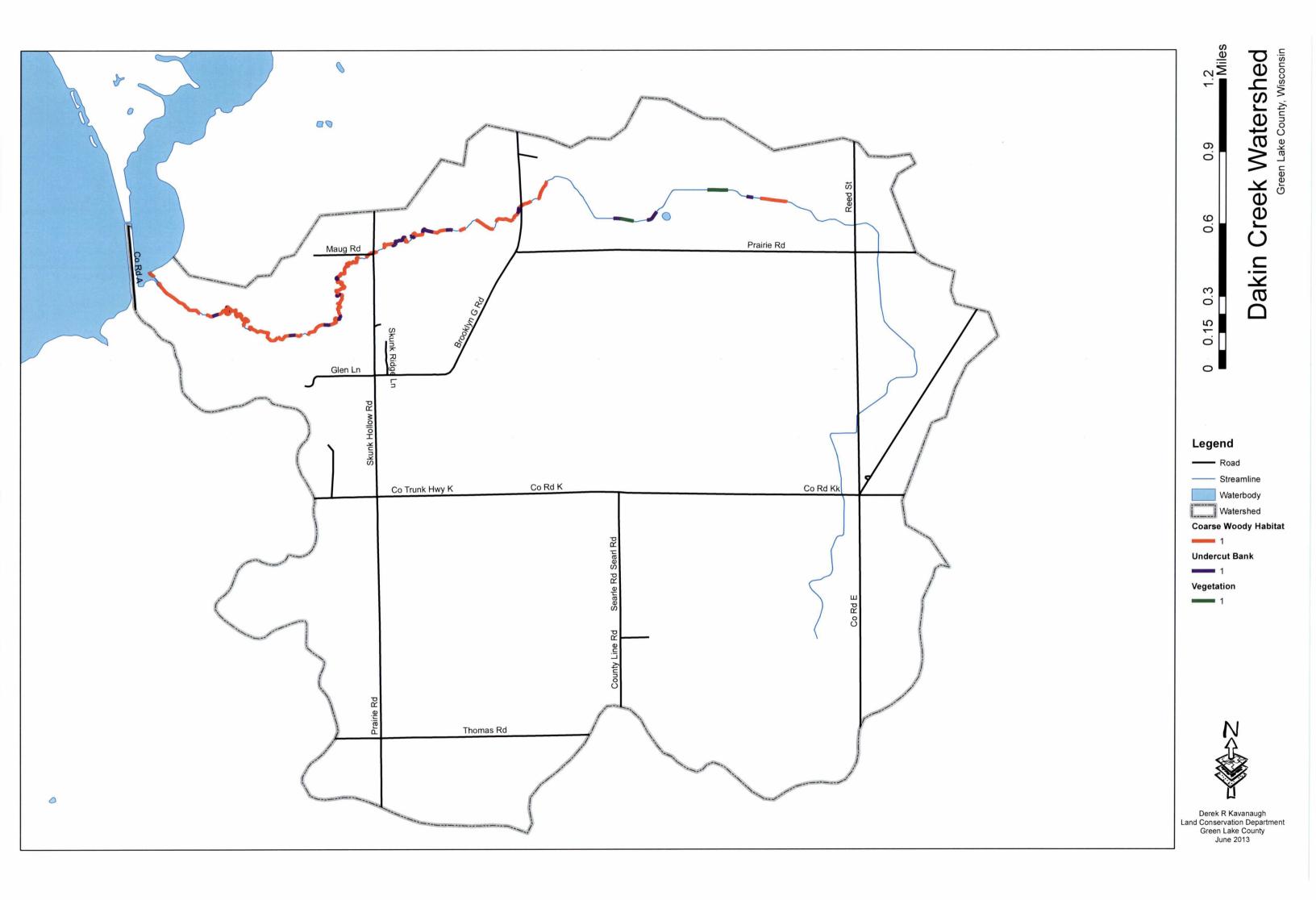












## Wuerches Creek Watershed

0.8 Miles

9.0

0.4

0 0.1 0.2

## Legend

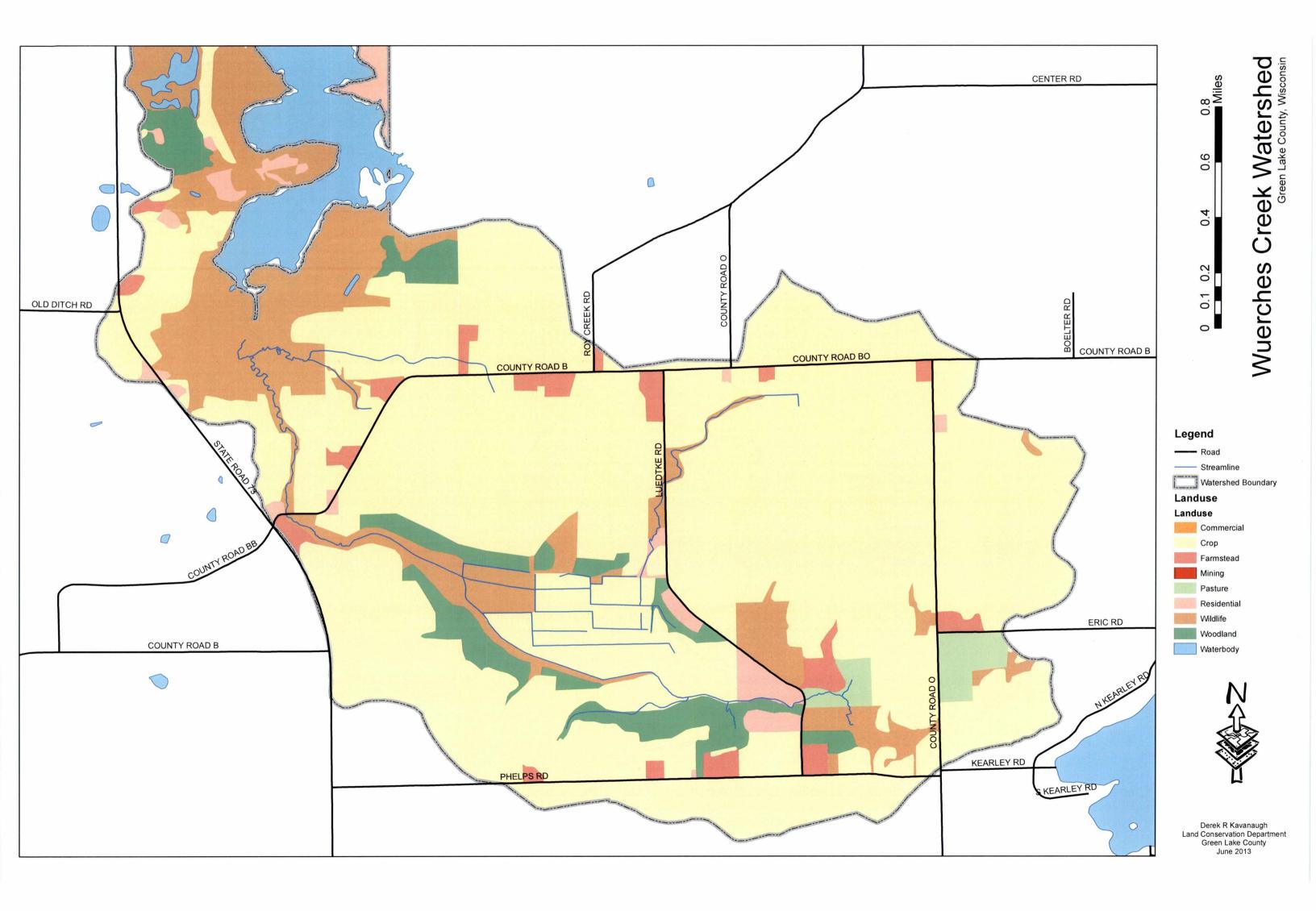
Road

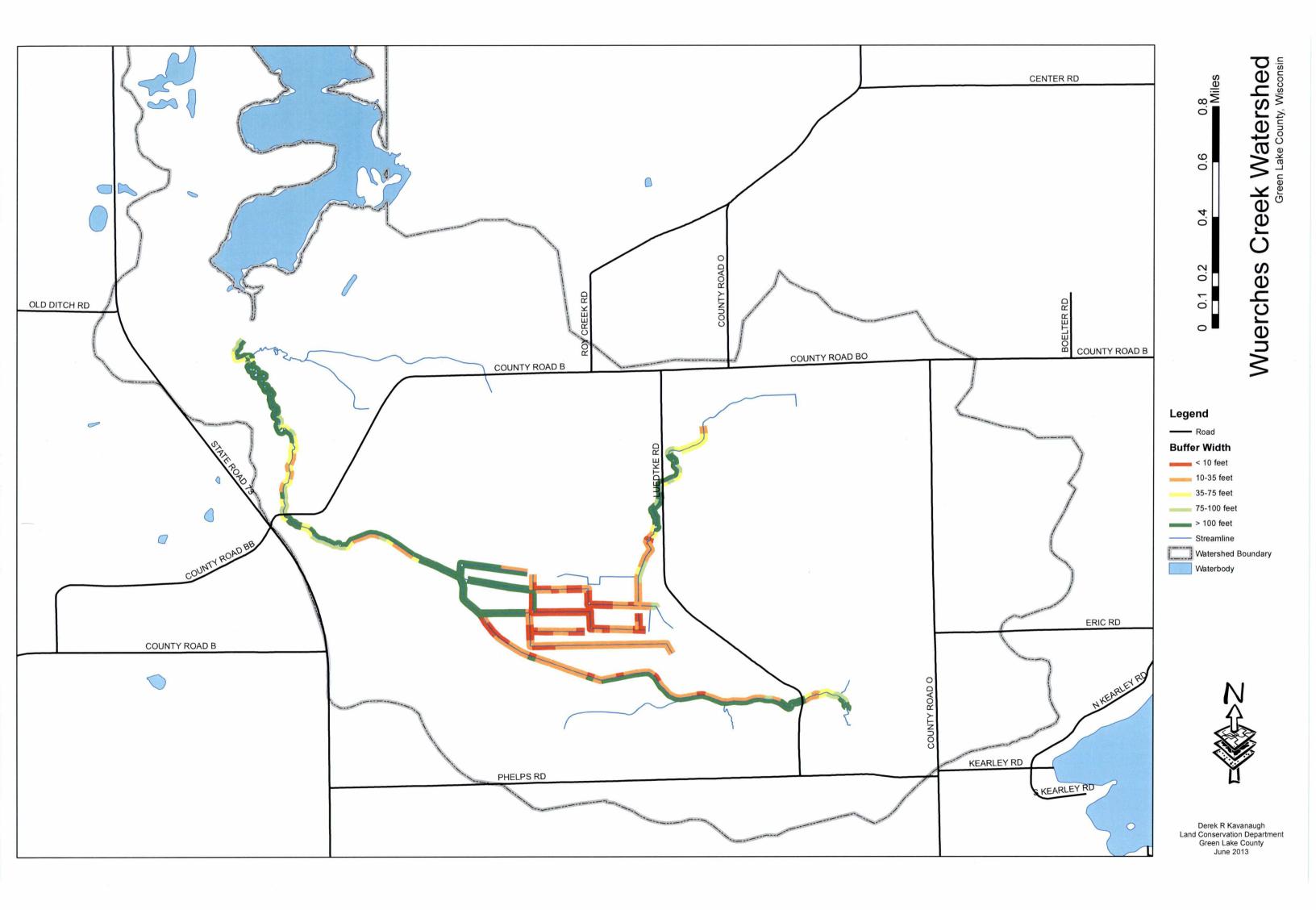
Streamline

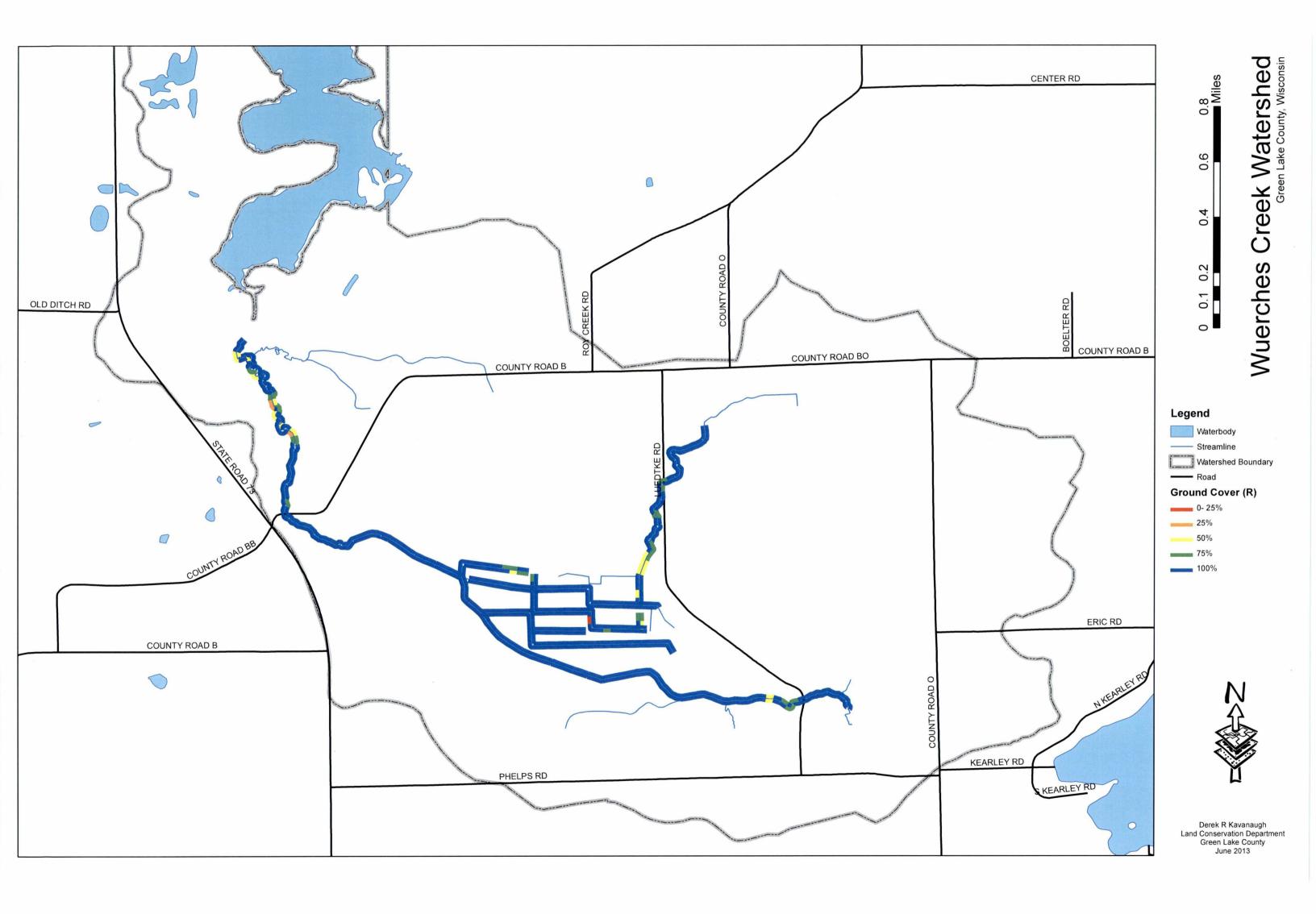
Watershed Boundary
Waterbody

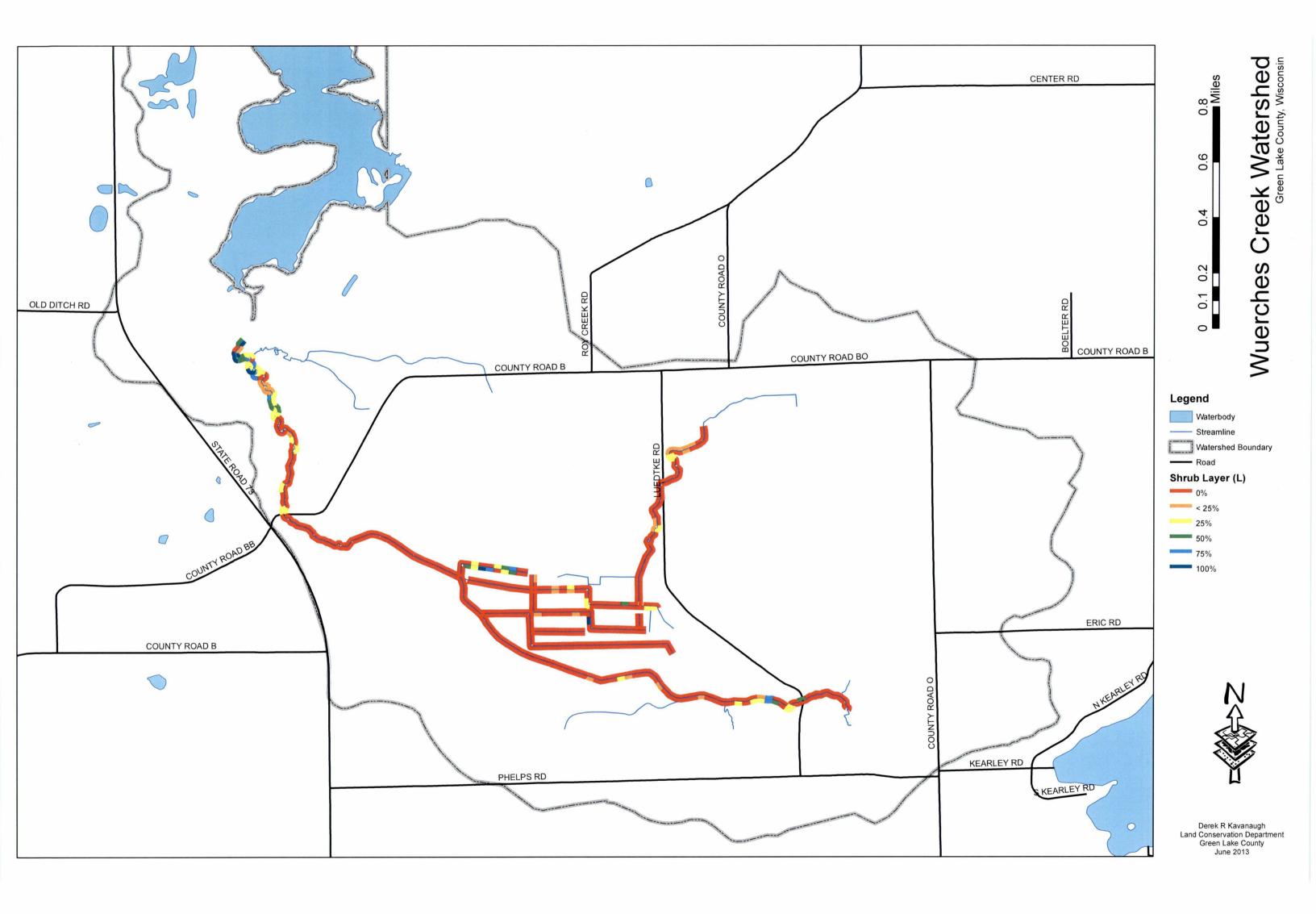


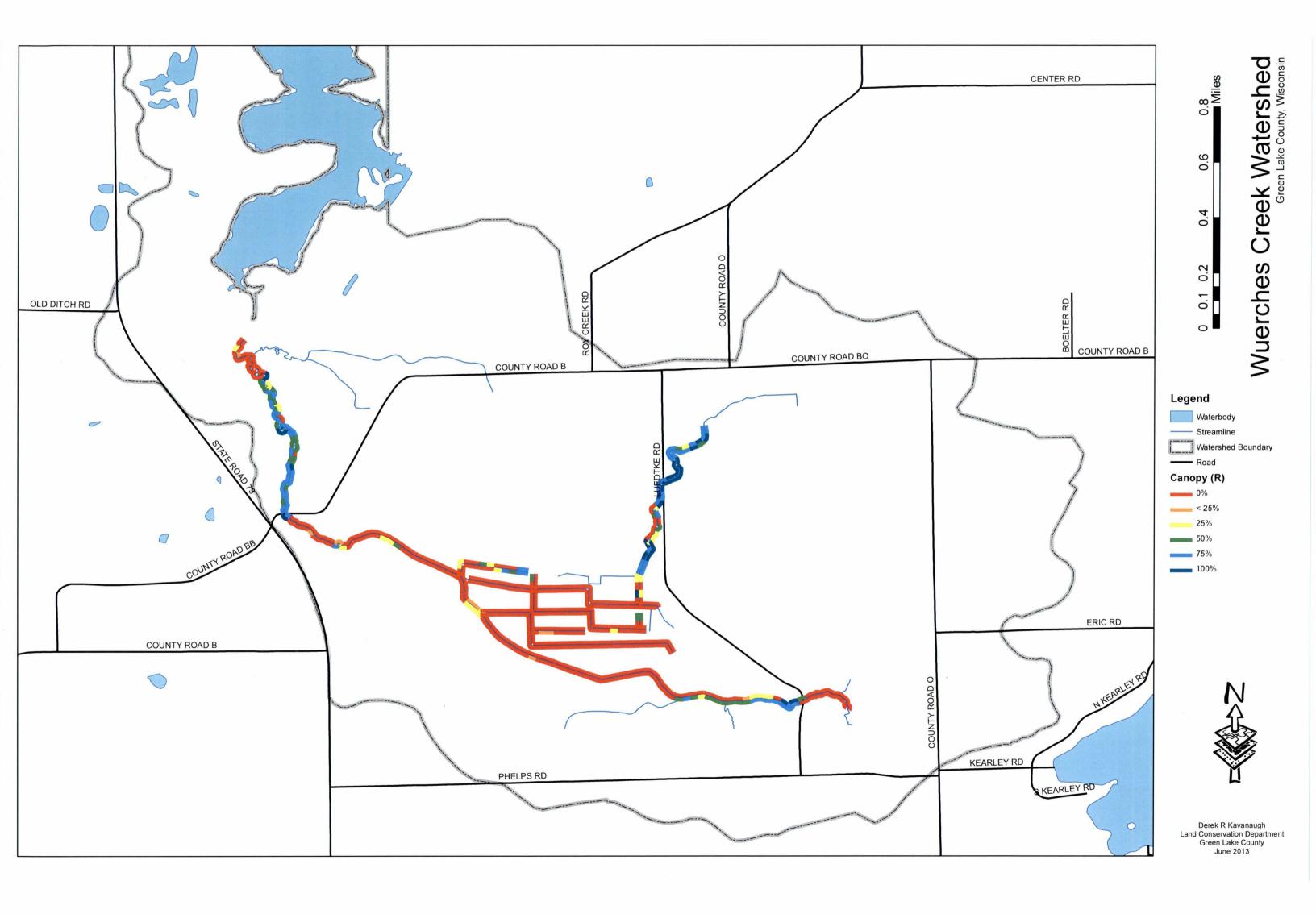
Derek R Kavanaugh Land Conservation Department Green Lake County June 2013

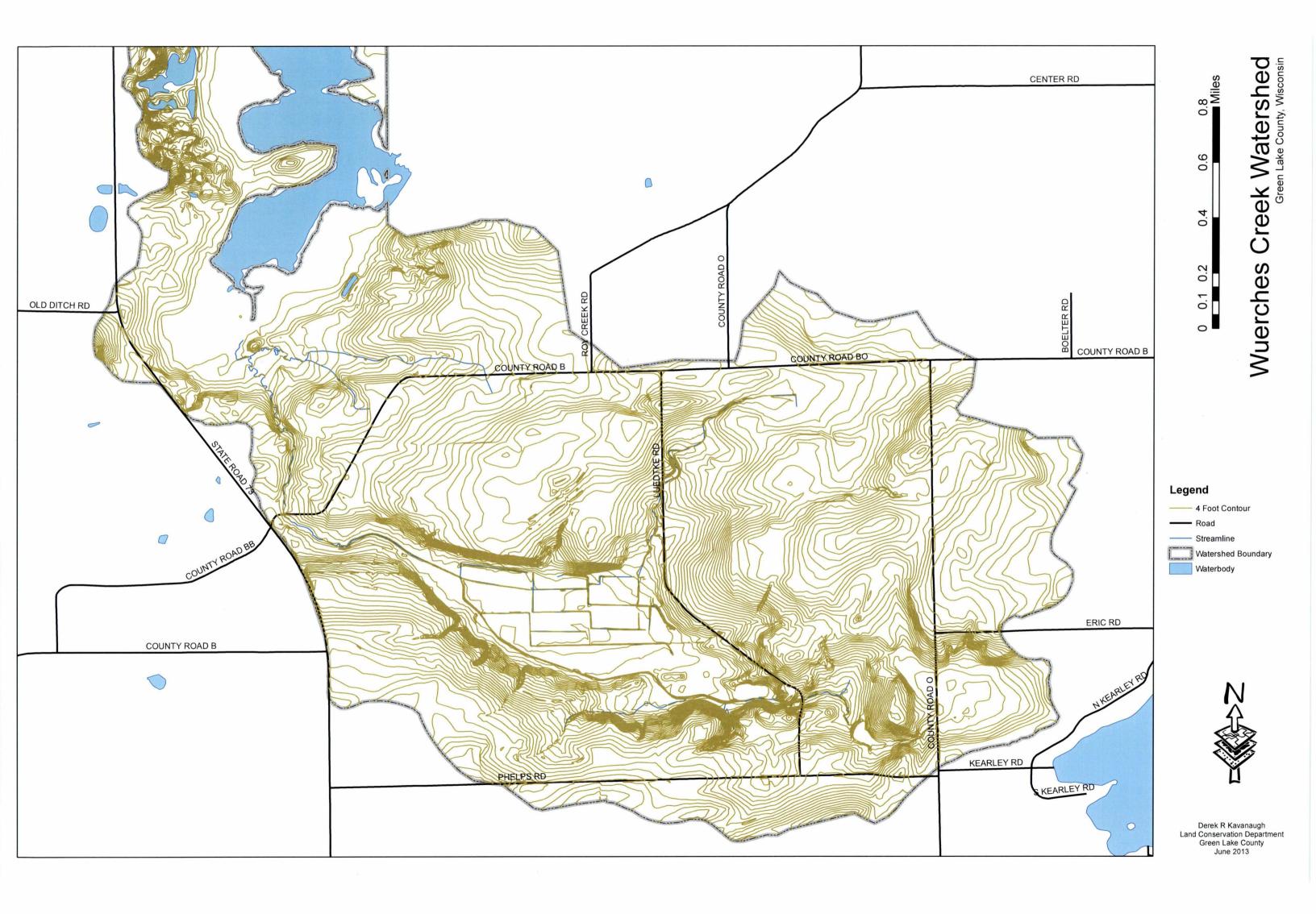


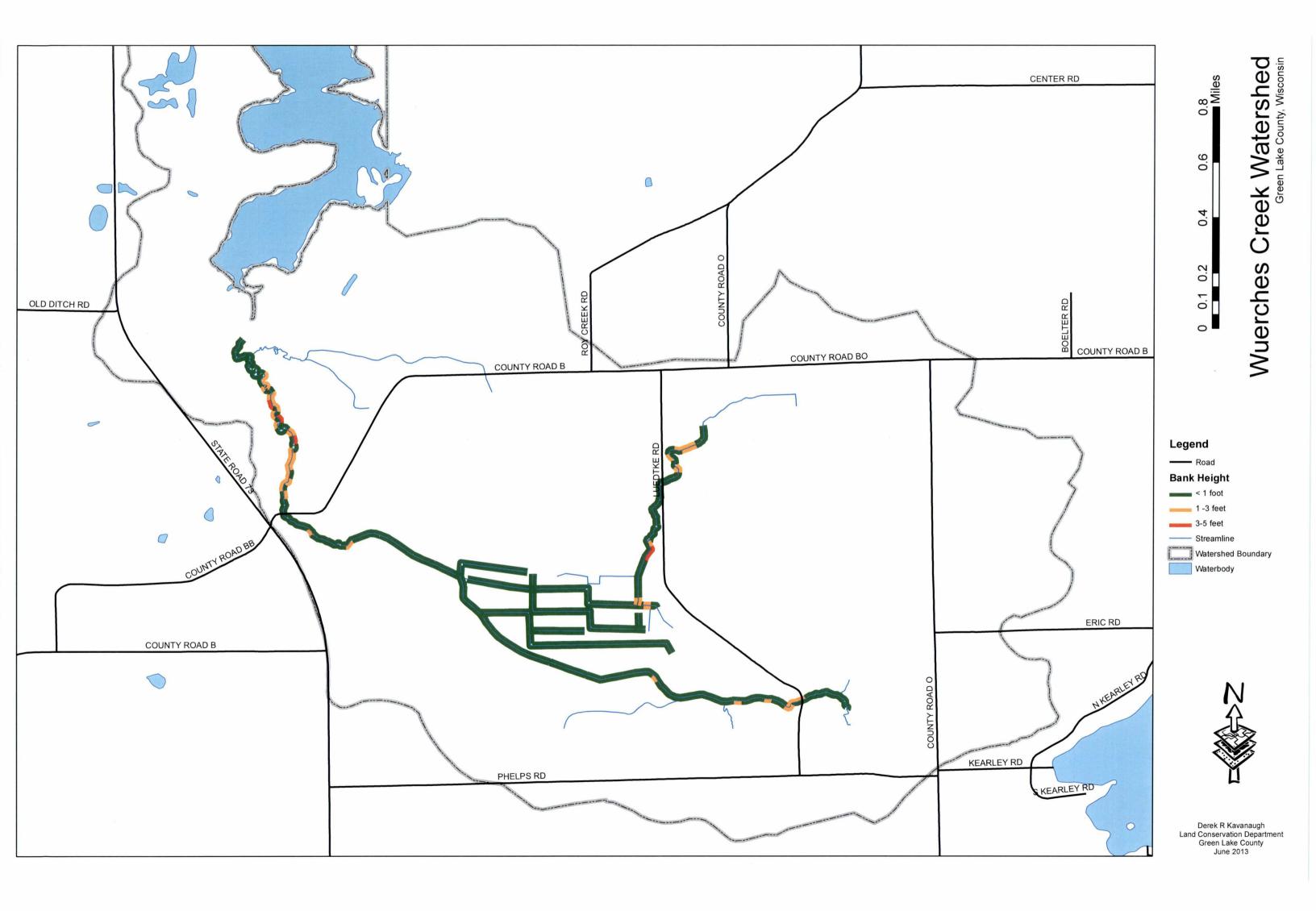


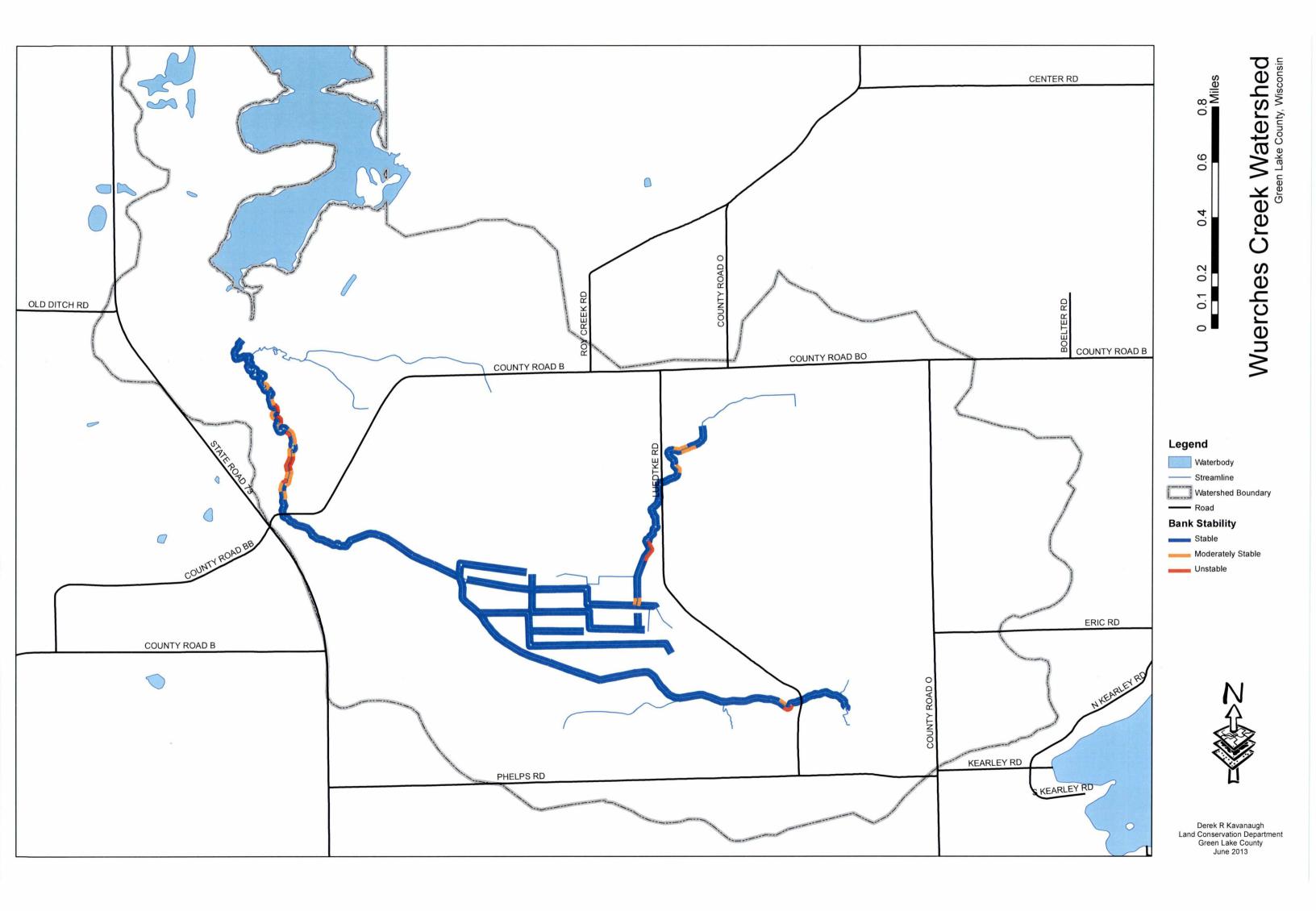


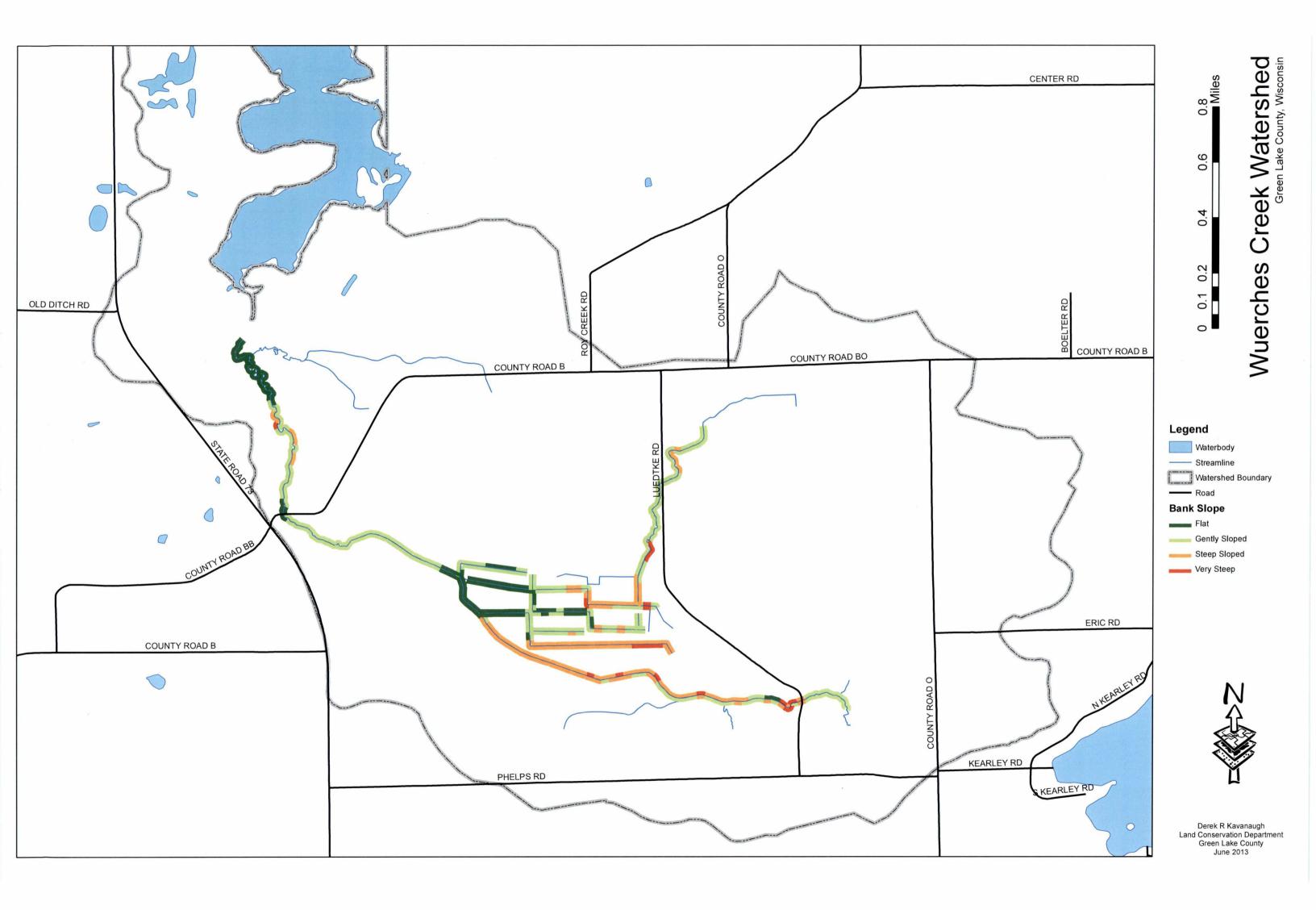


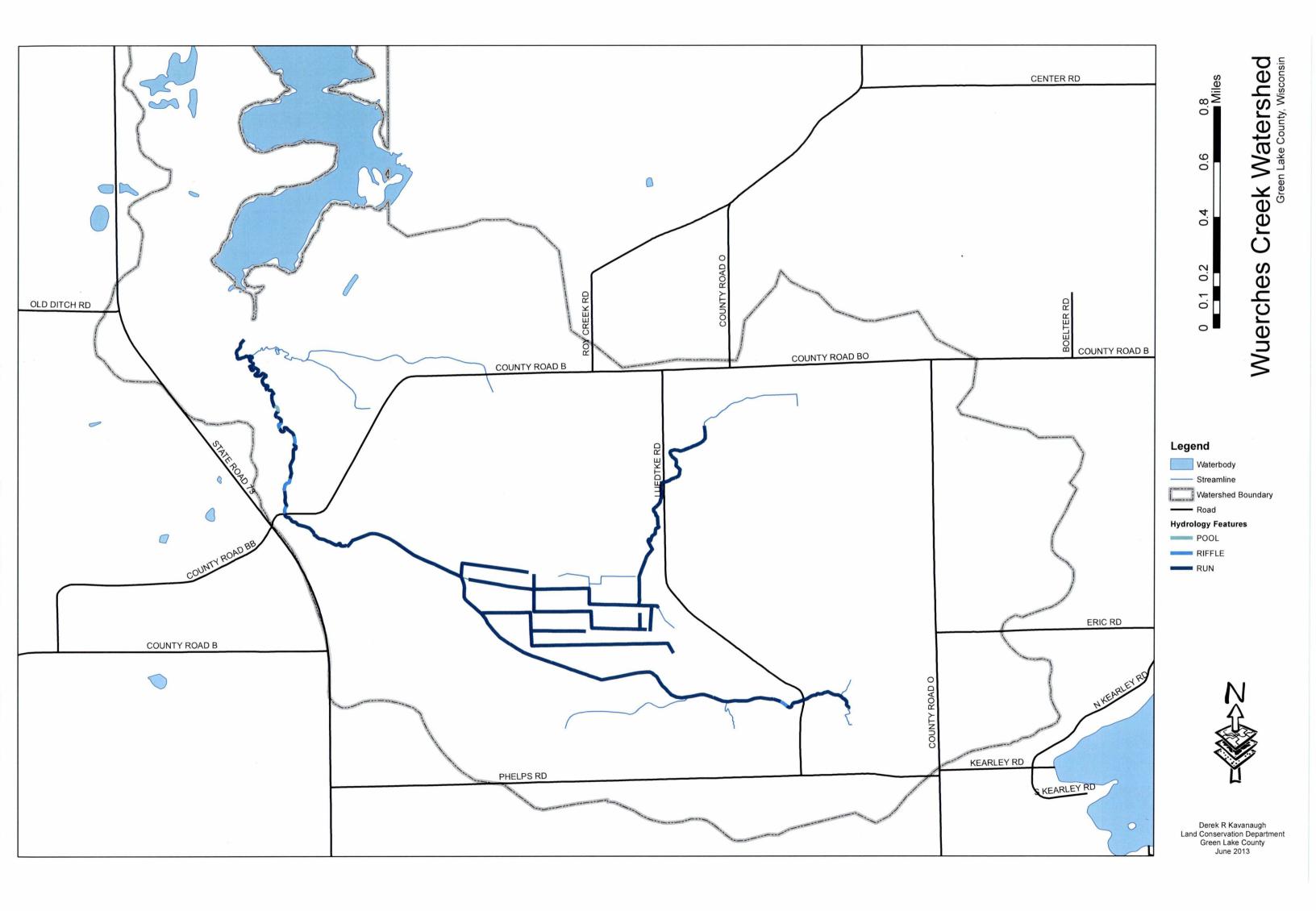


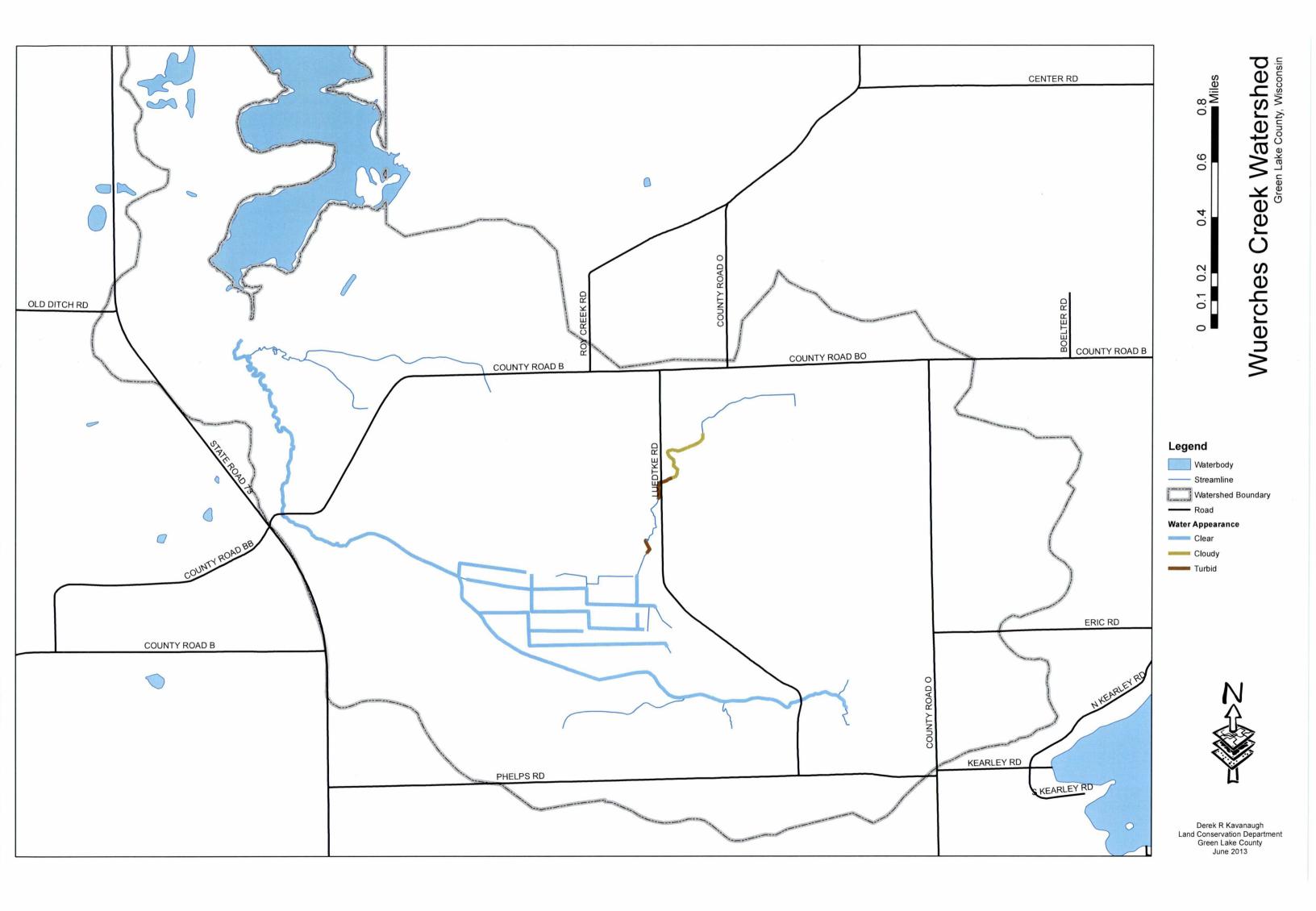


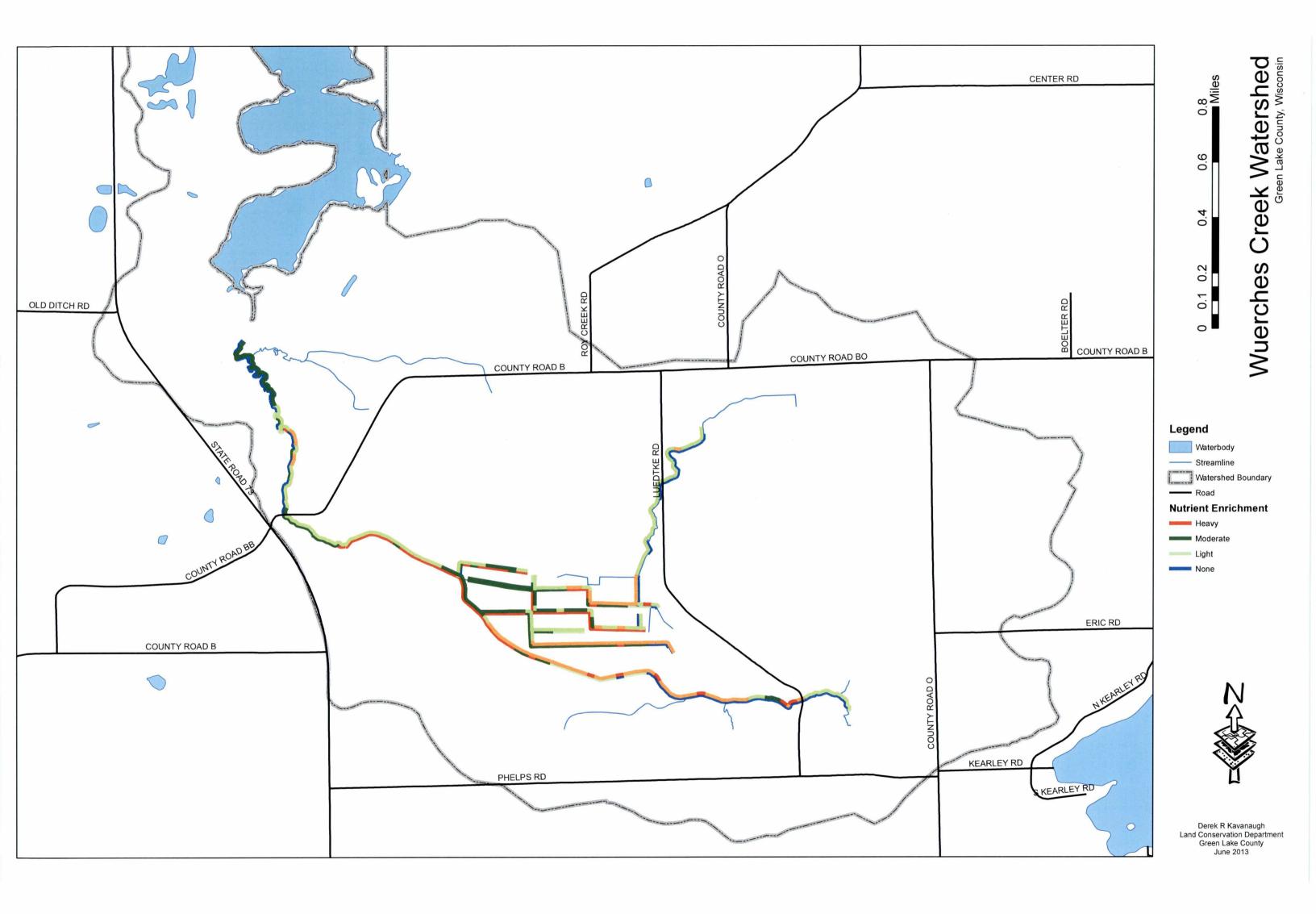


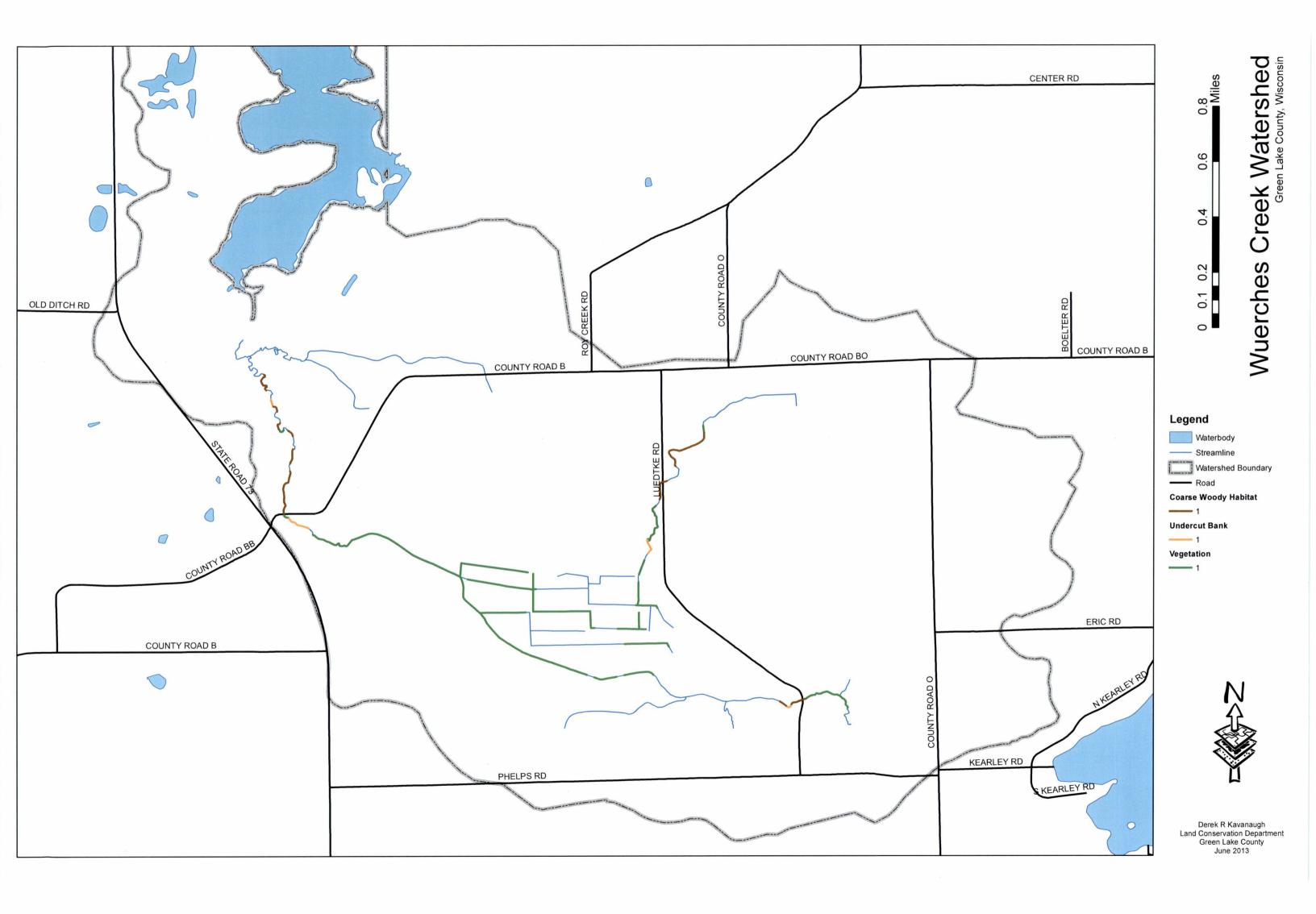














## Hill Creek Watershed Green Lake County, Wisconsin

Waterbody

Watershed Boundary

Landuse\_Hill

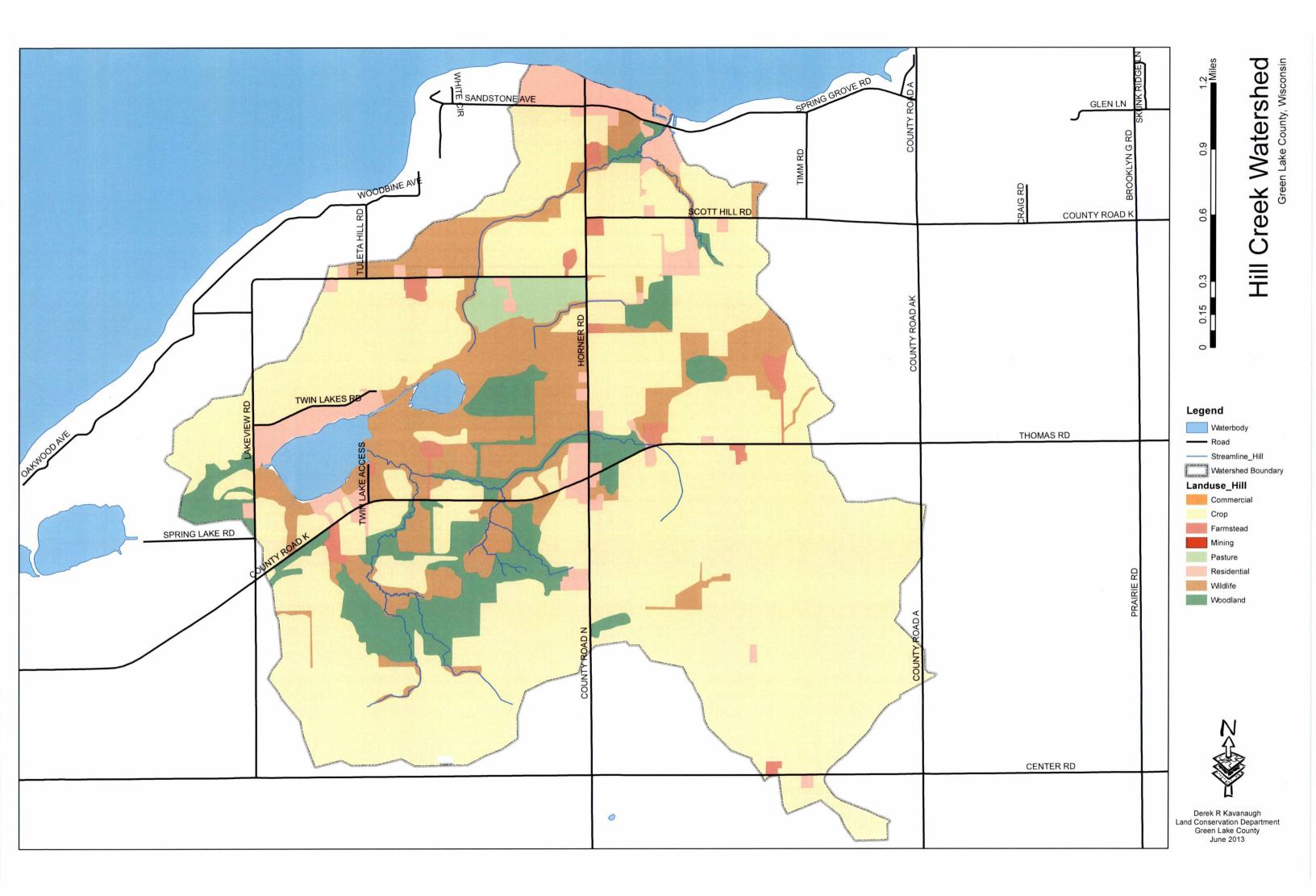
Commercial

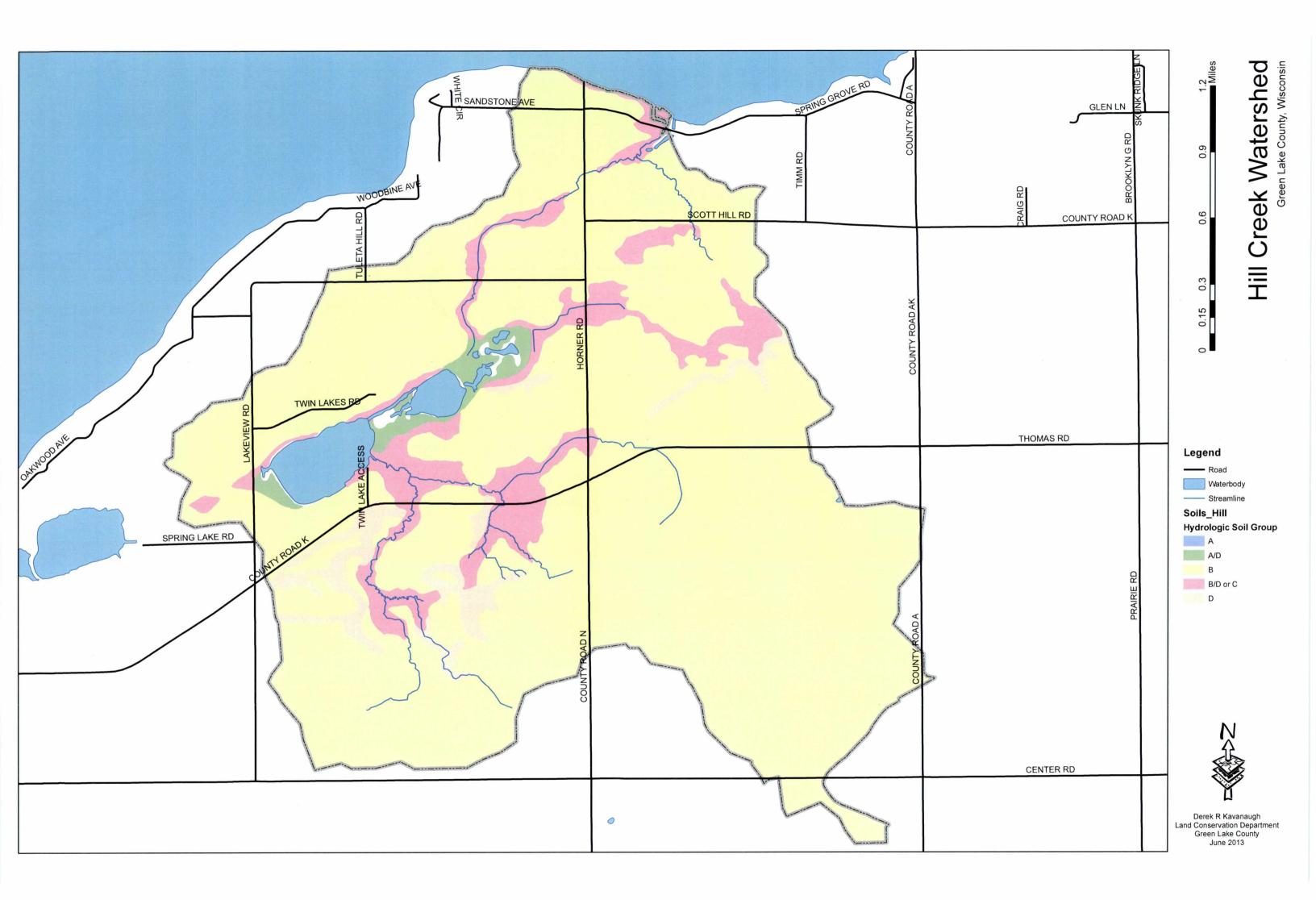
Farmstead

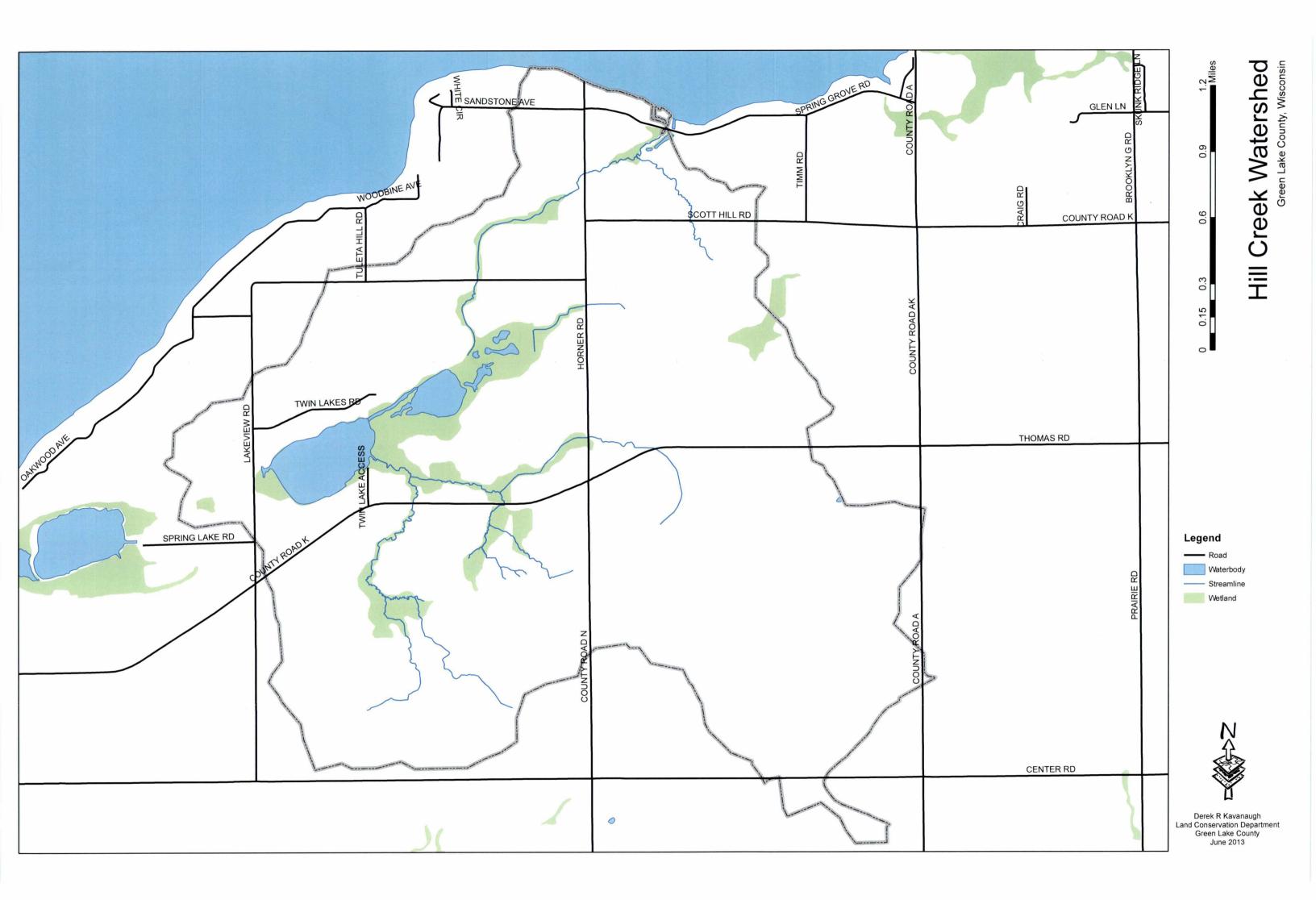
Residential

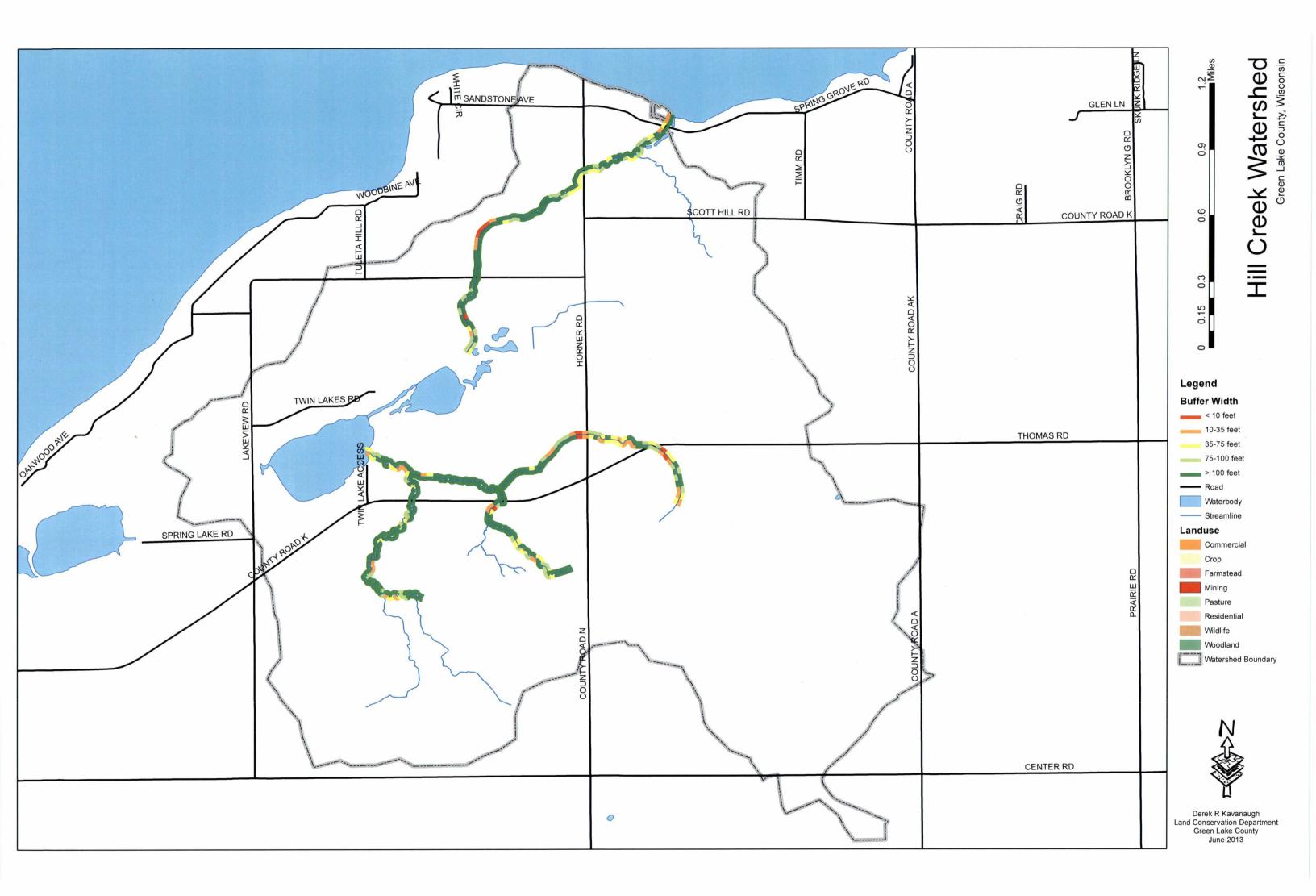


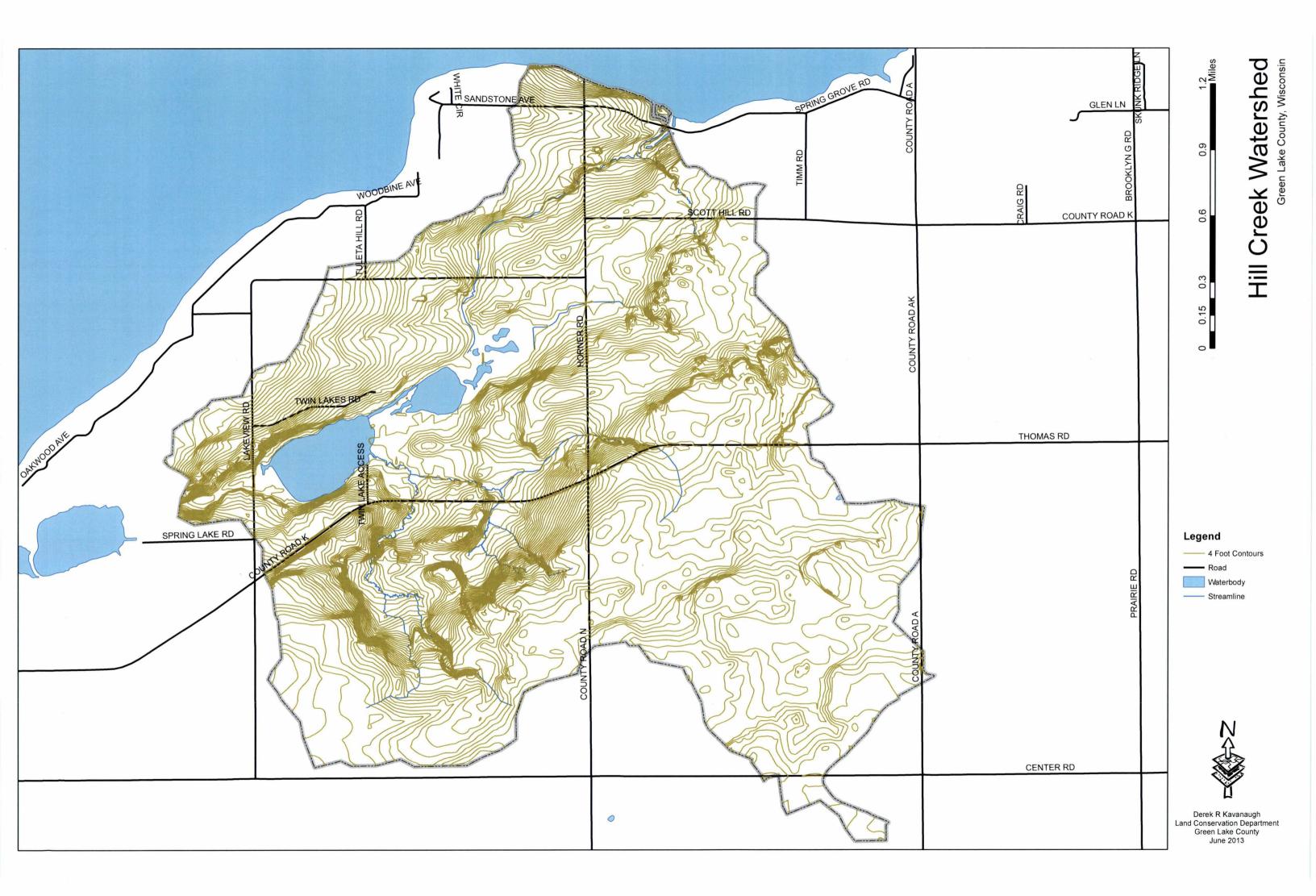
Derek R Kavanaugh Land Conservation Department Green Lake County June 2013

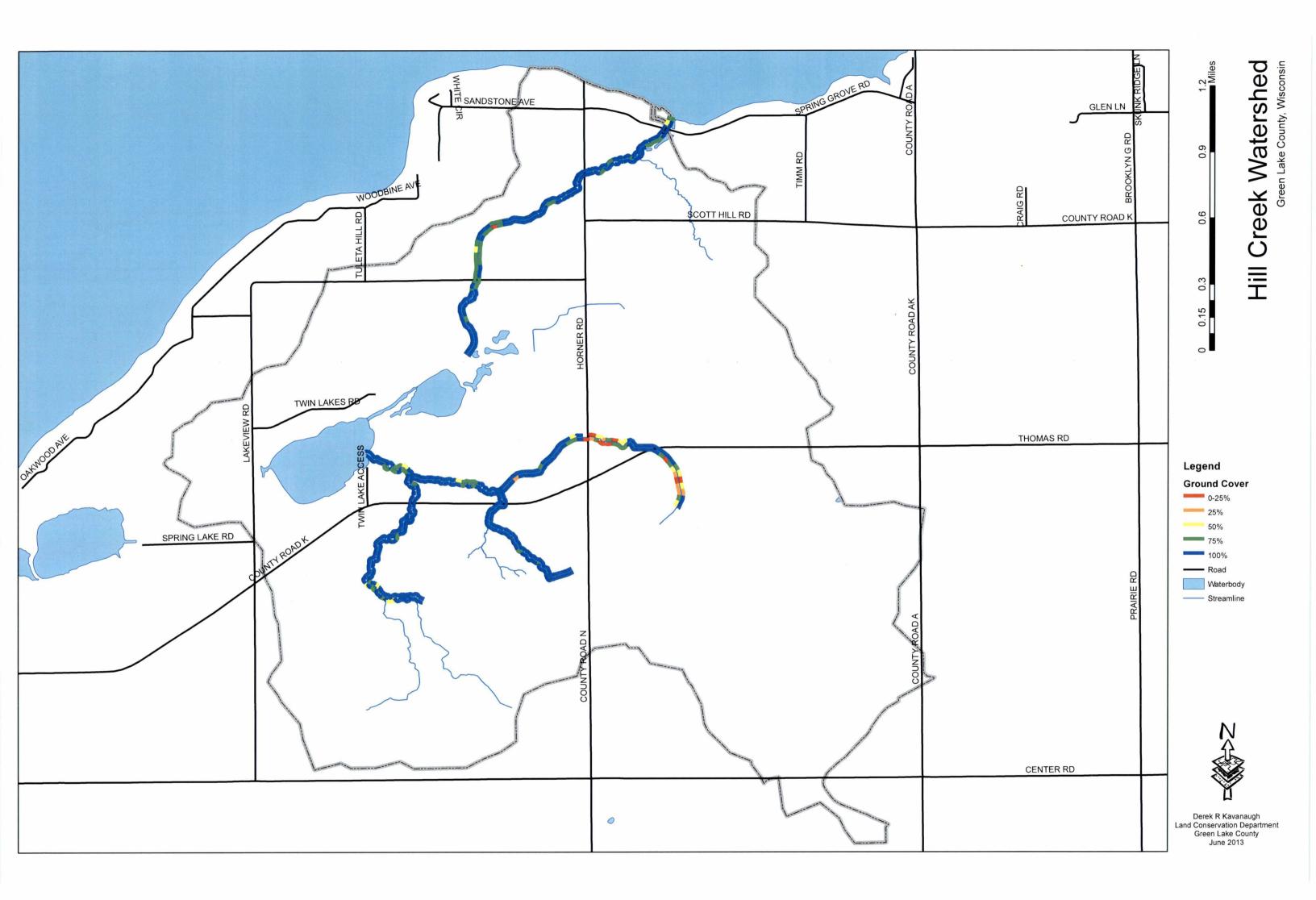


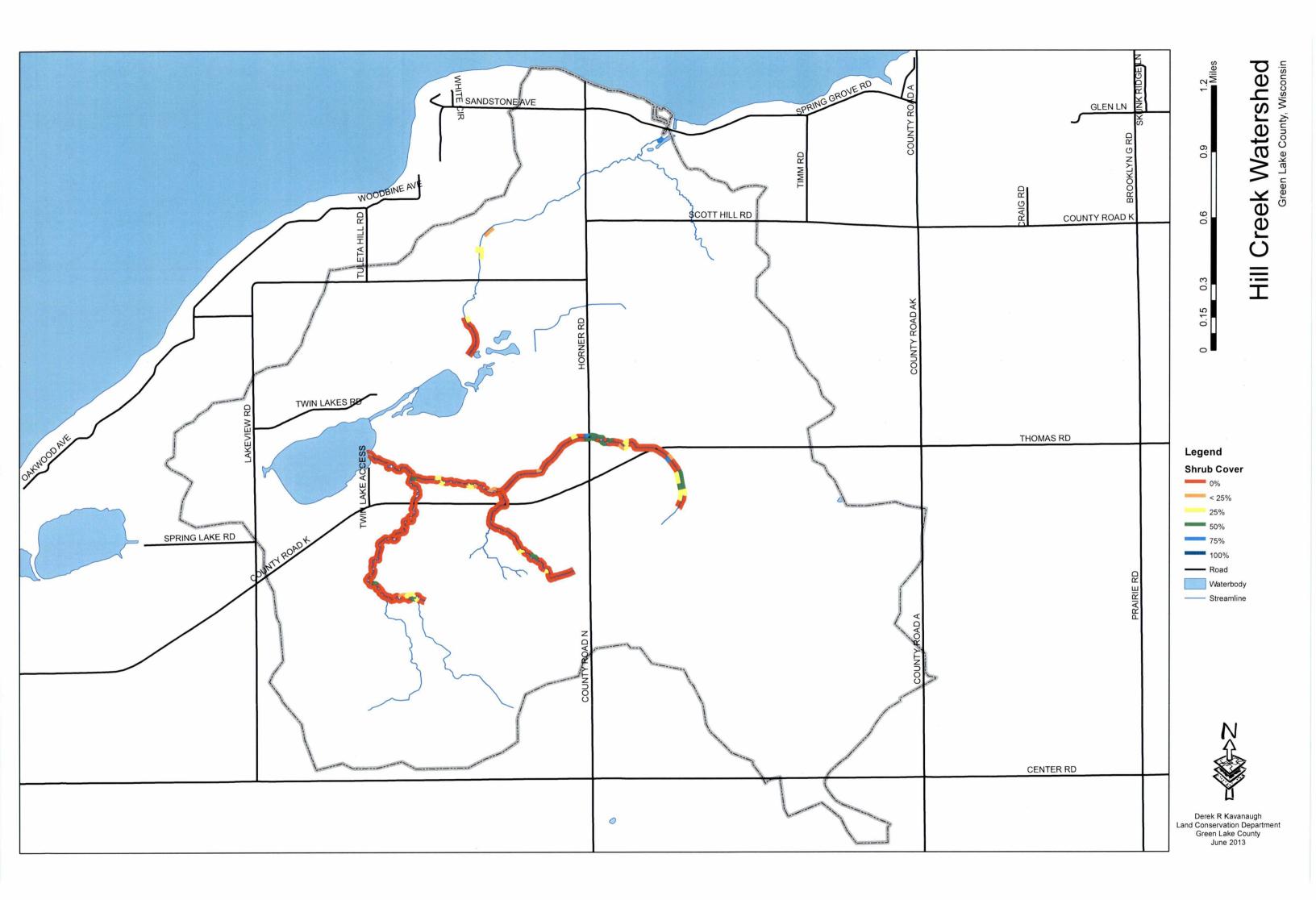


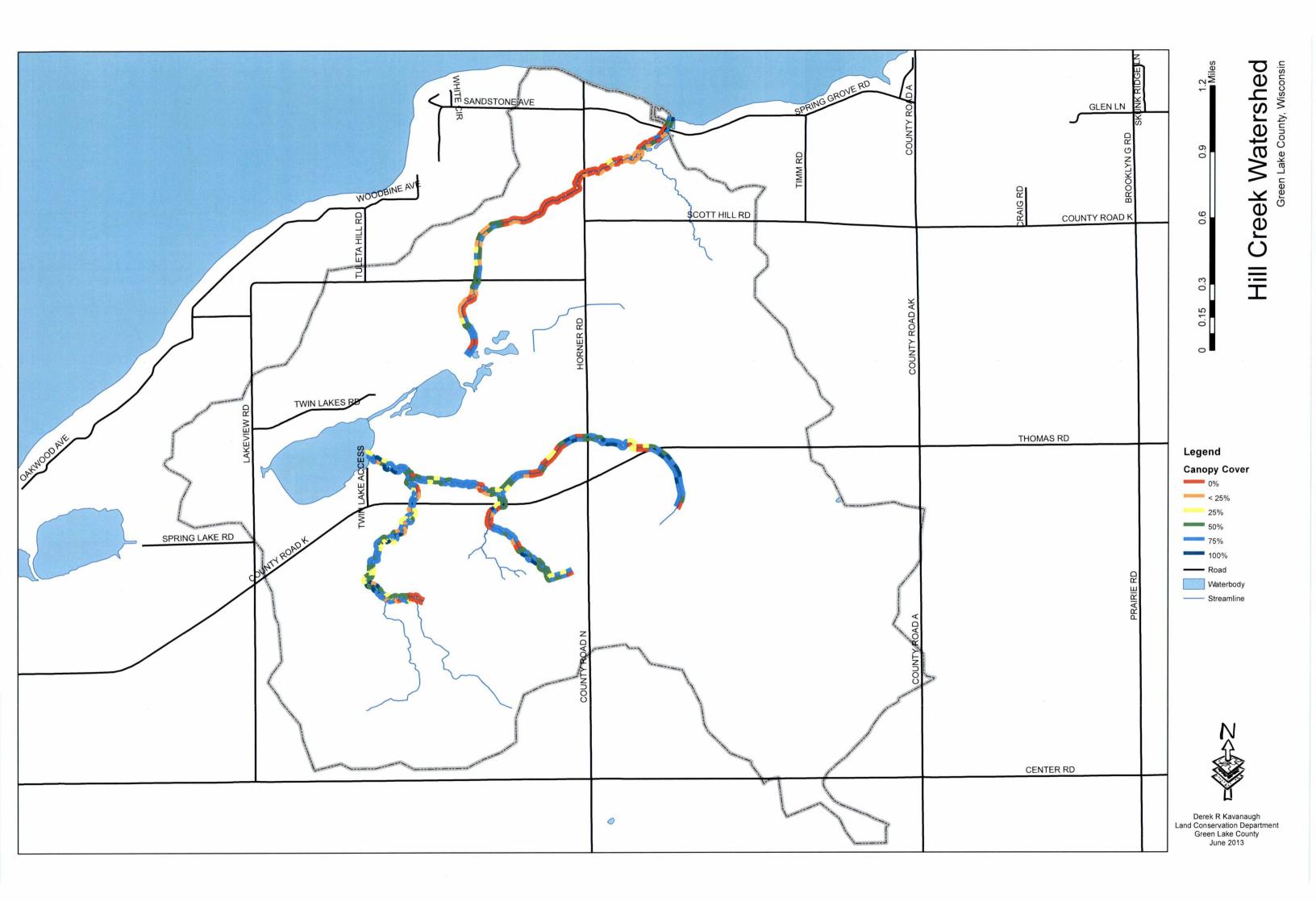


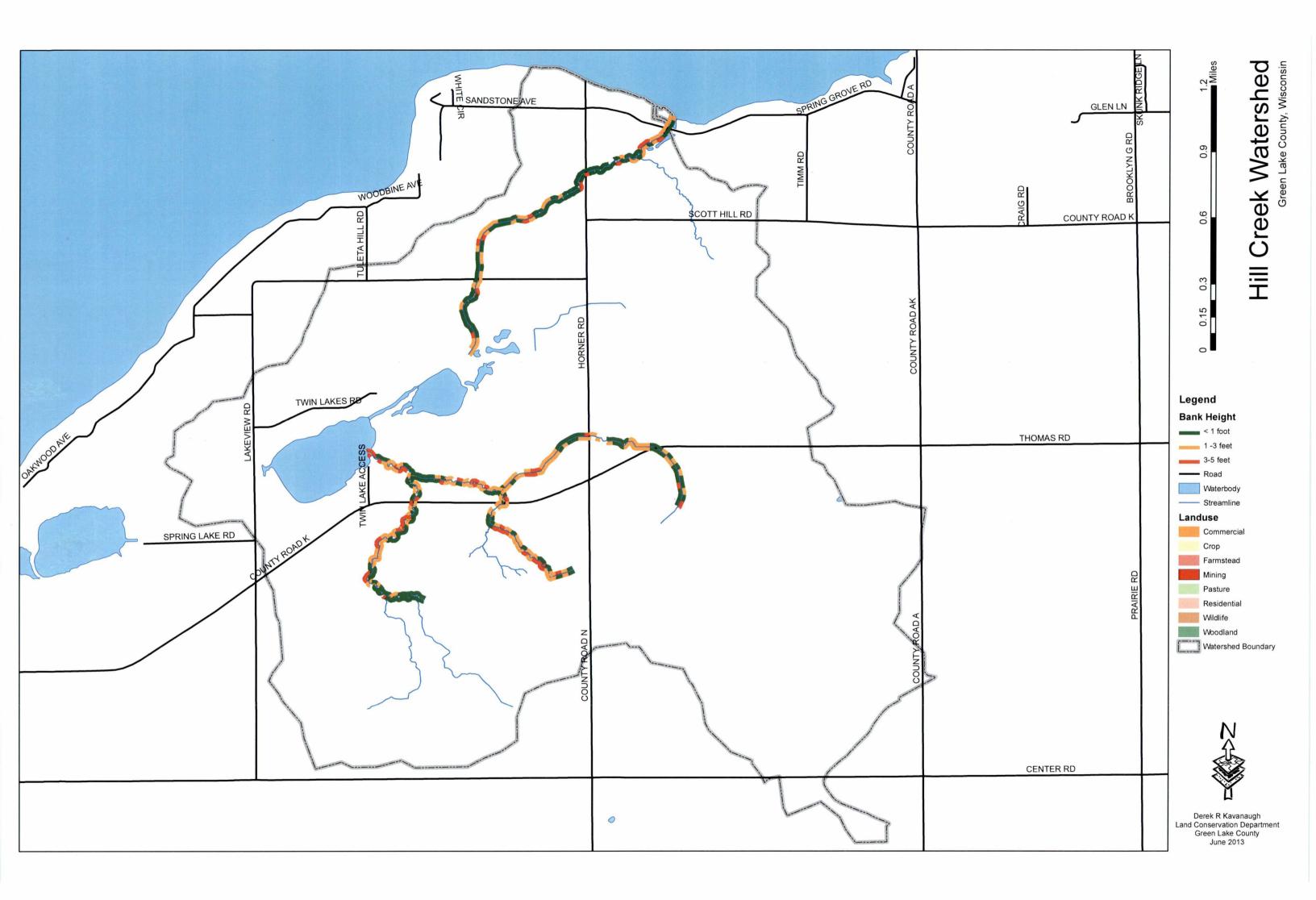


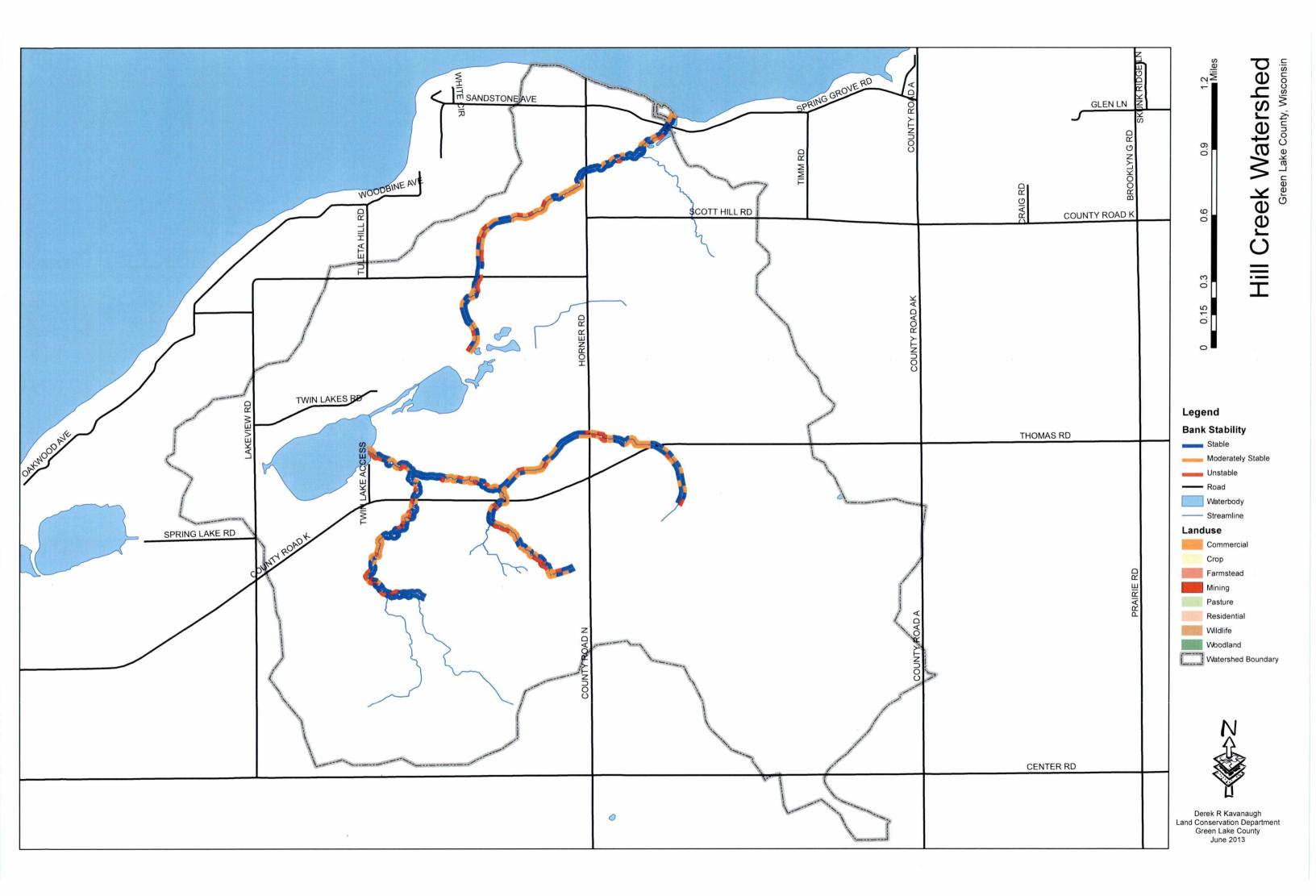


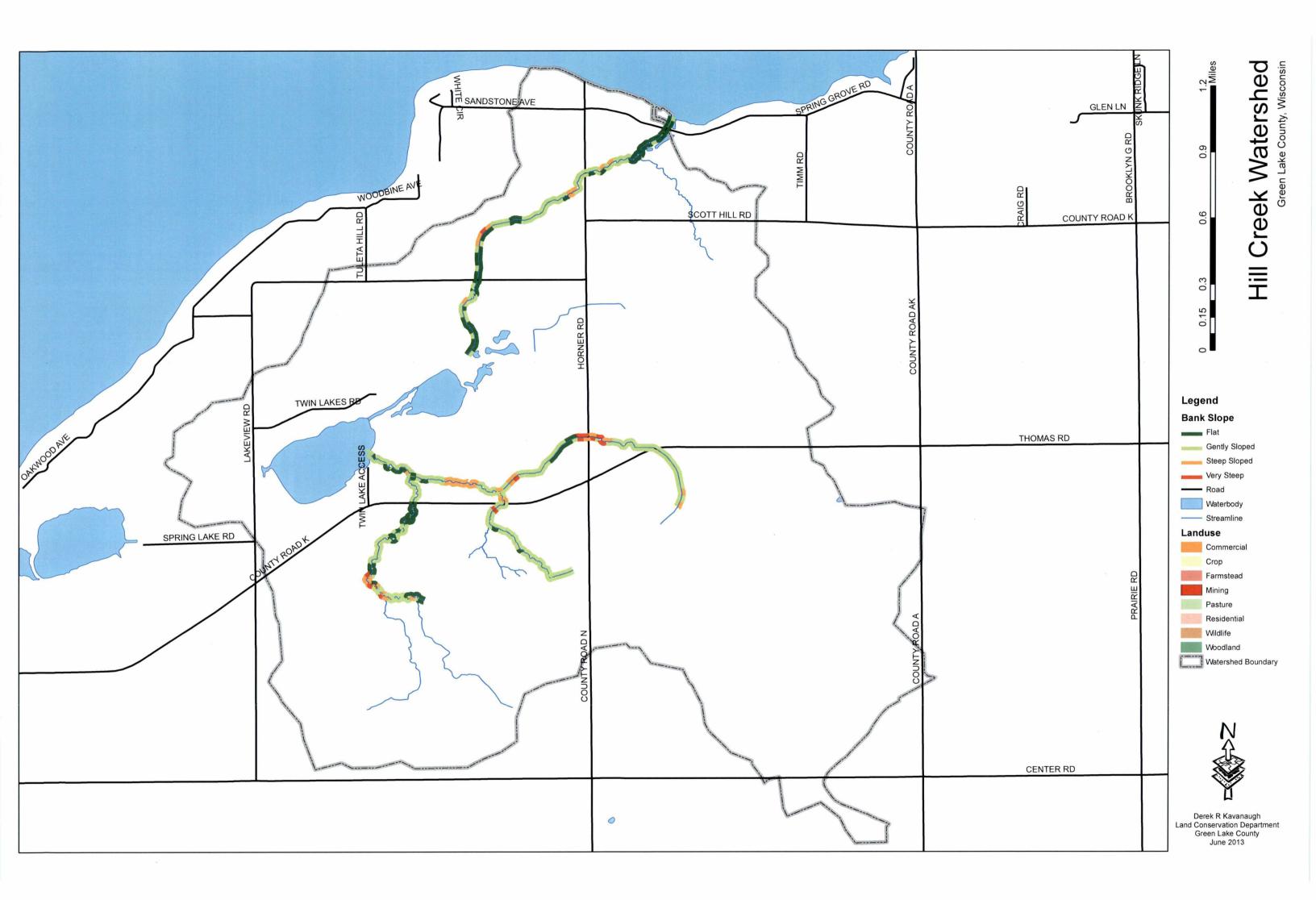


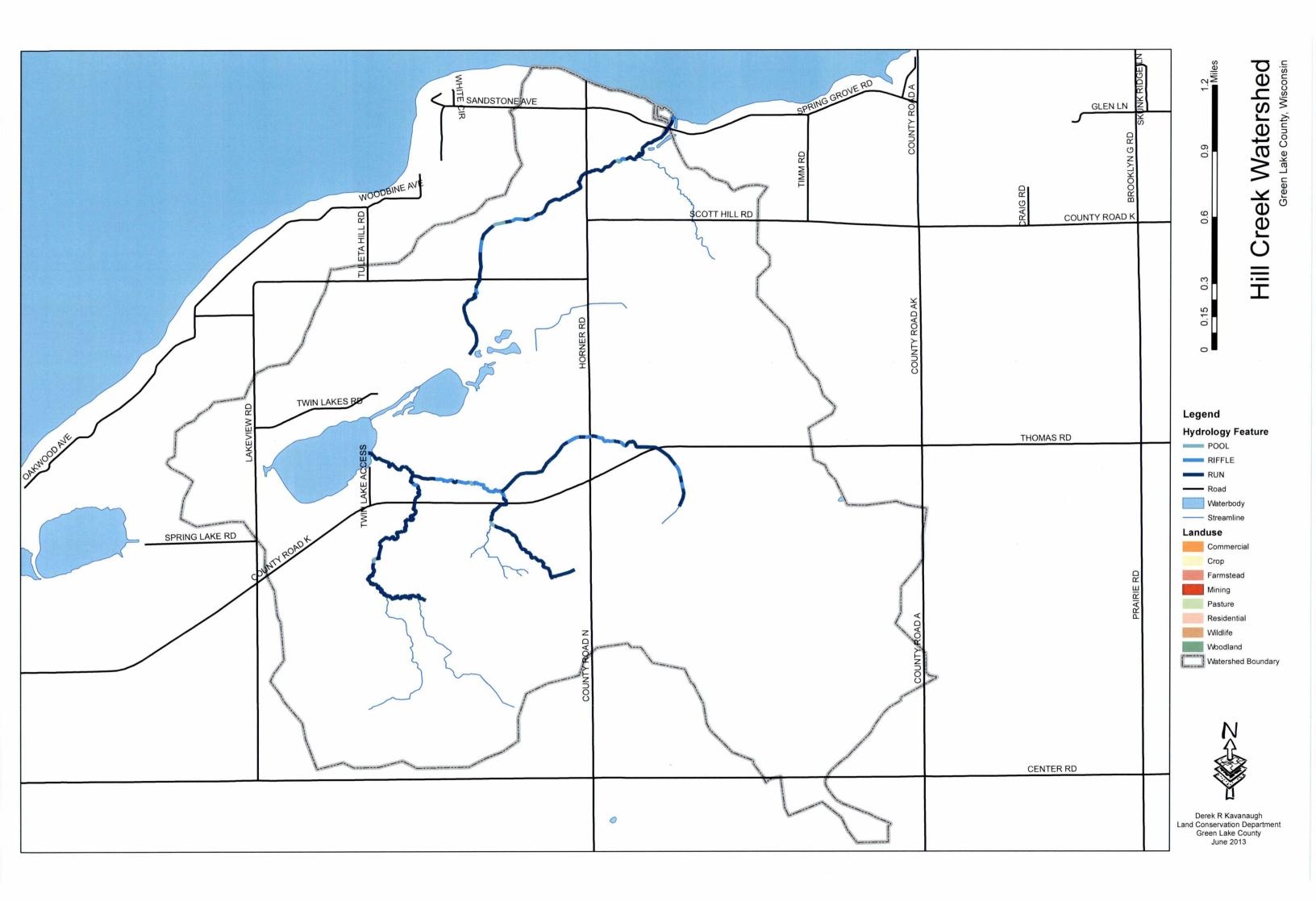


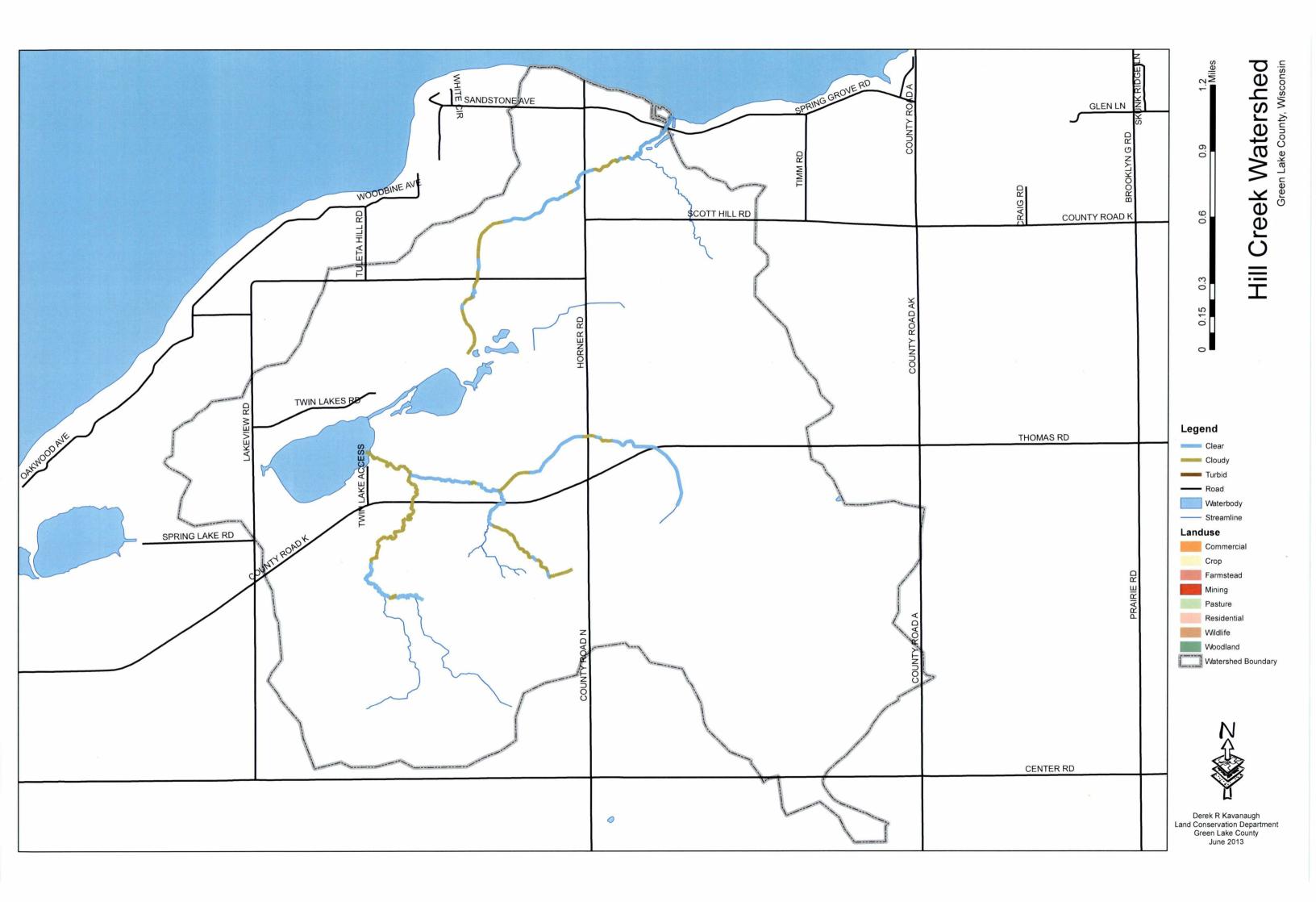


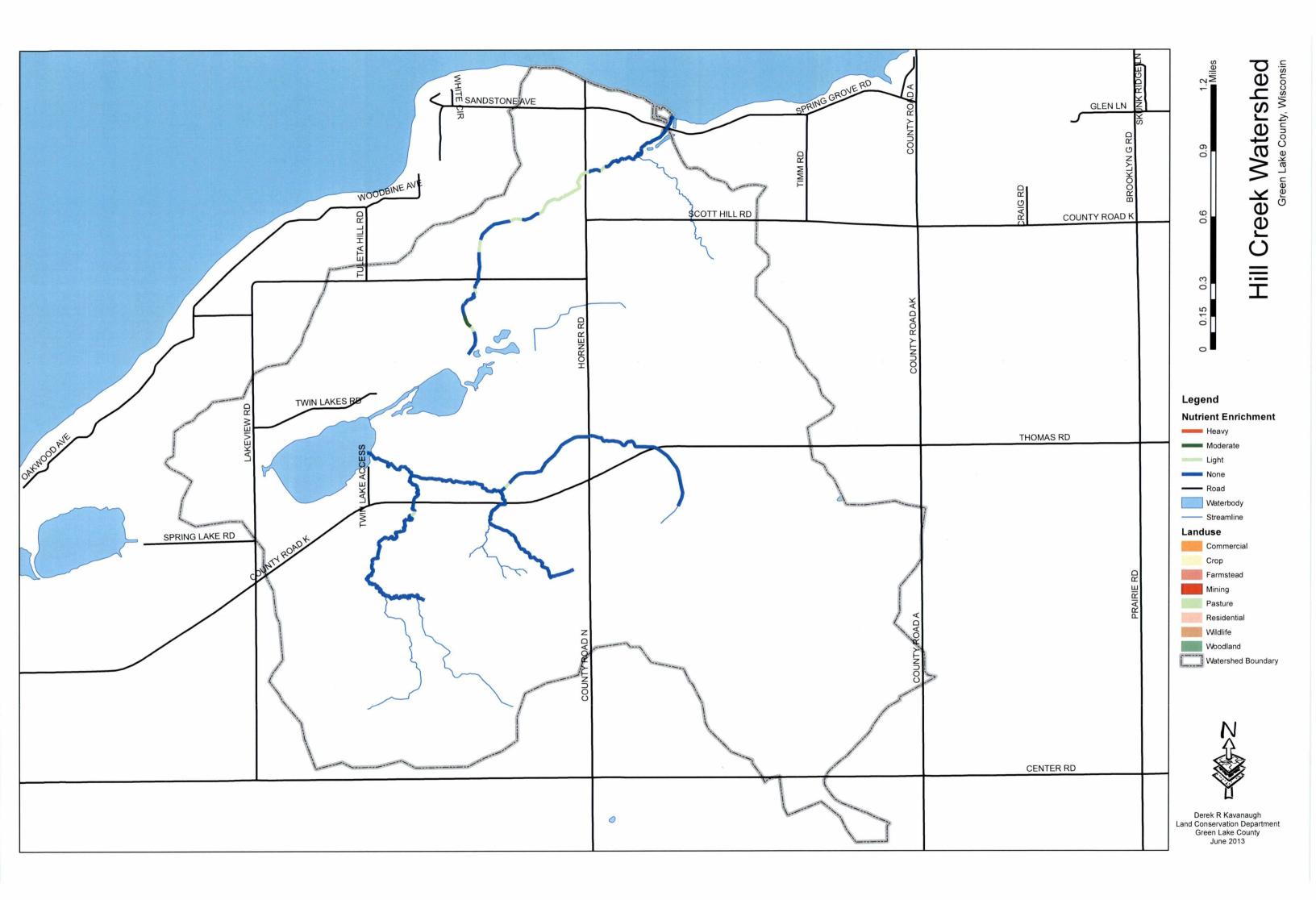


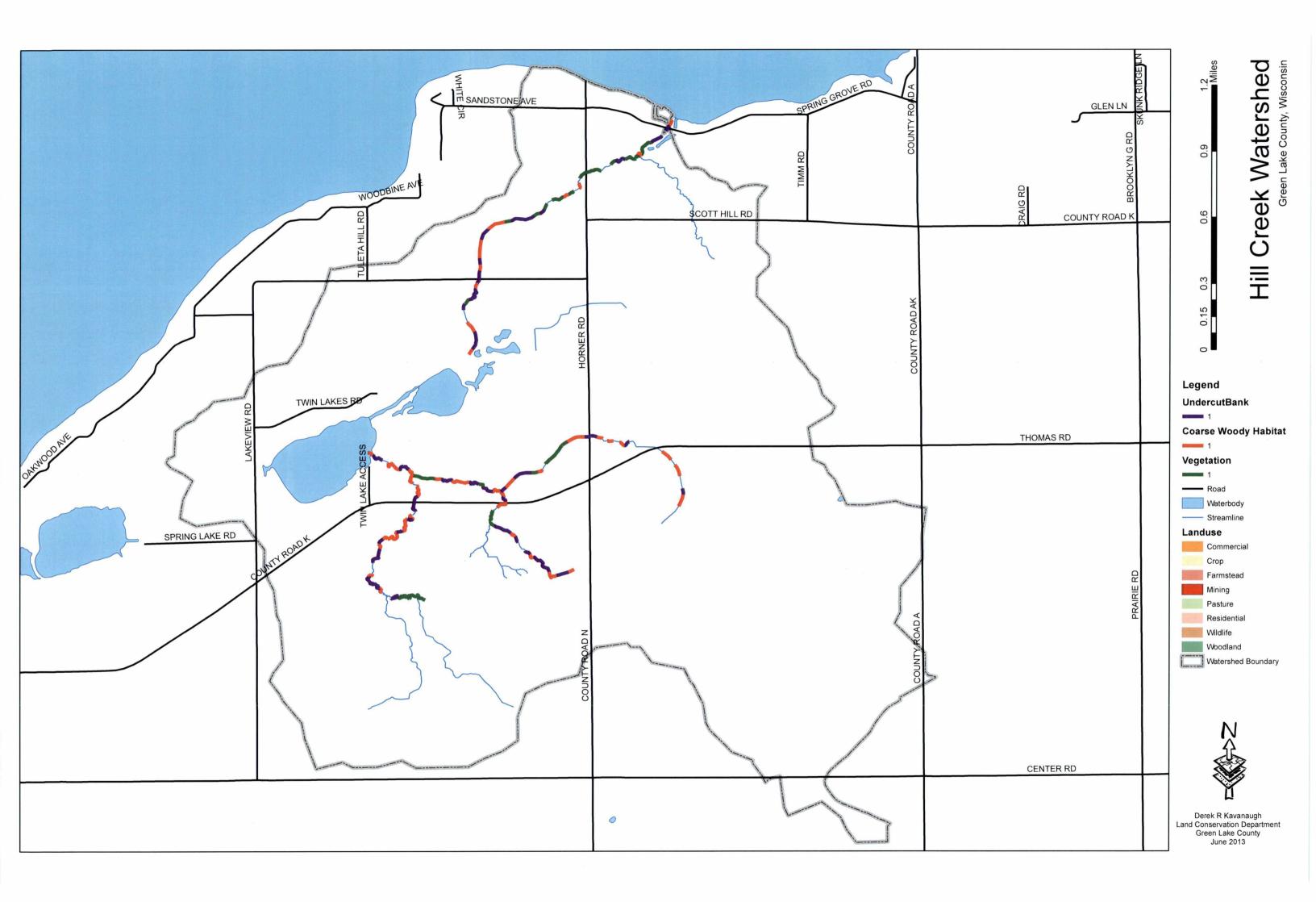


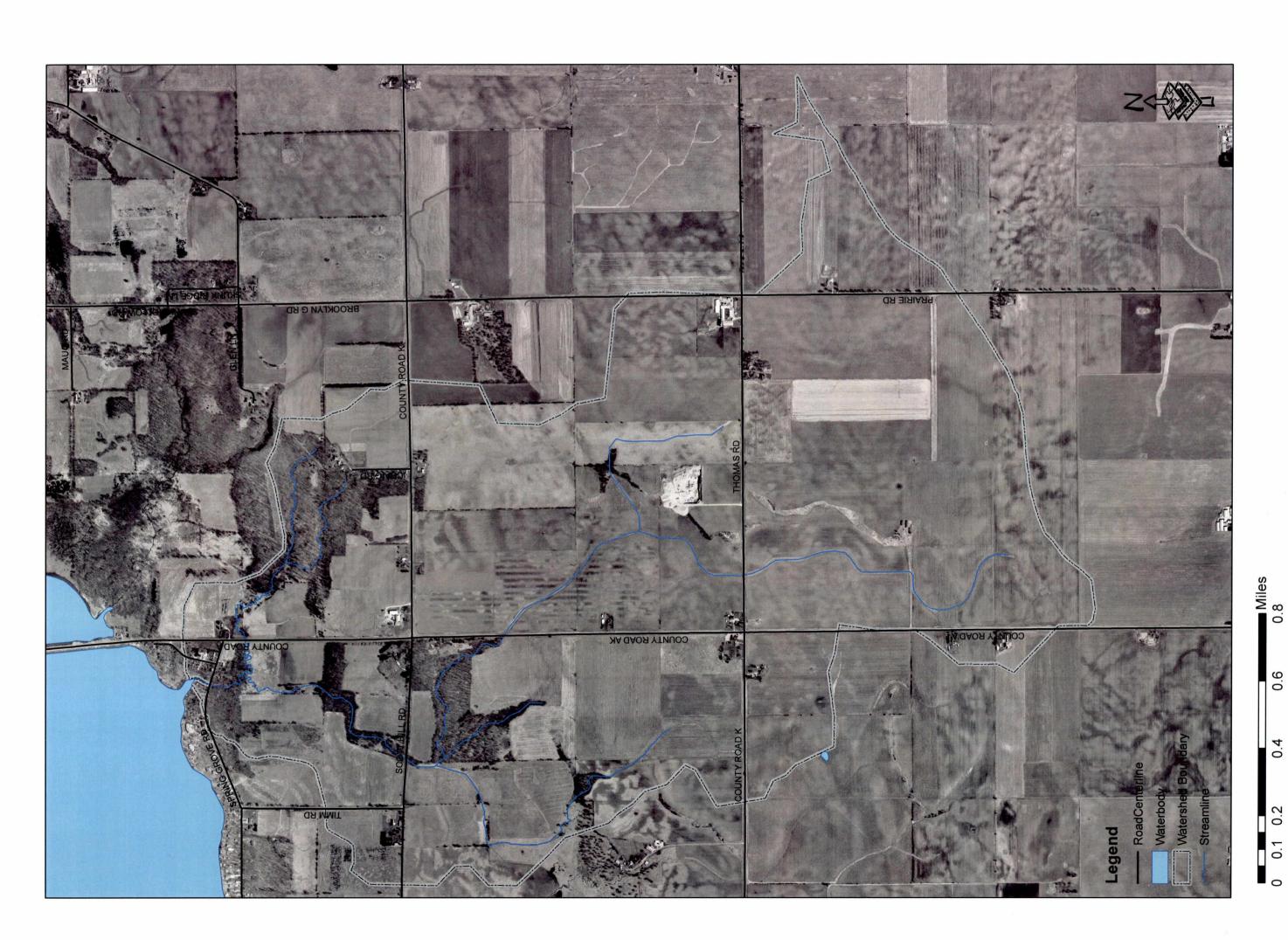


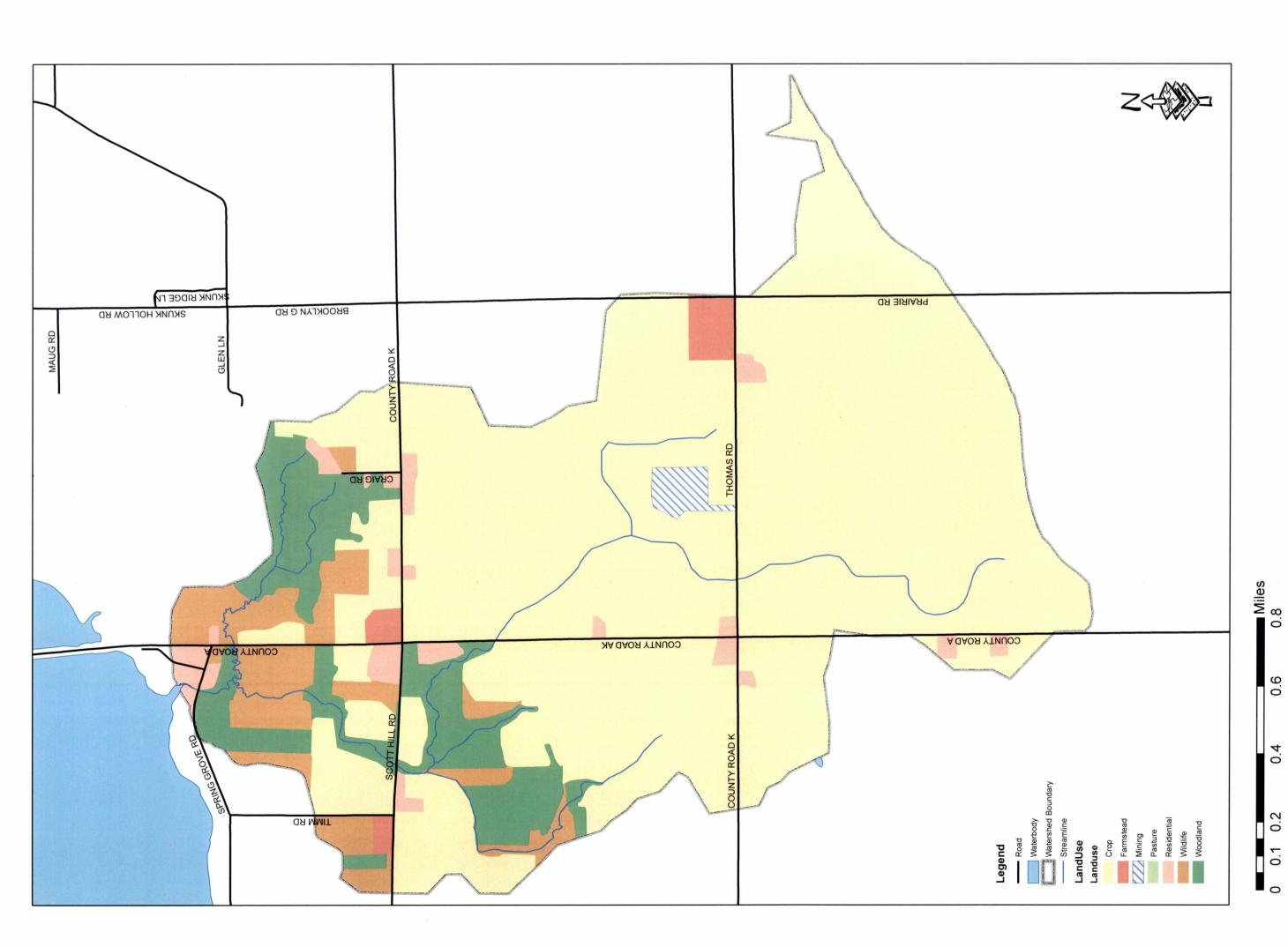


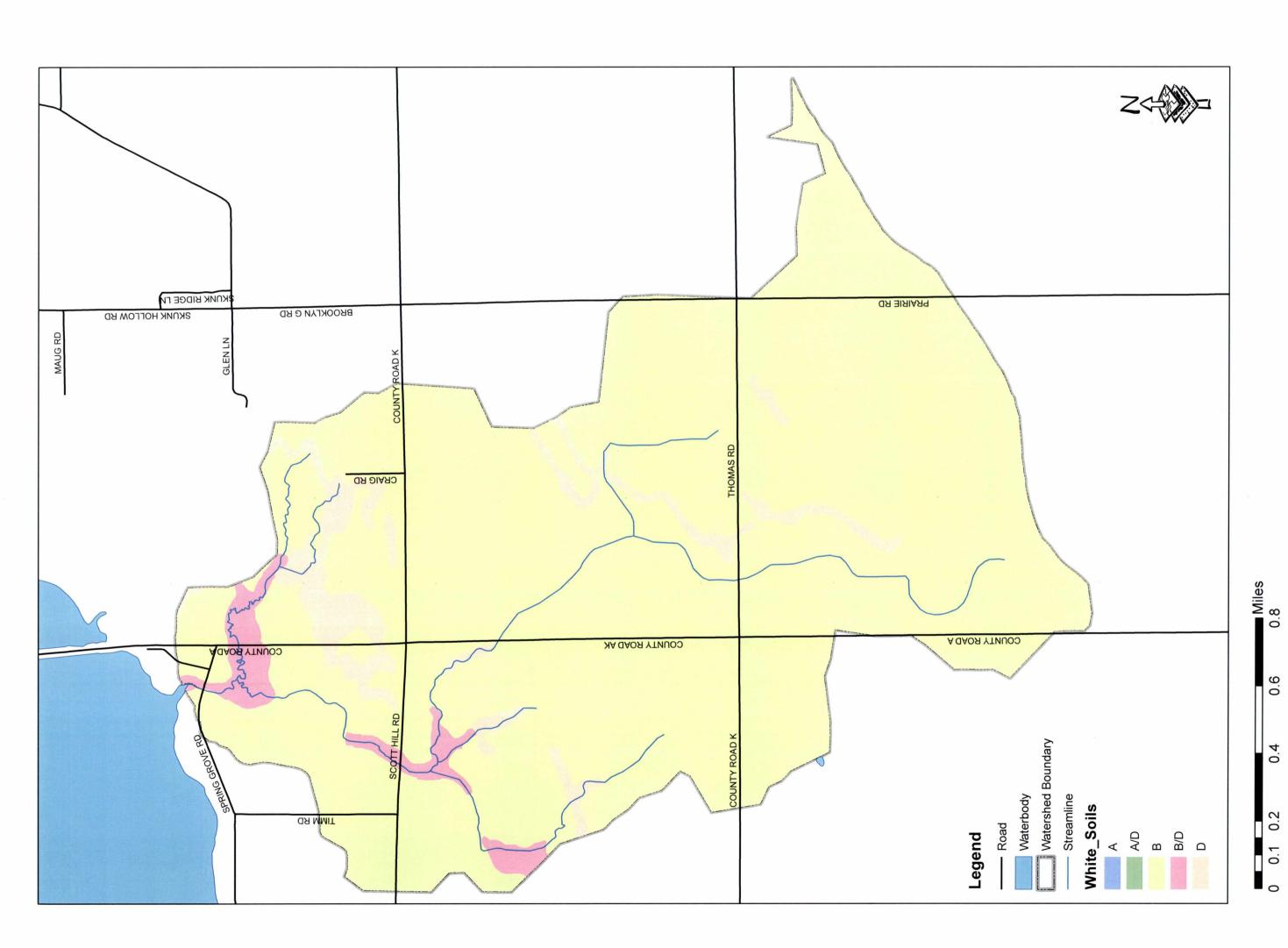


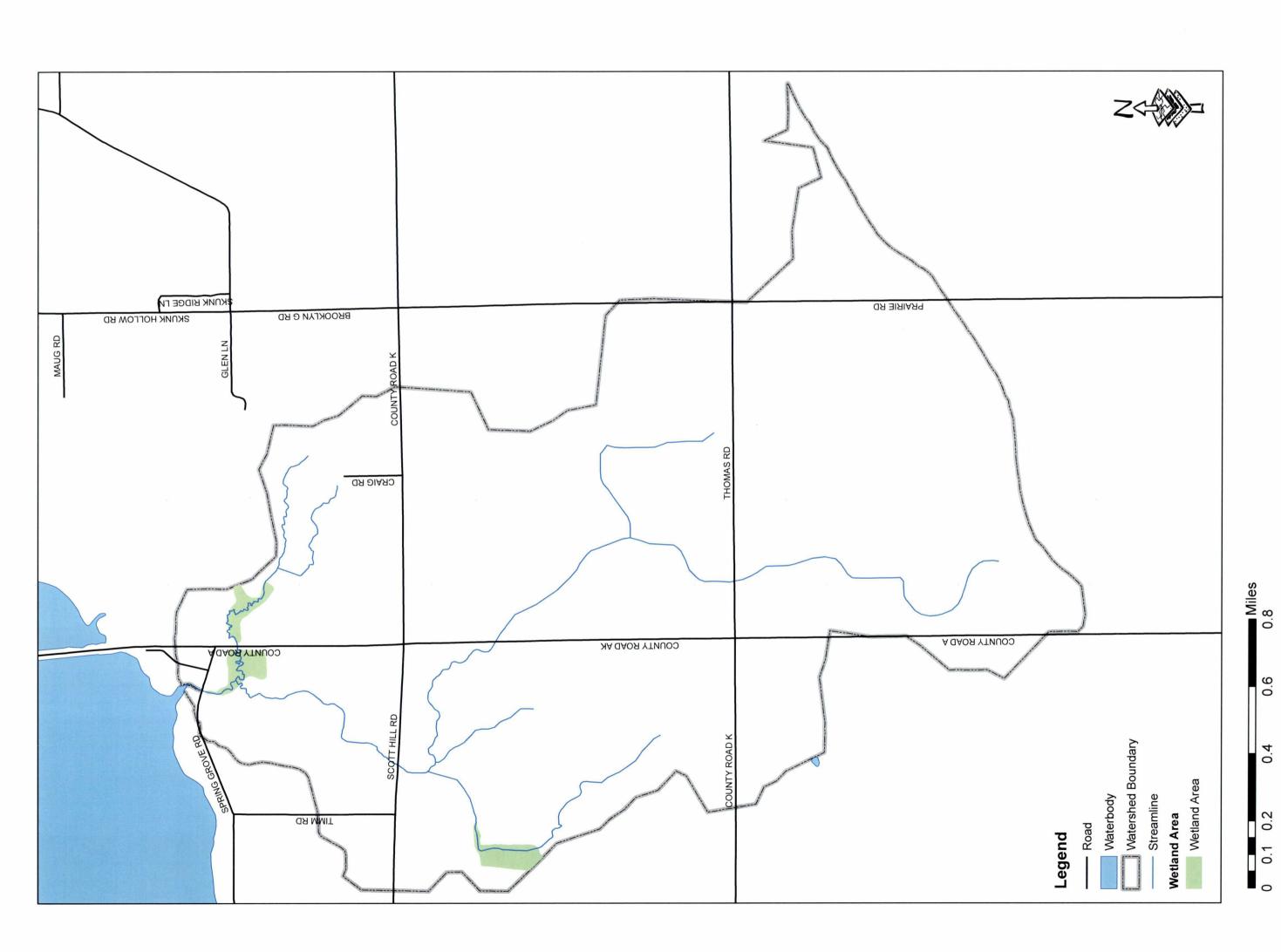


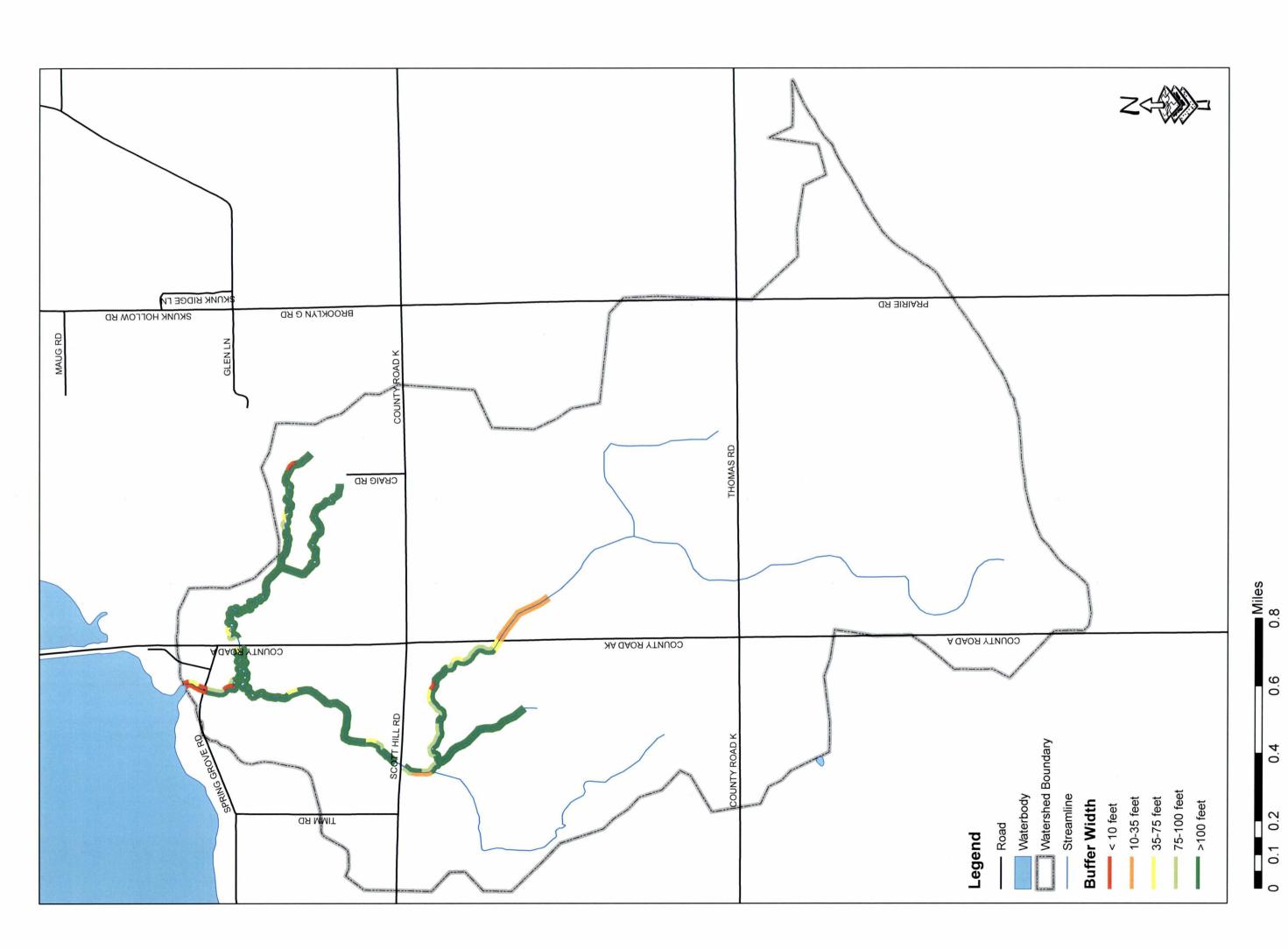


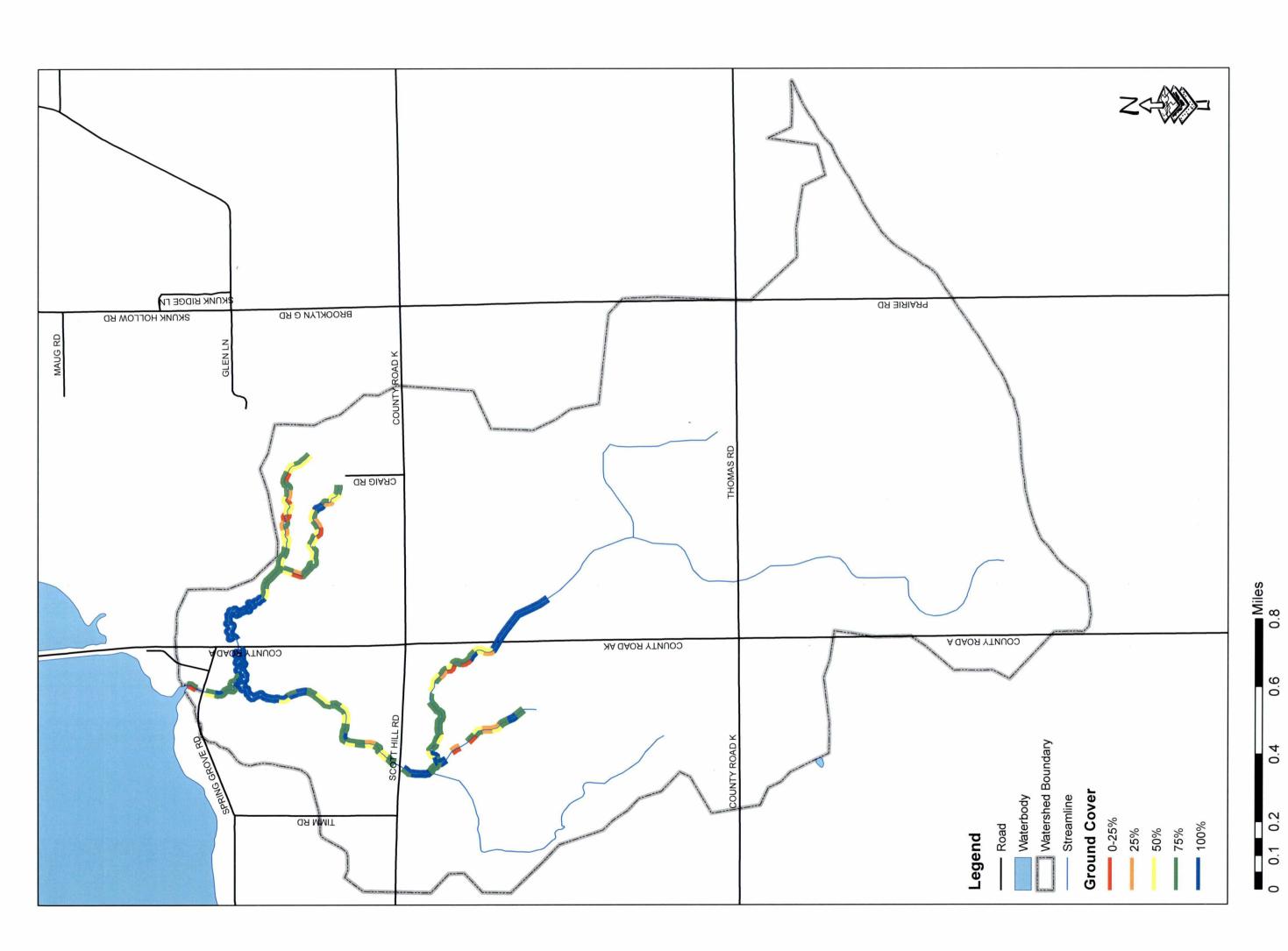


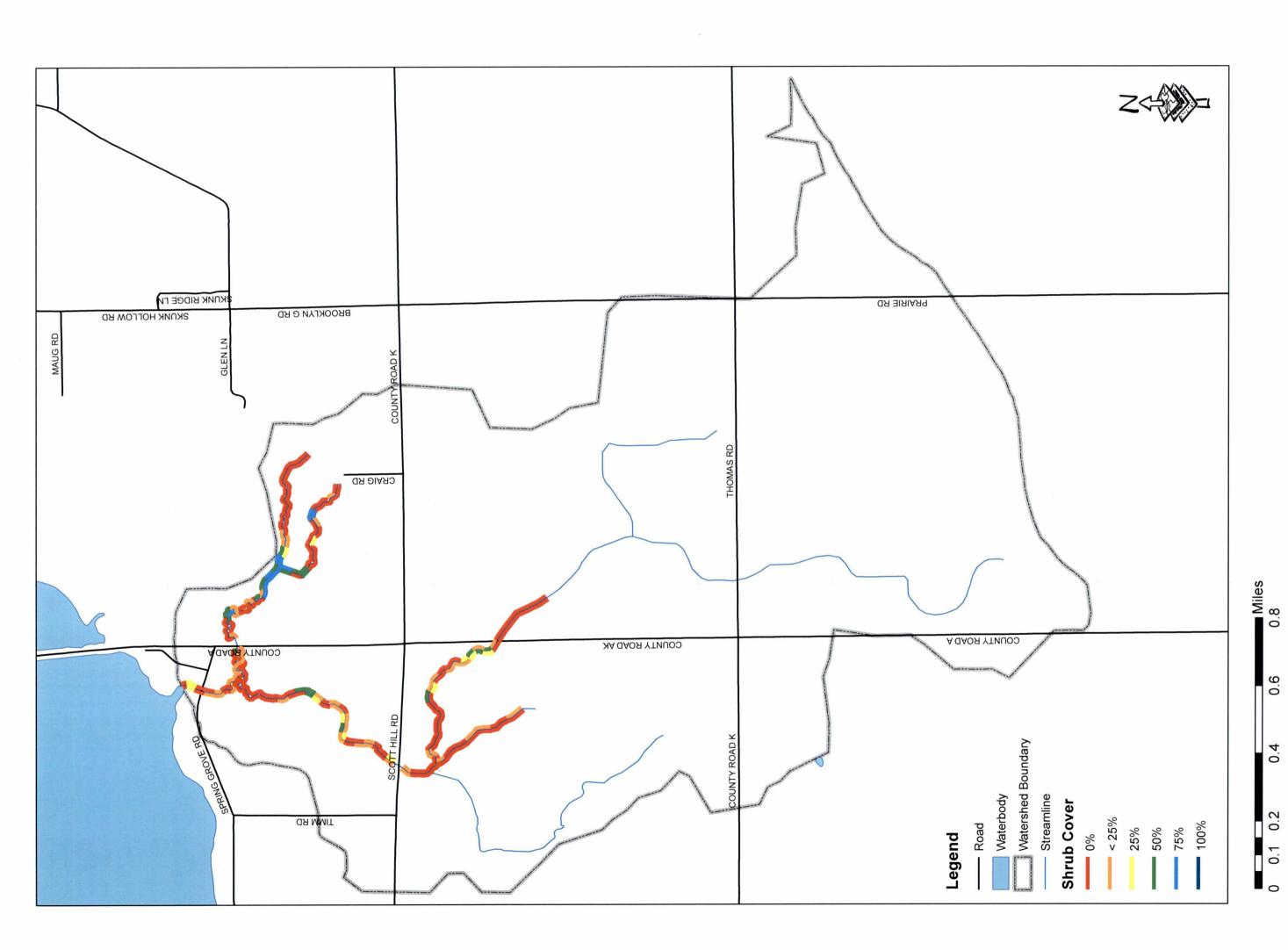


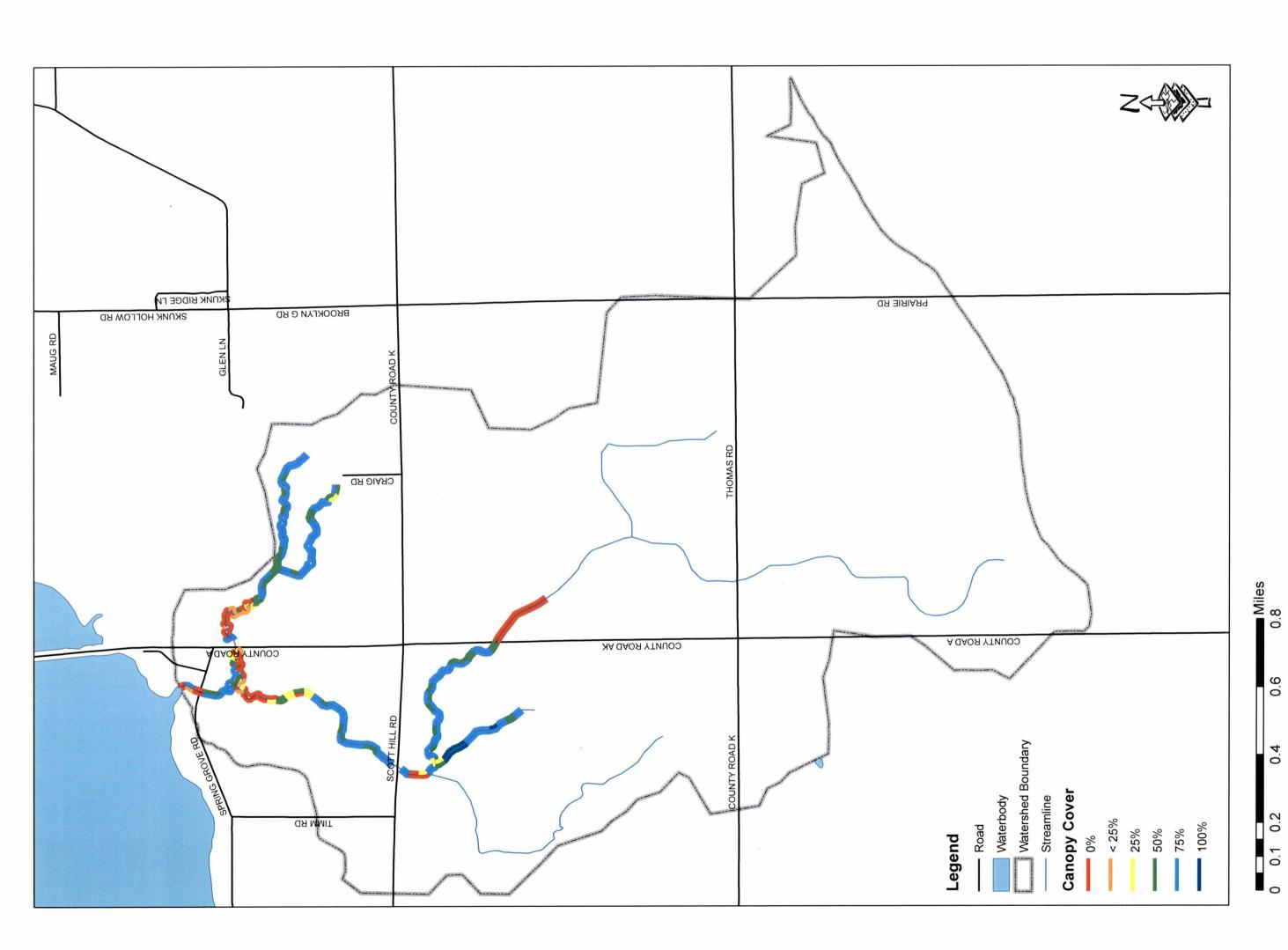


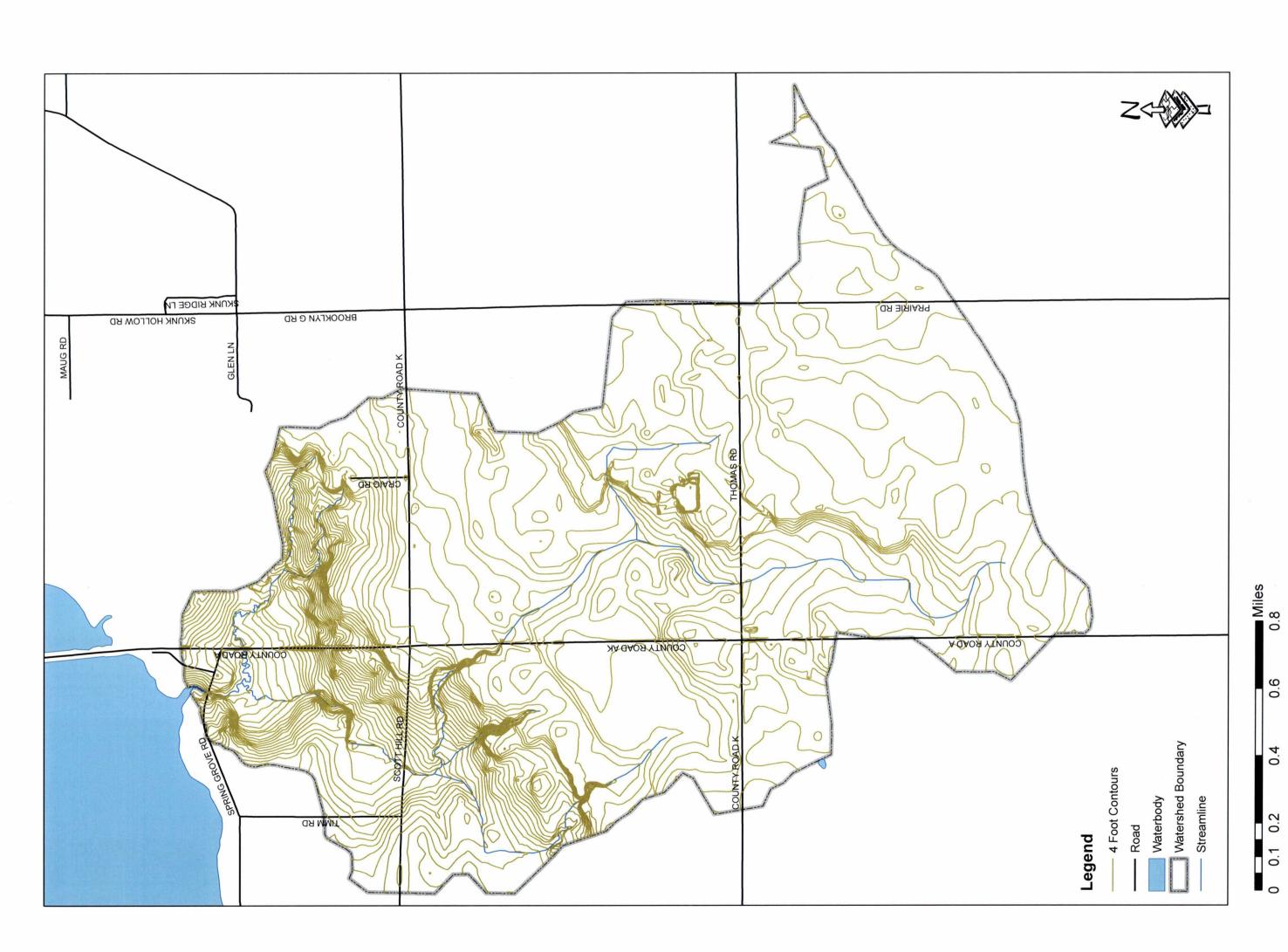


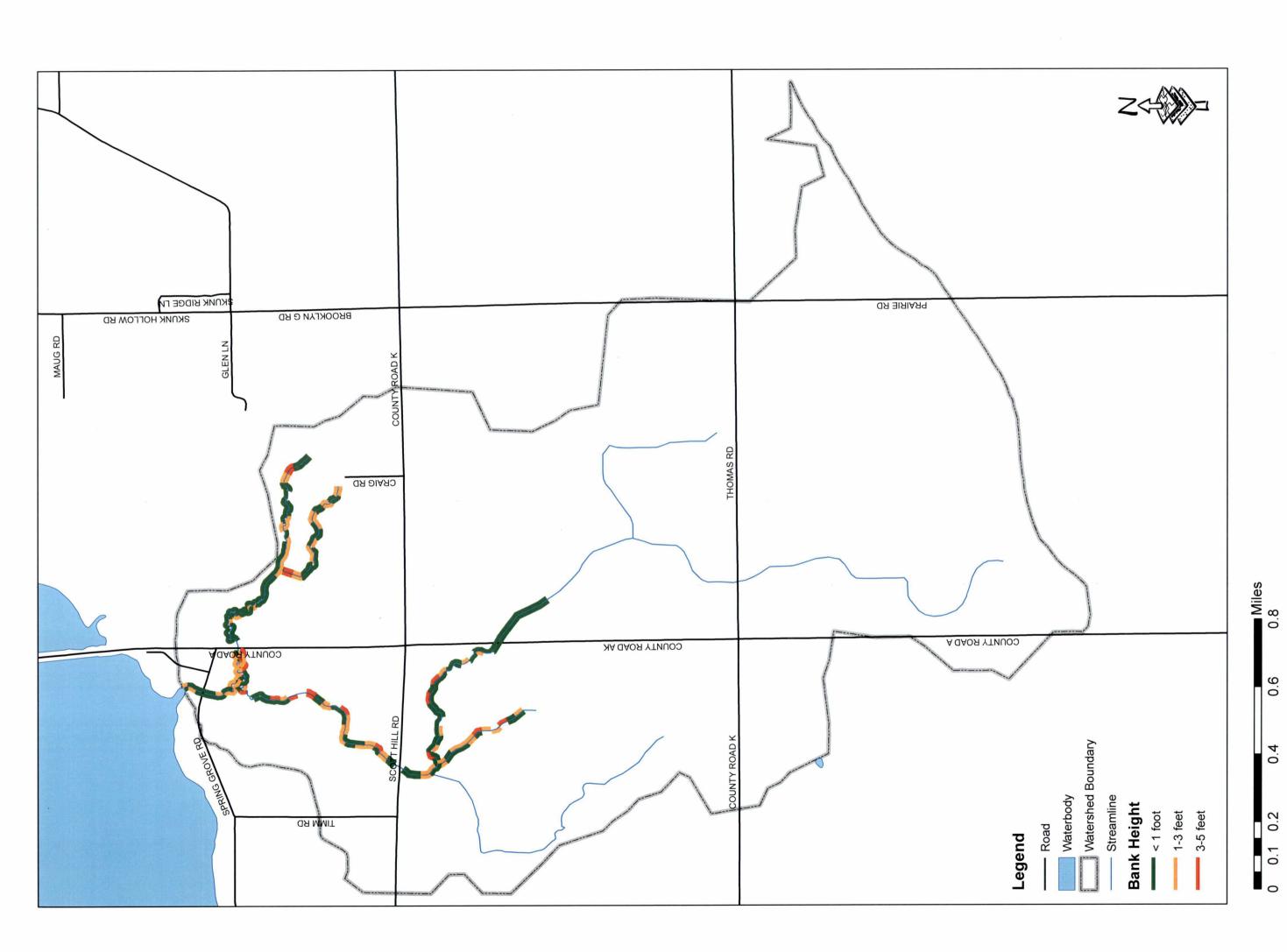


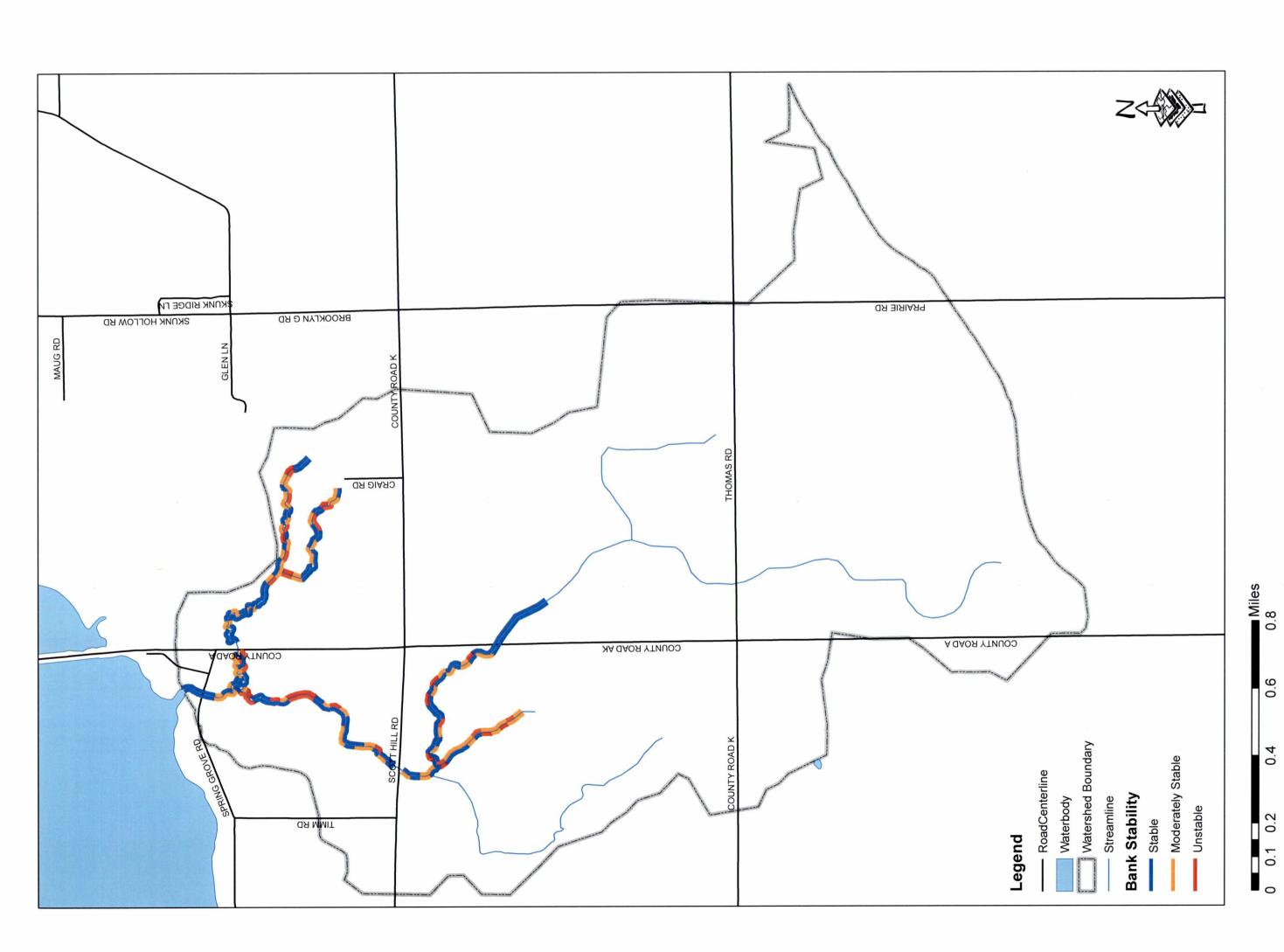


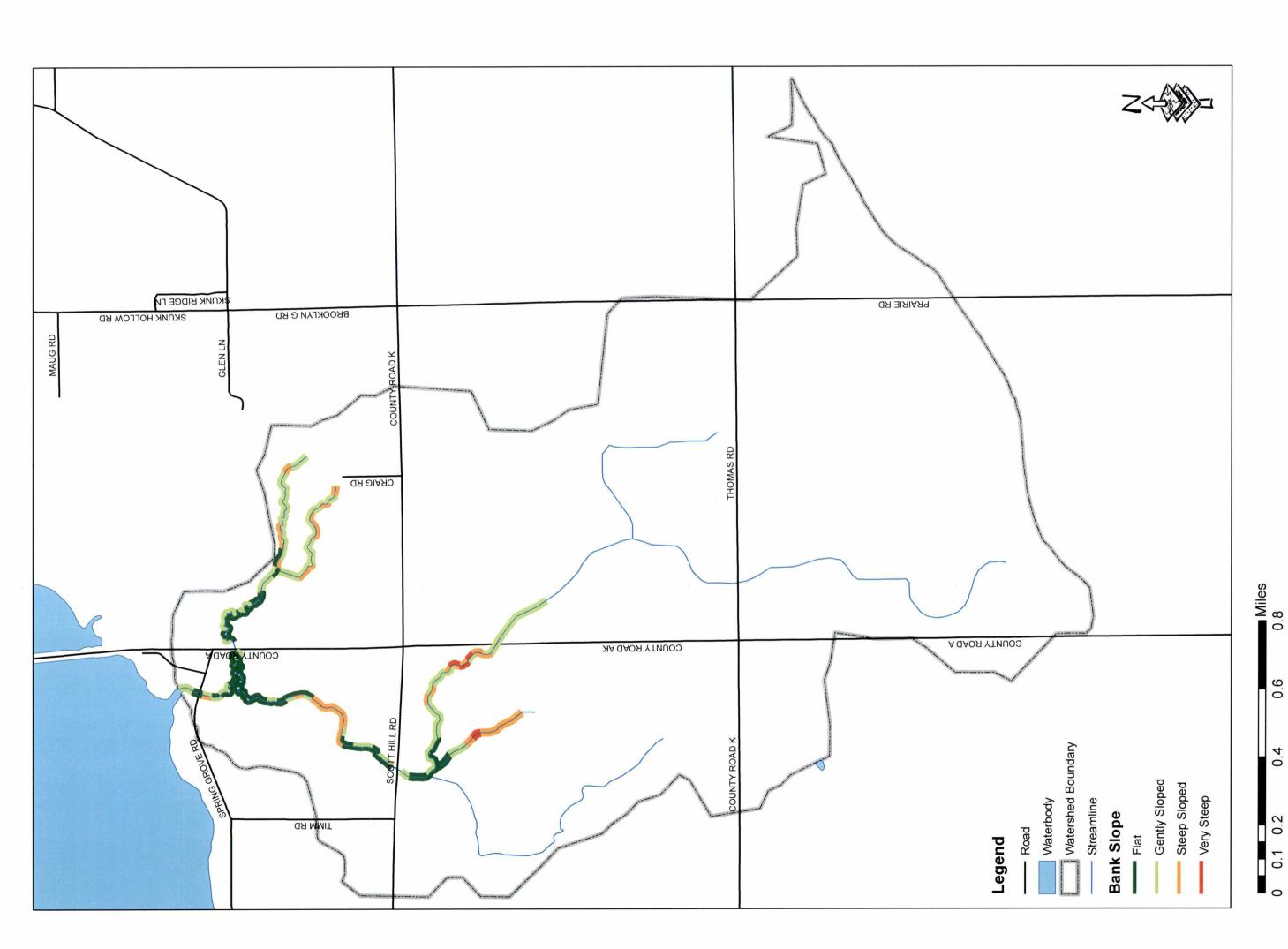


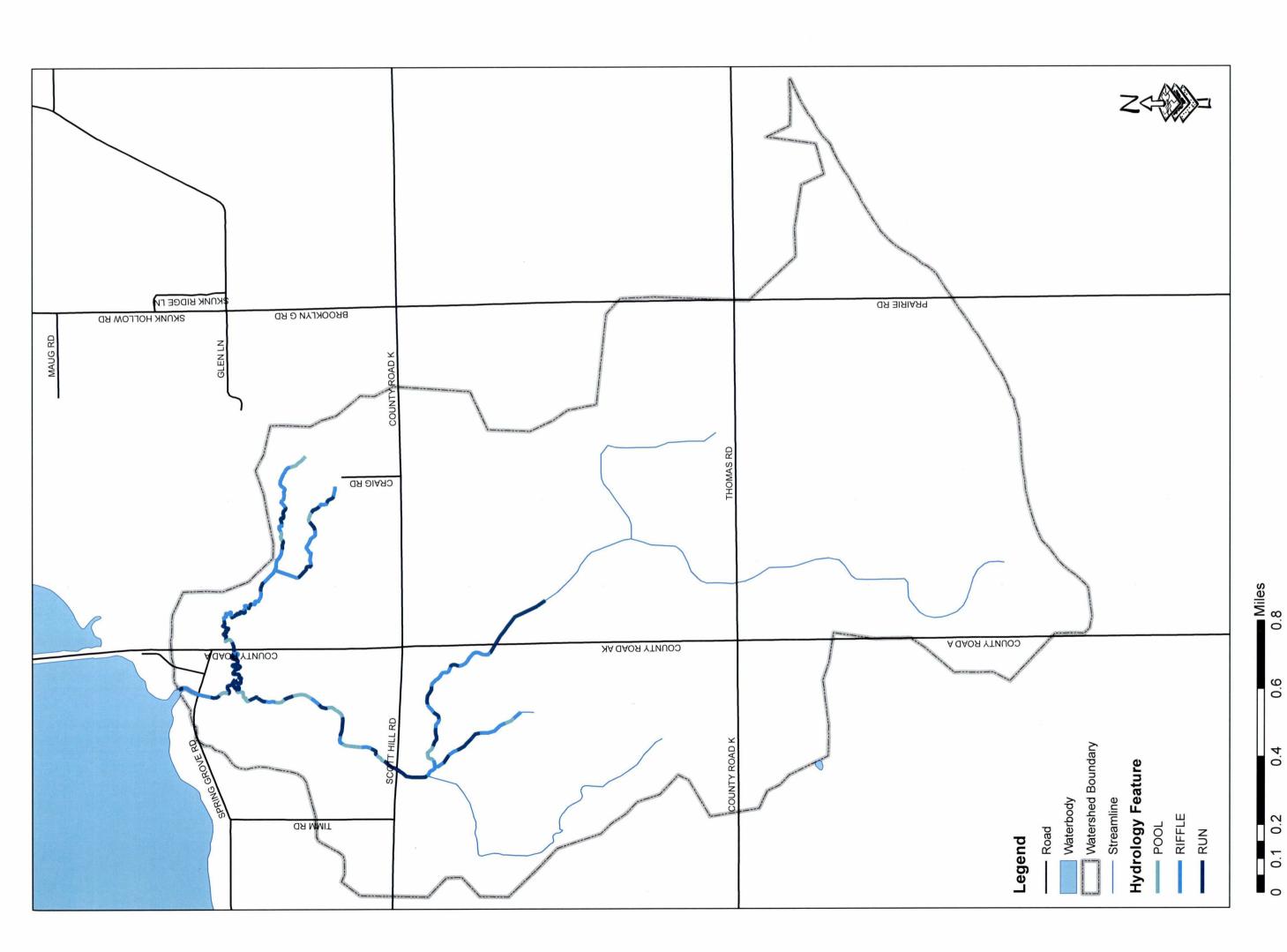


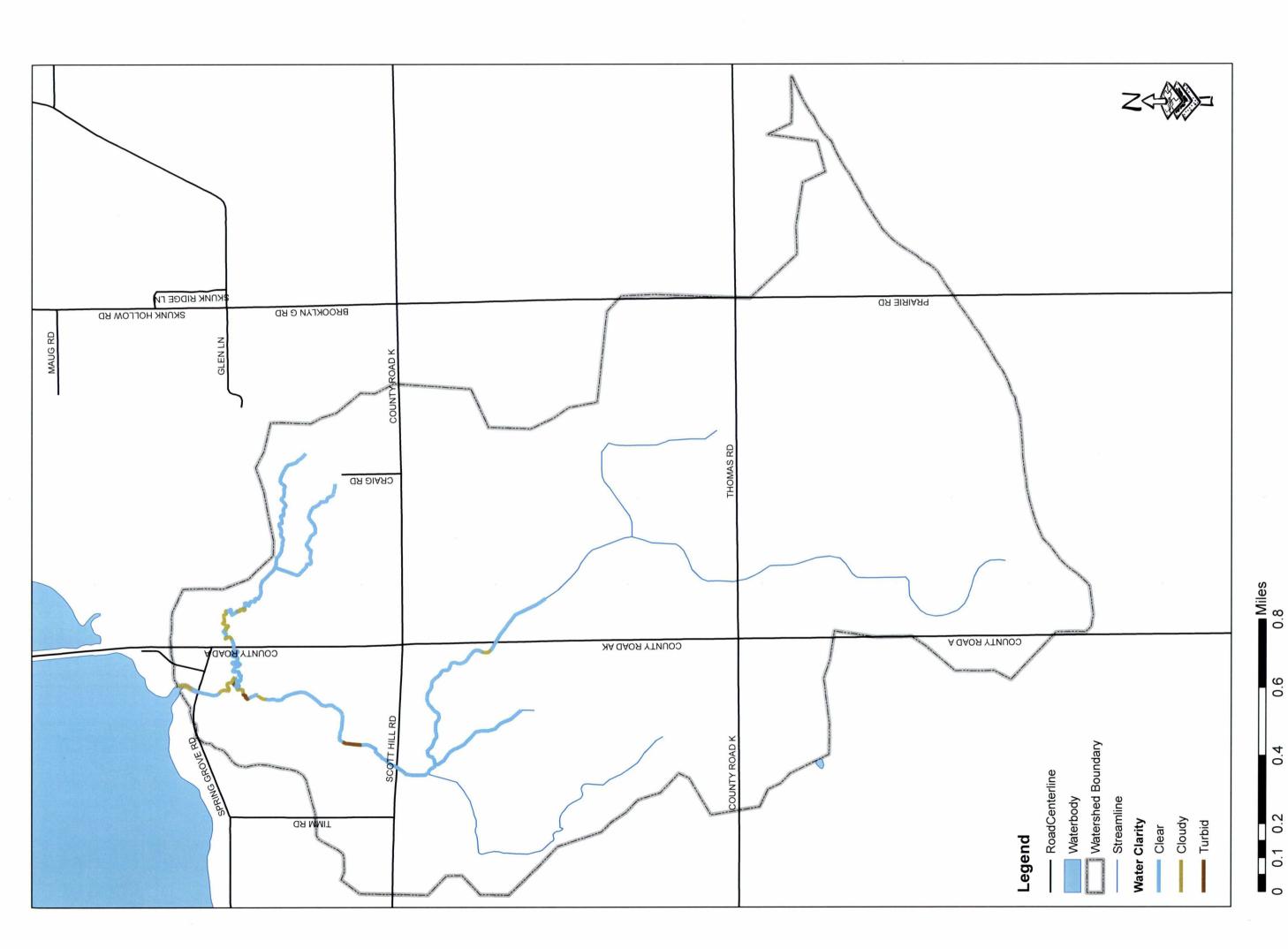


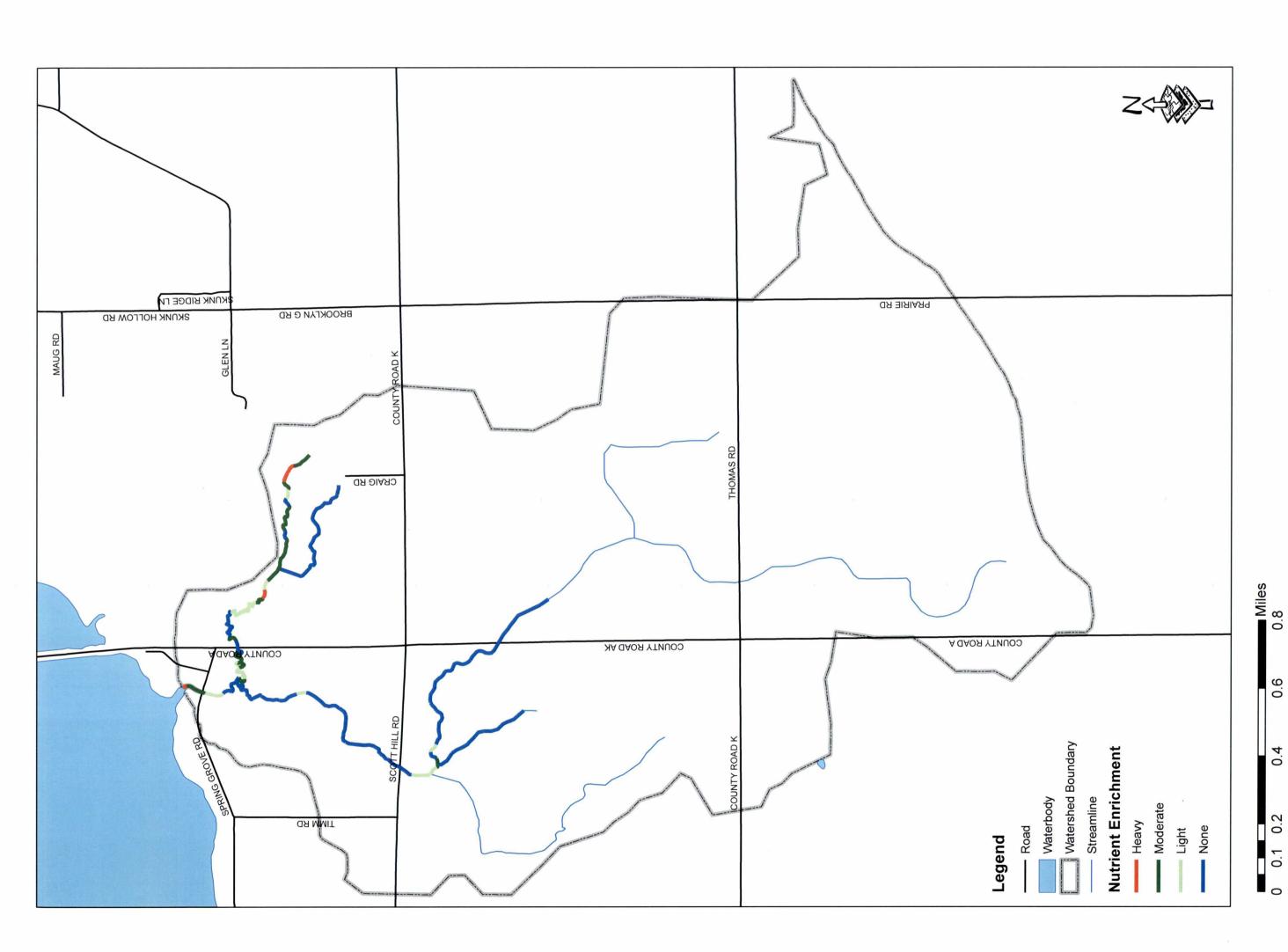


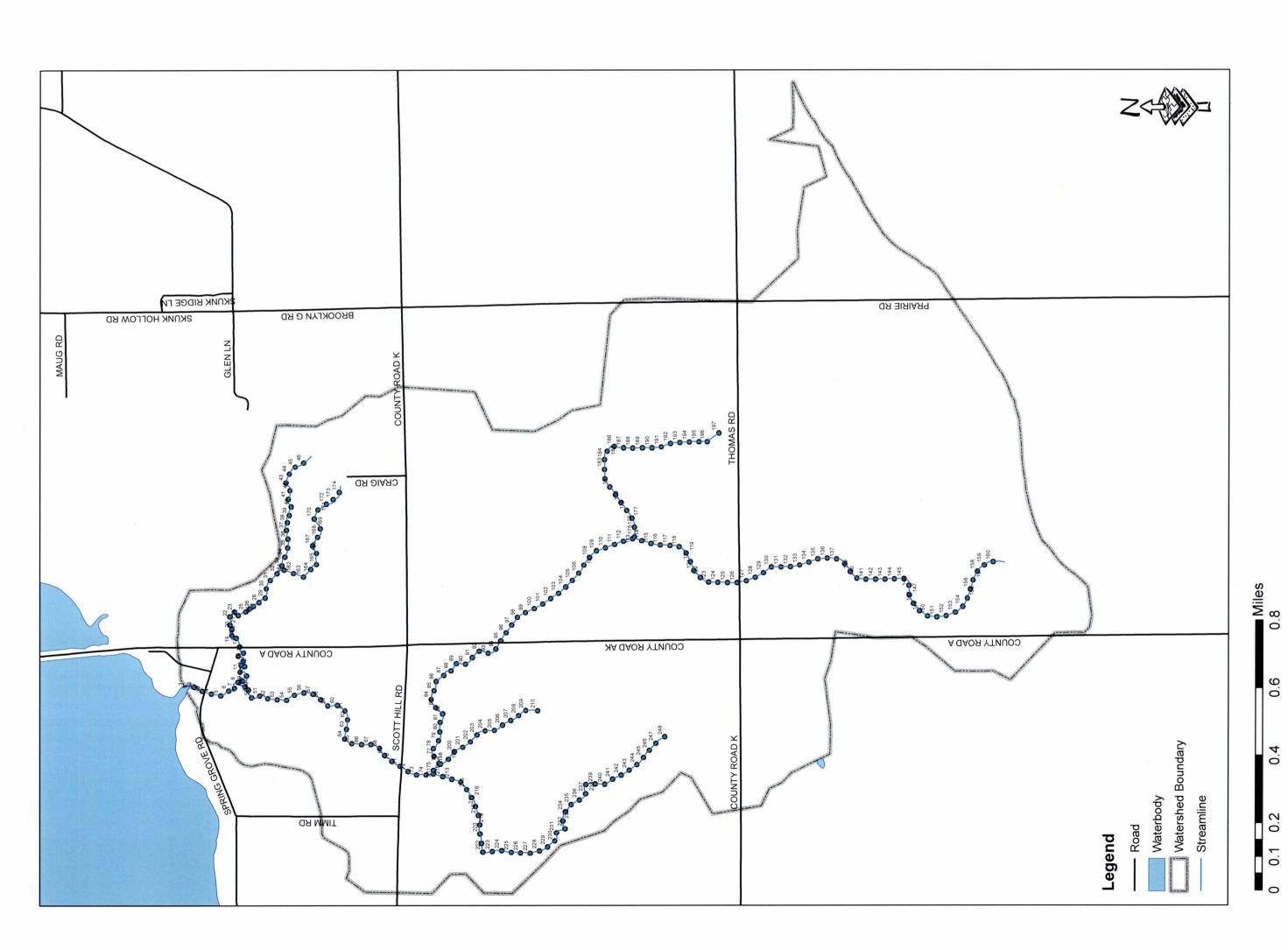


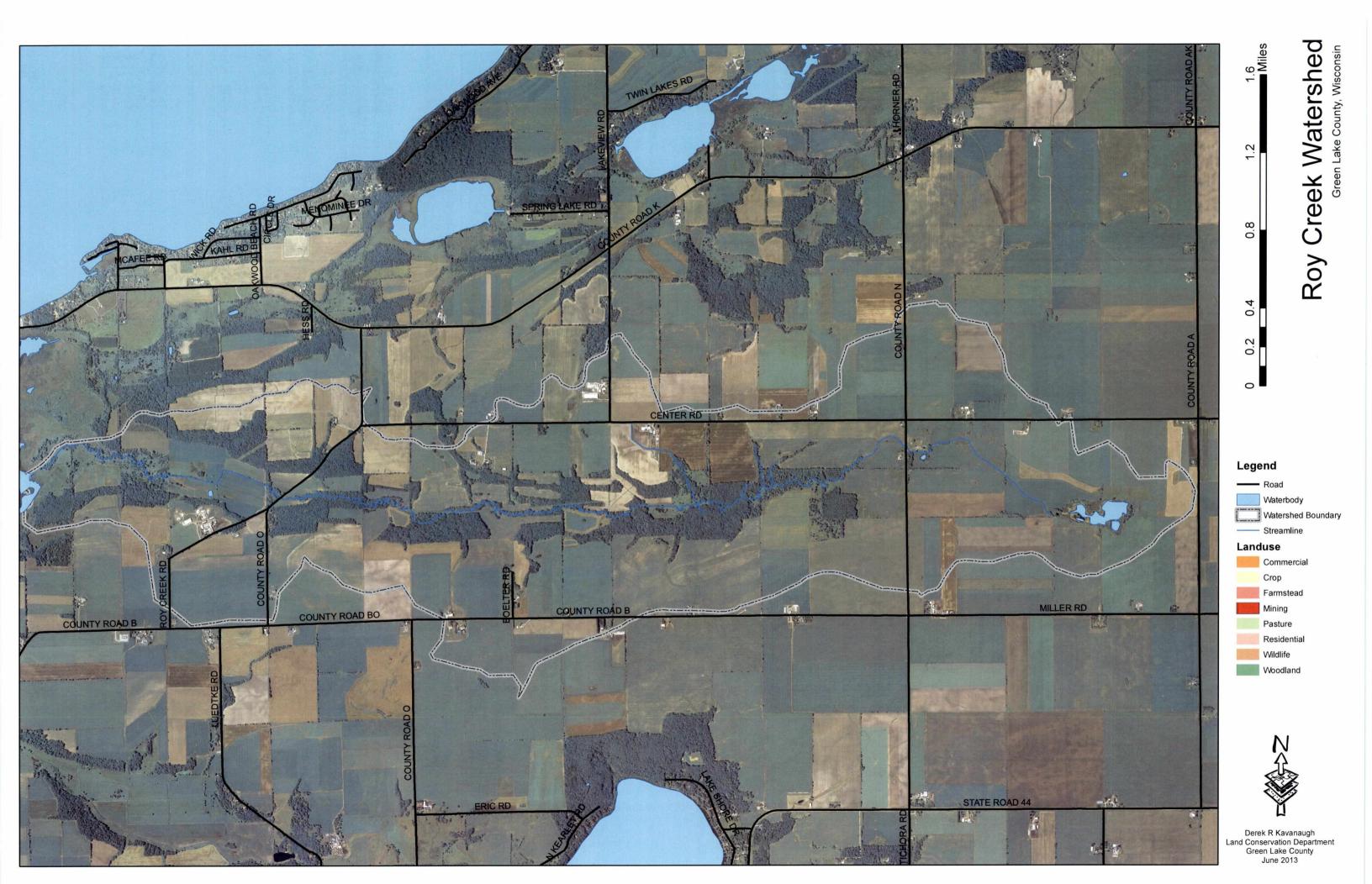


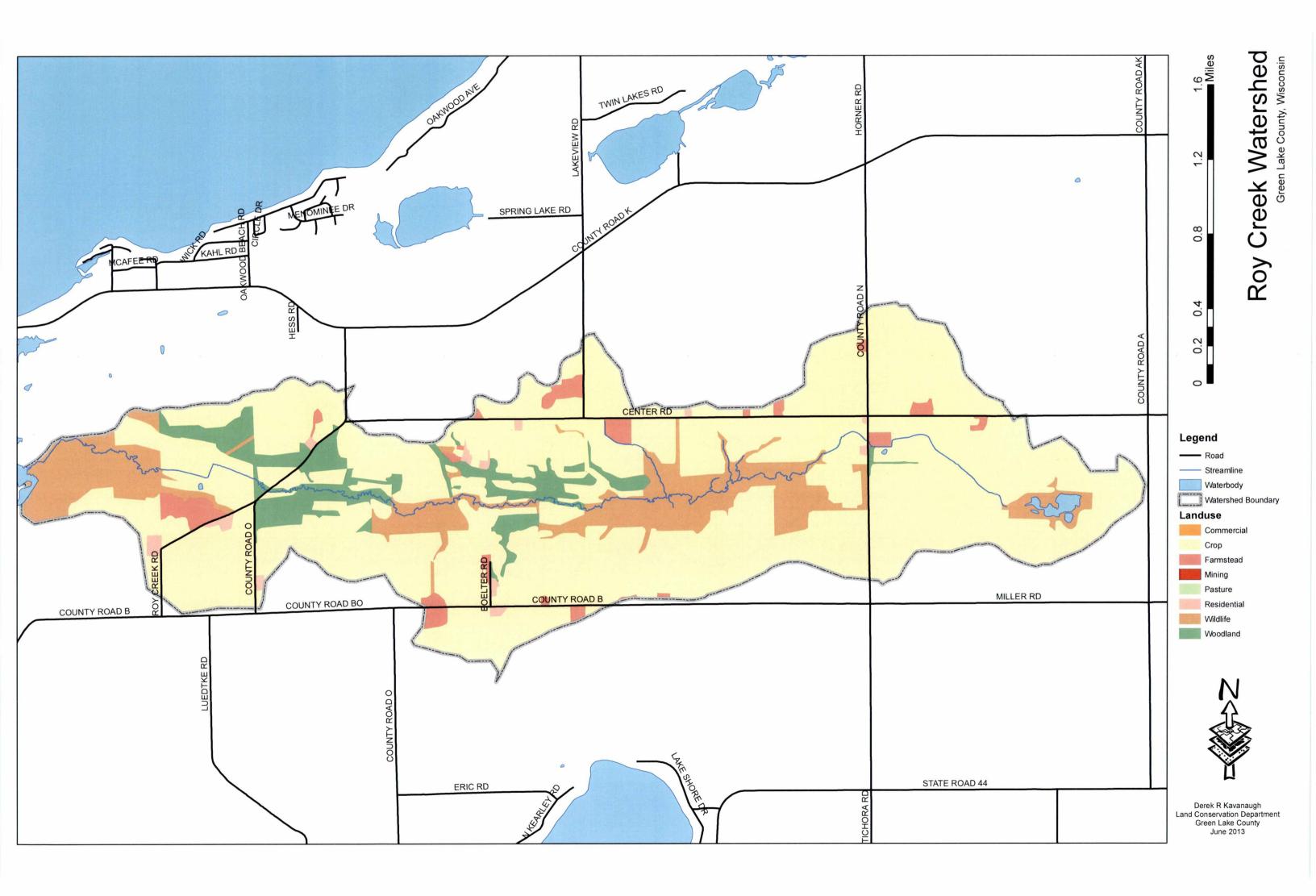


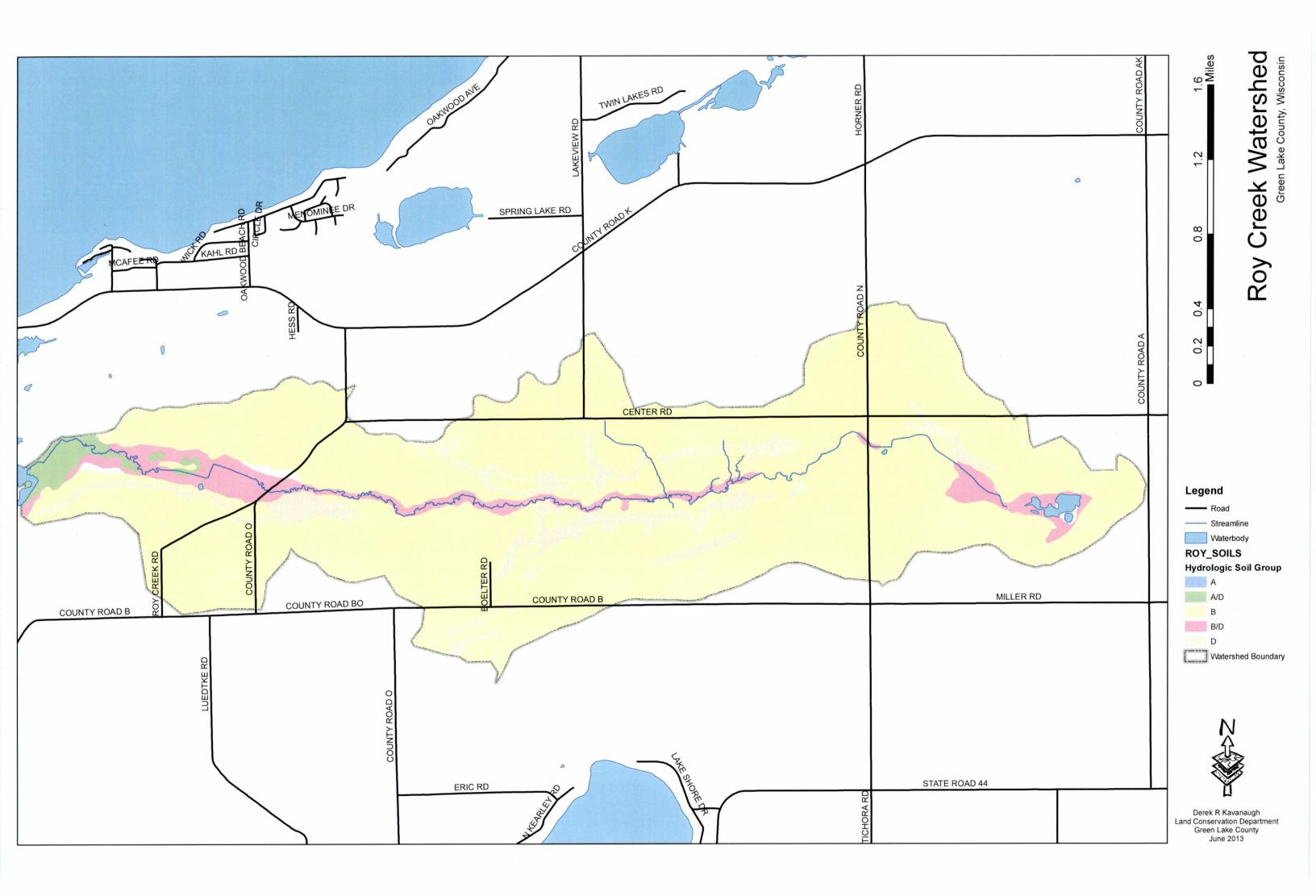


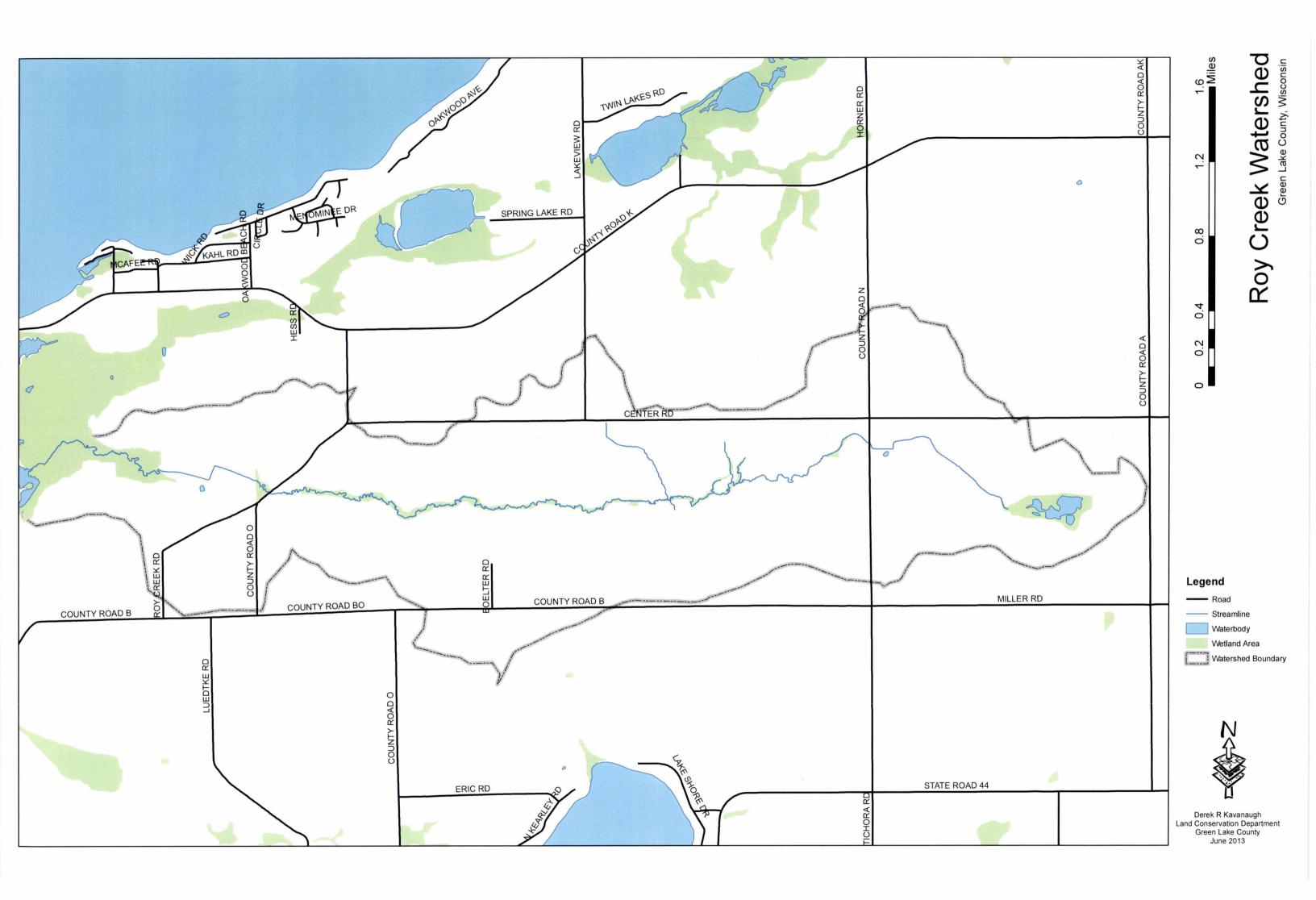


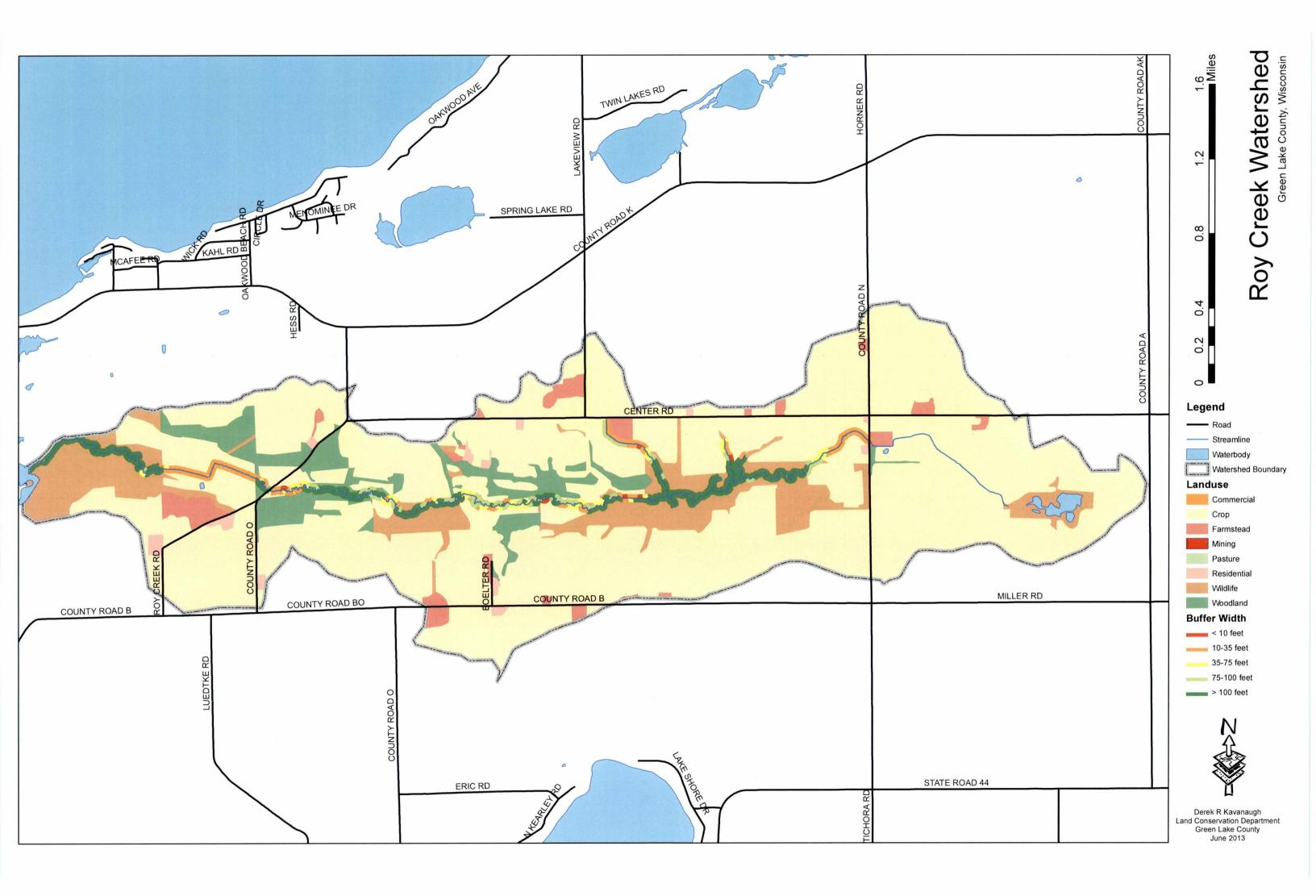


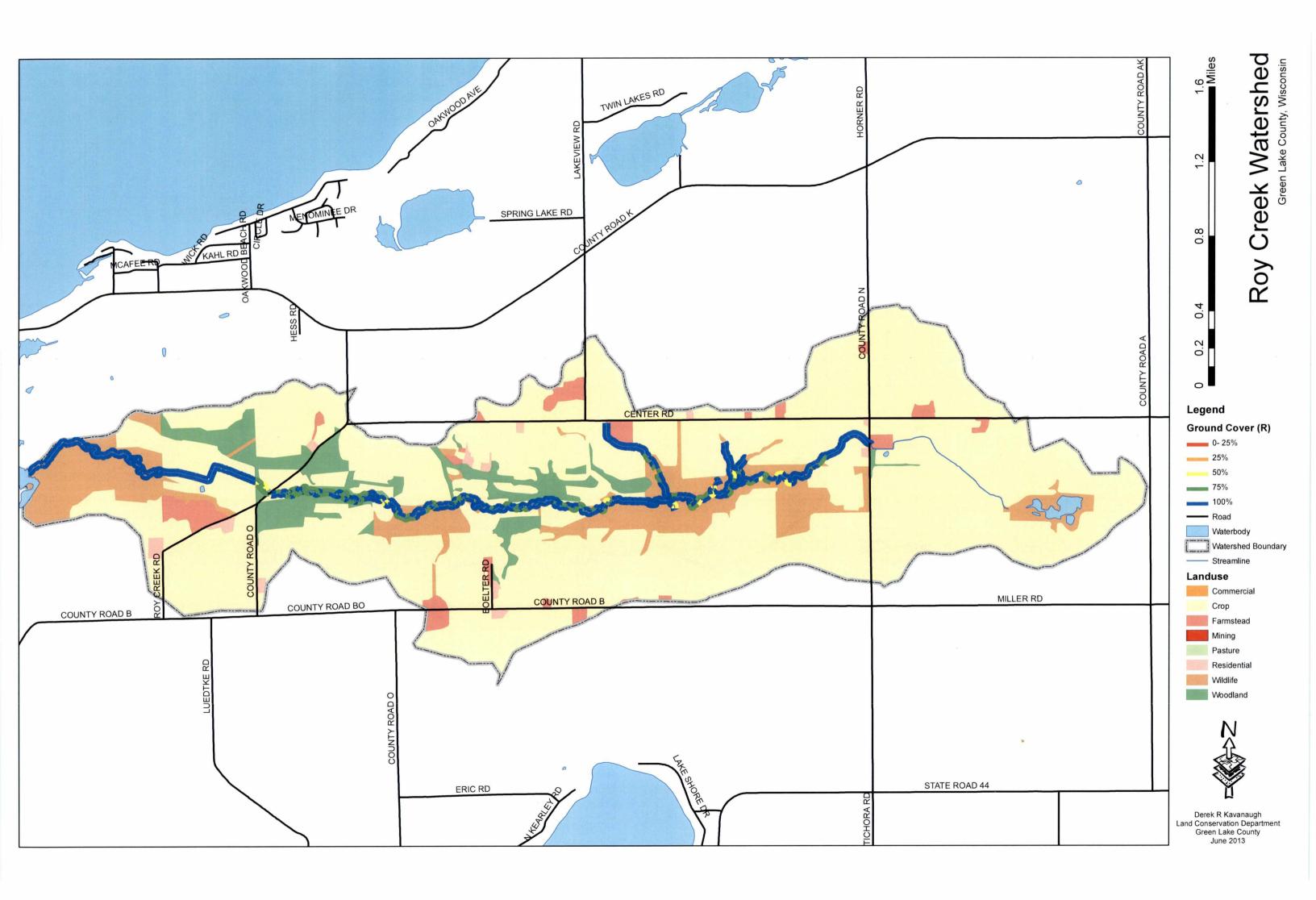


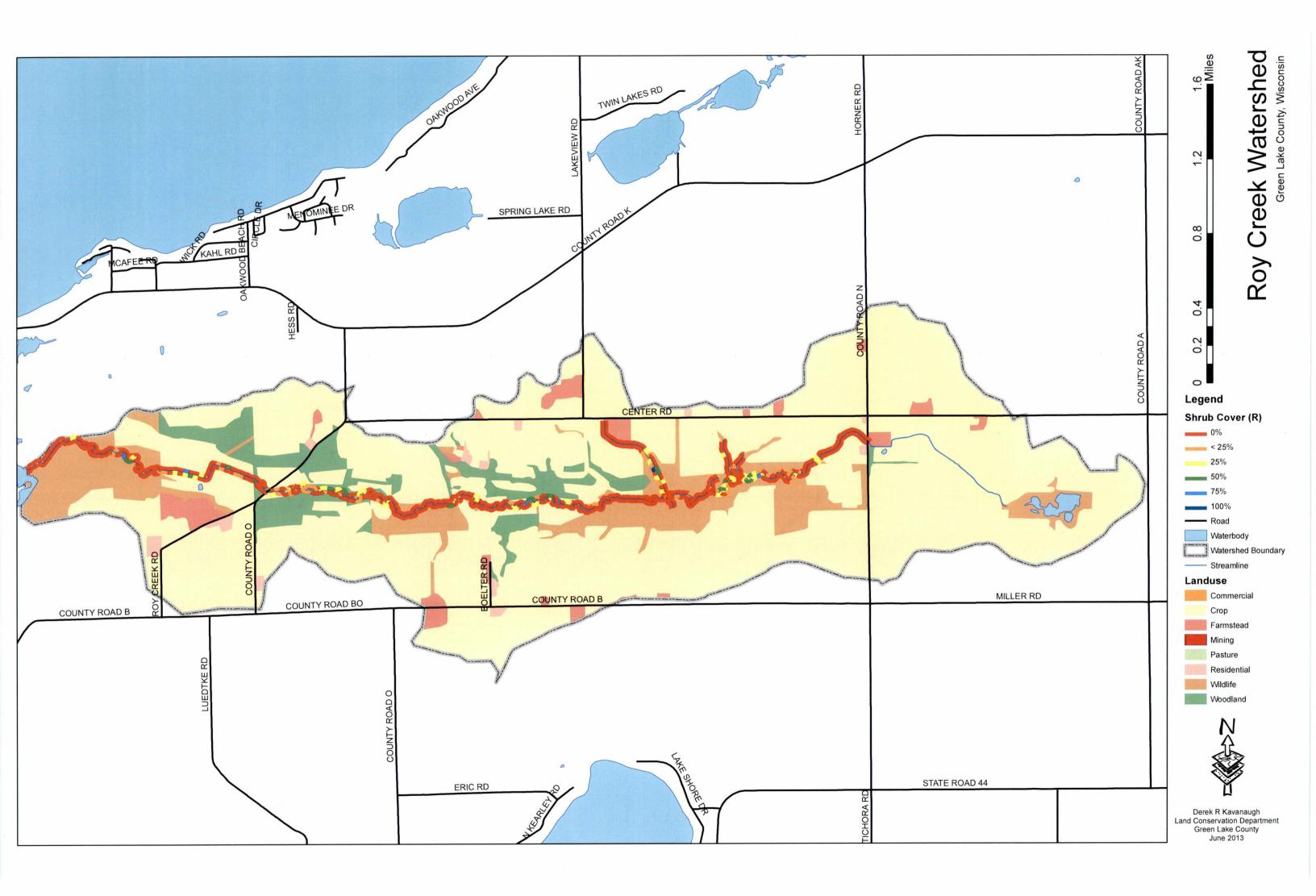


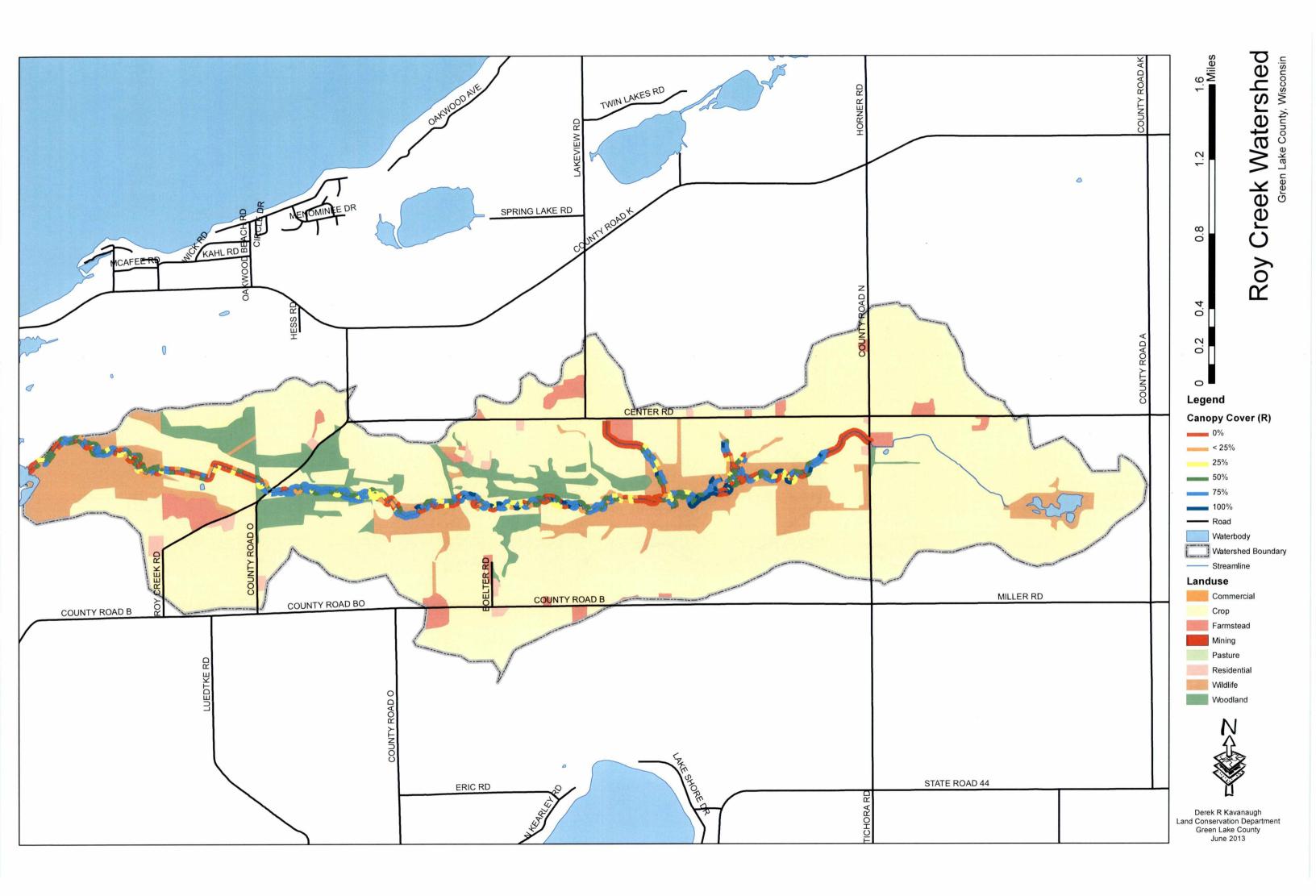


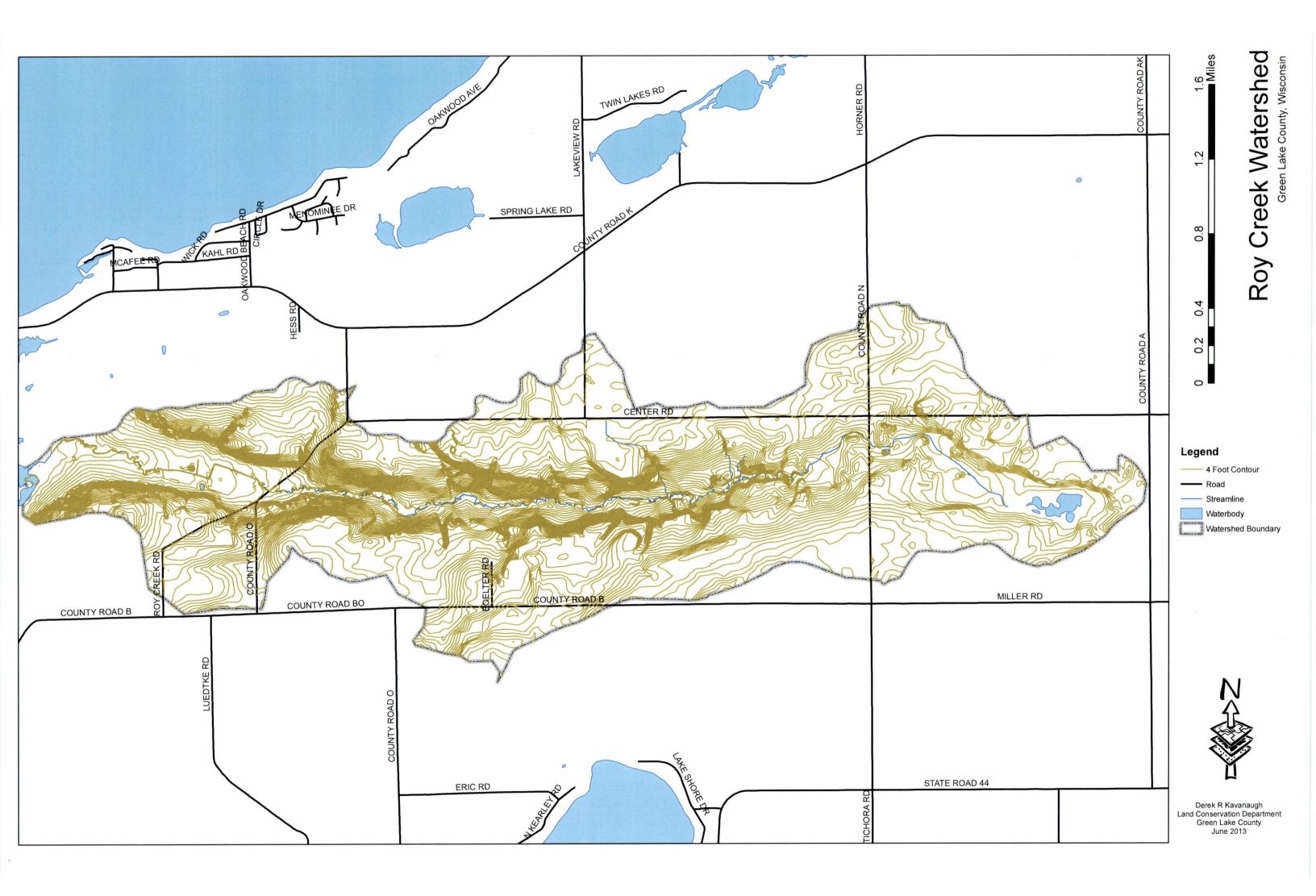


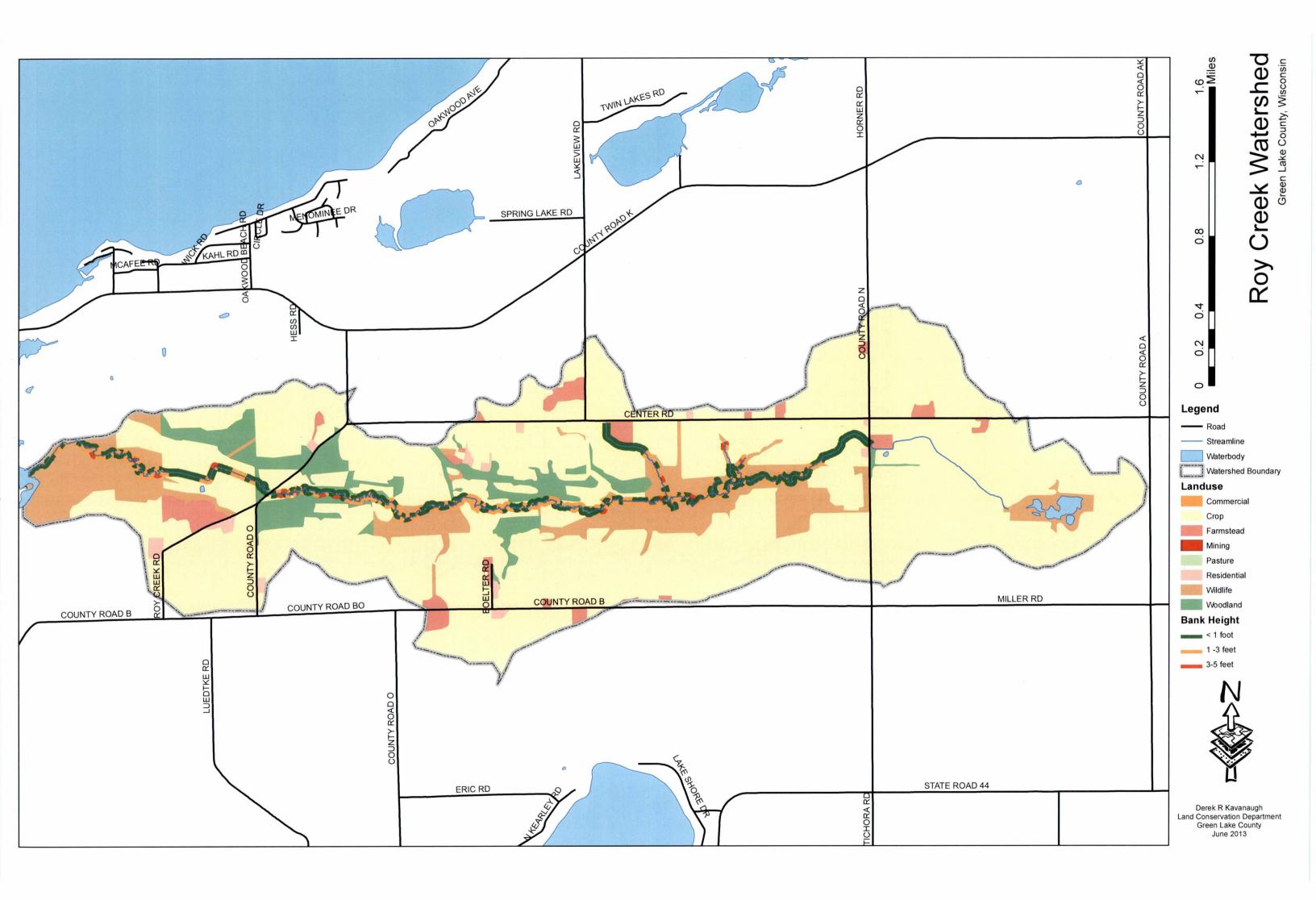


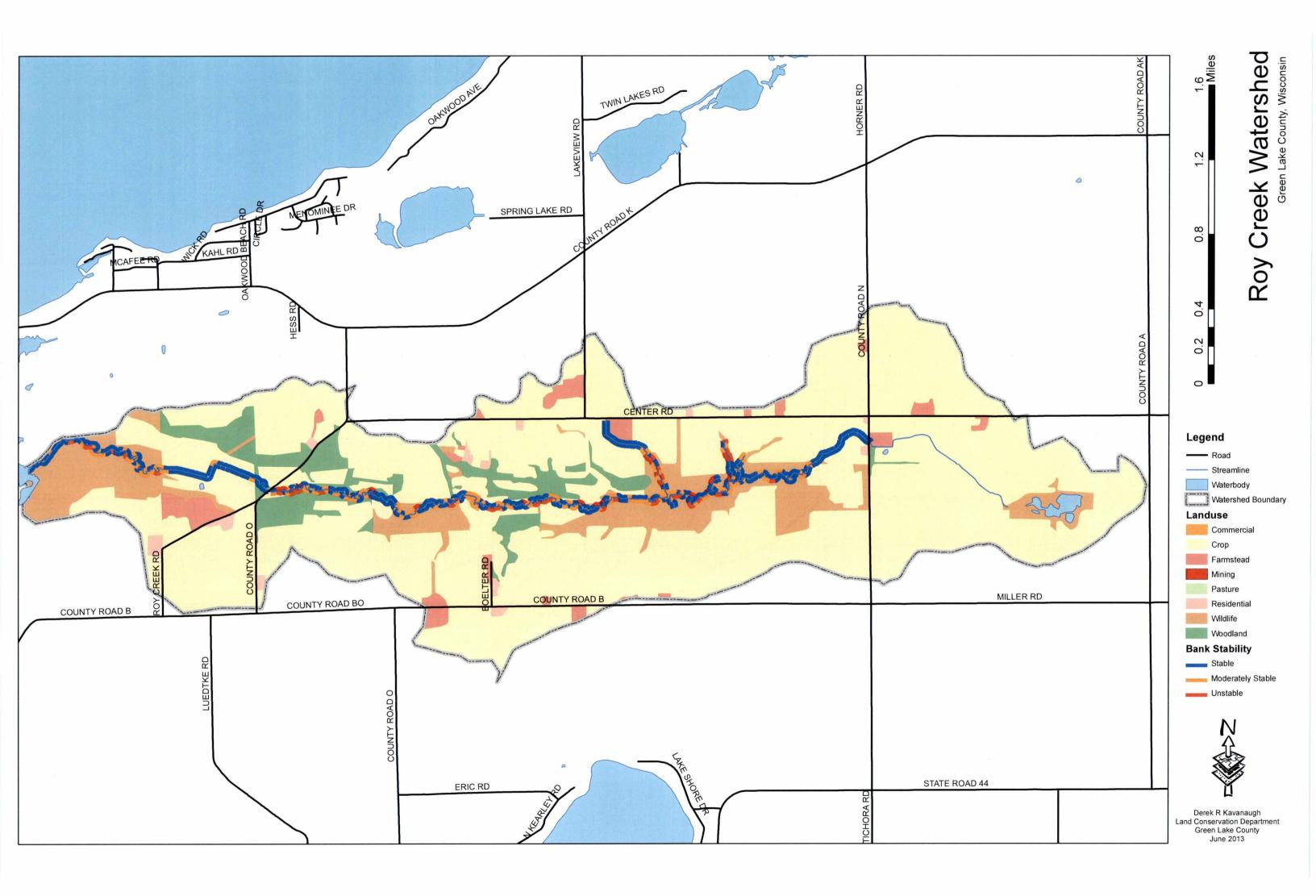


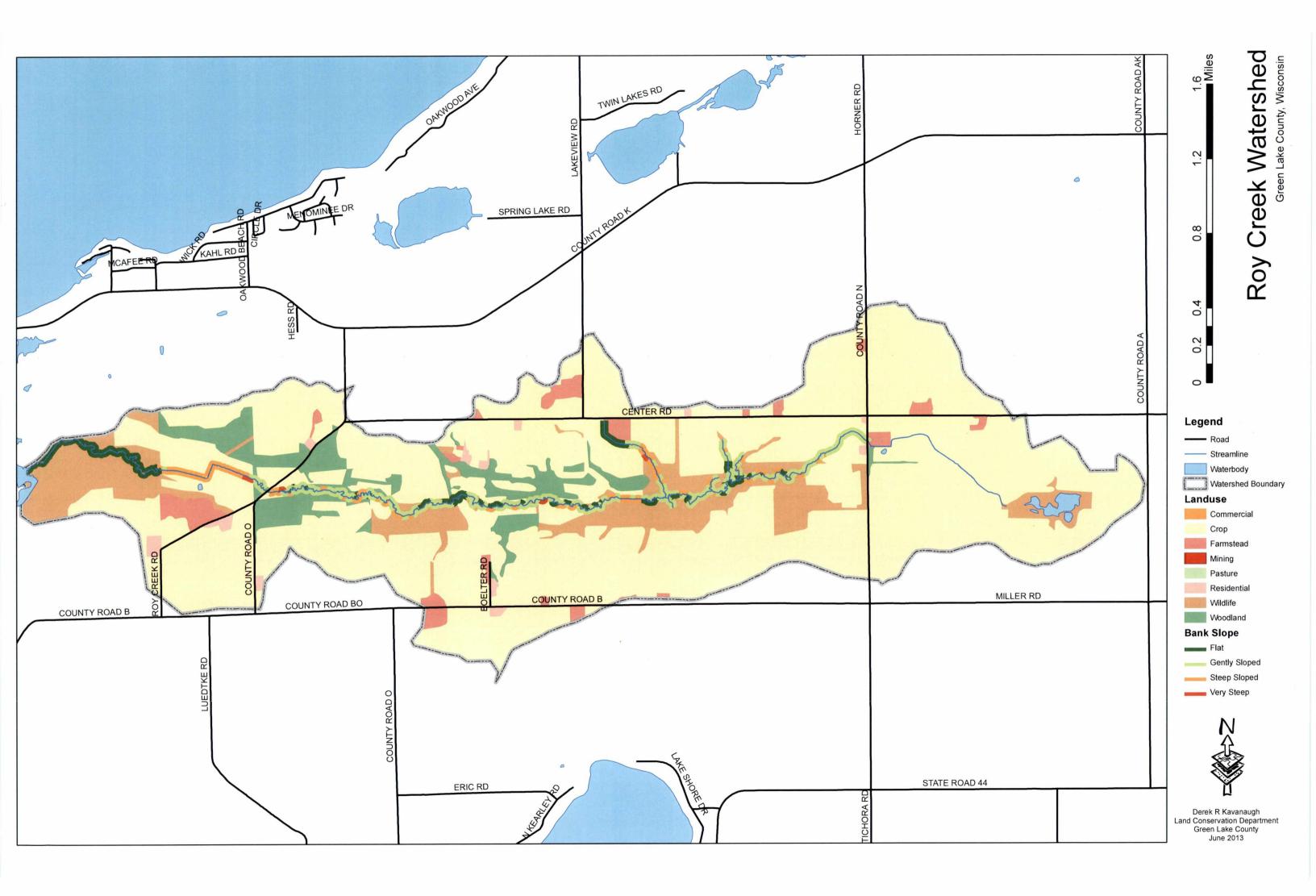


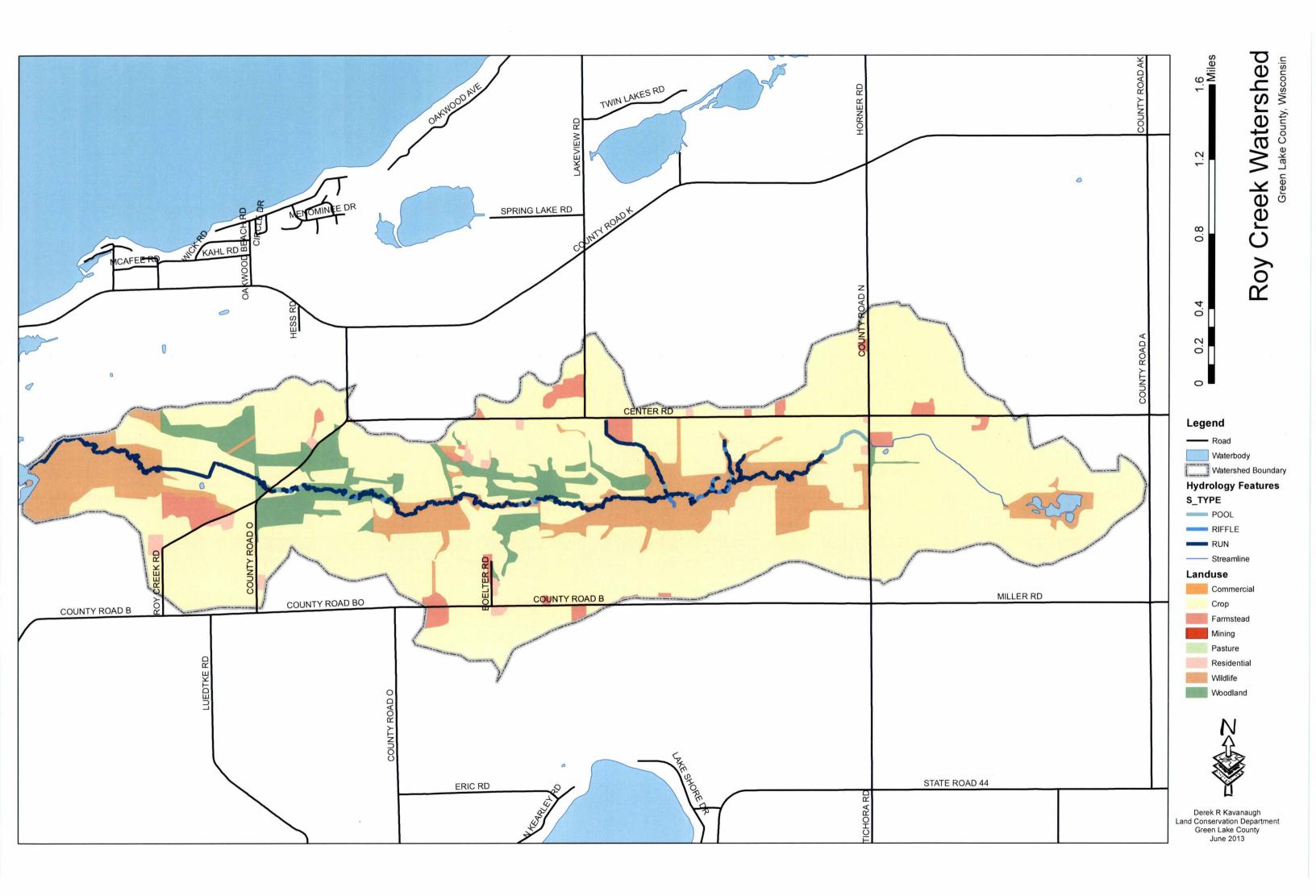


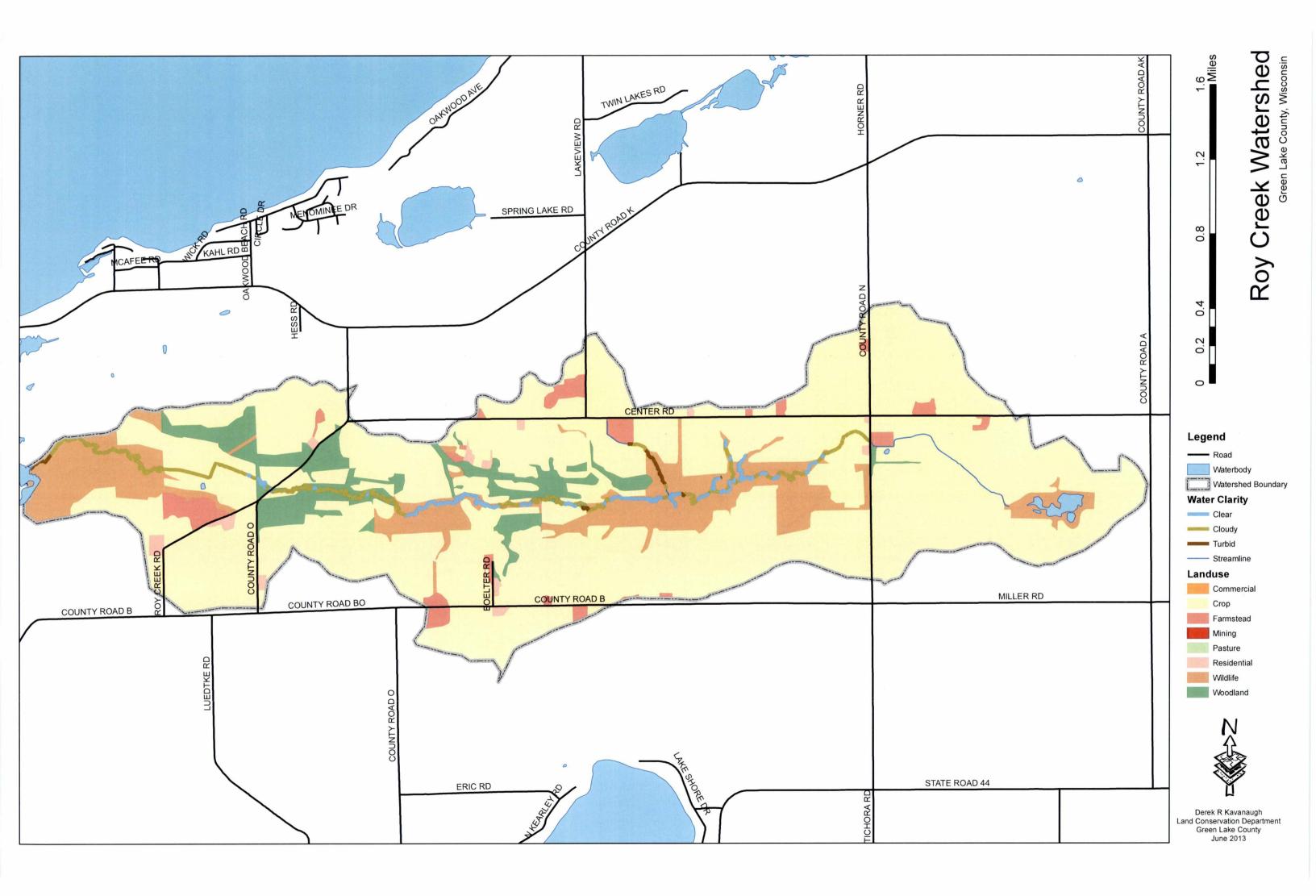


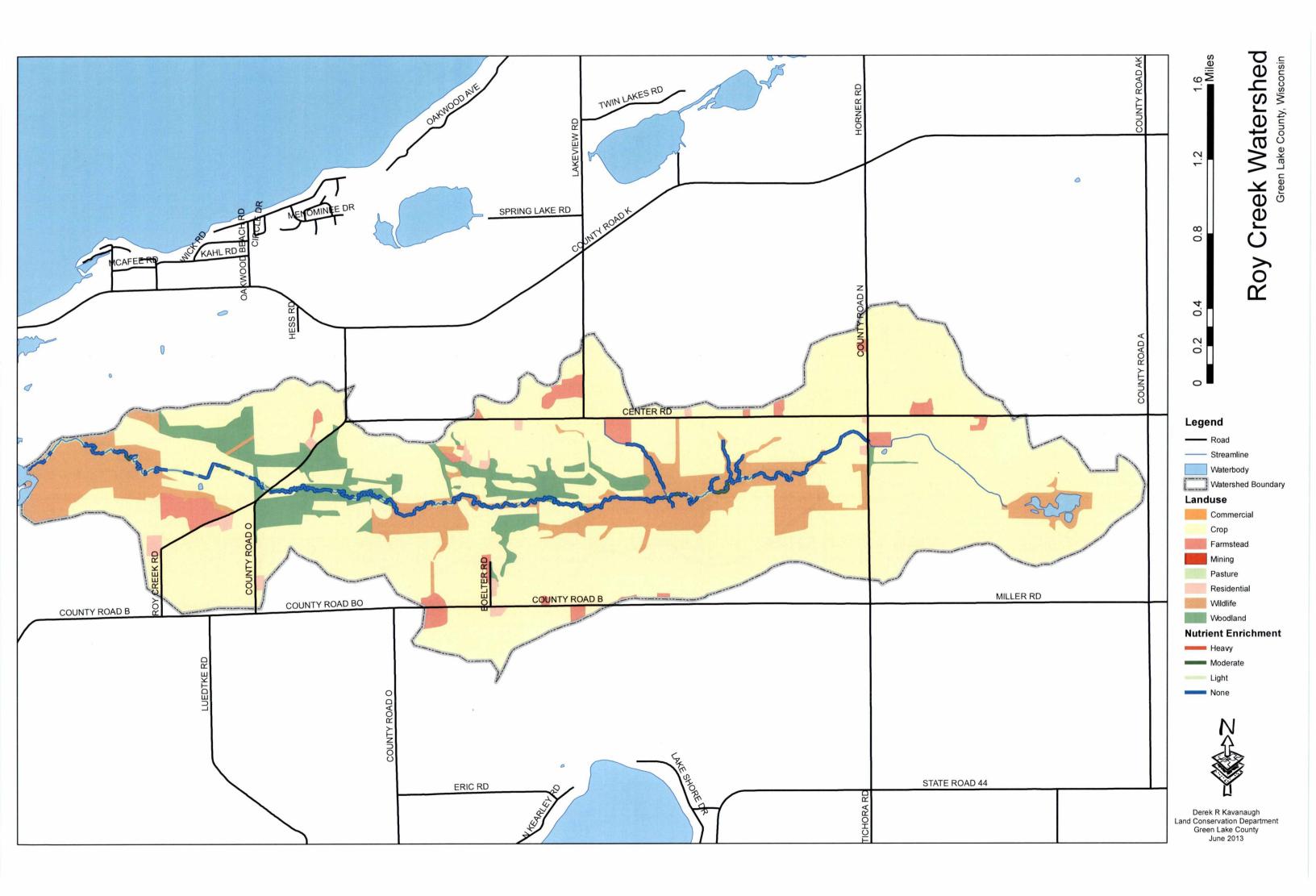


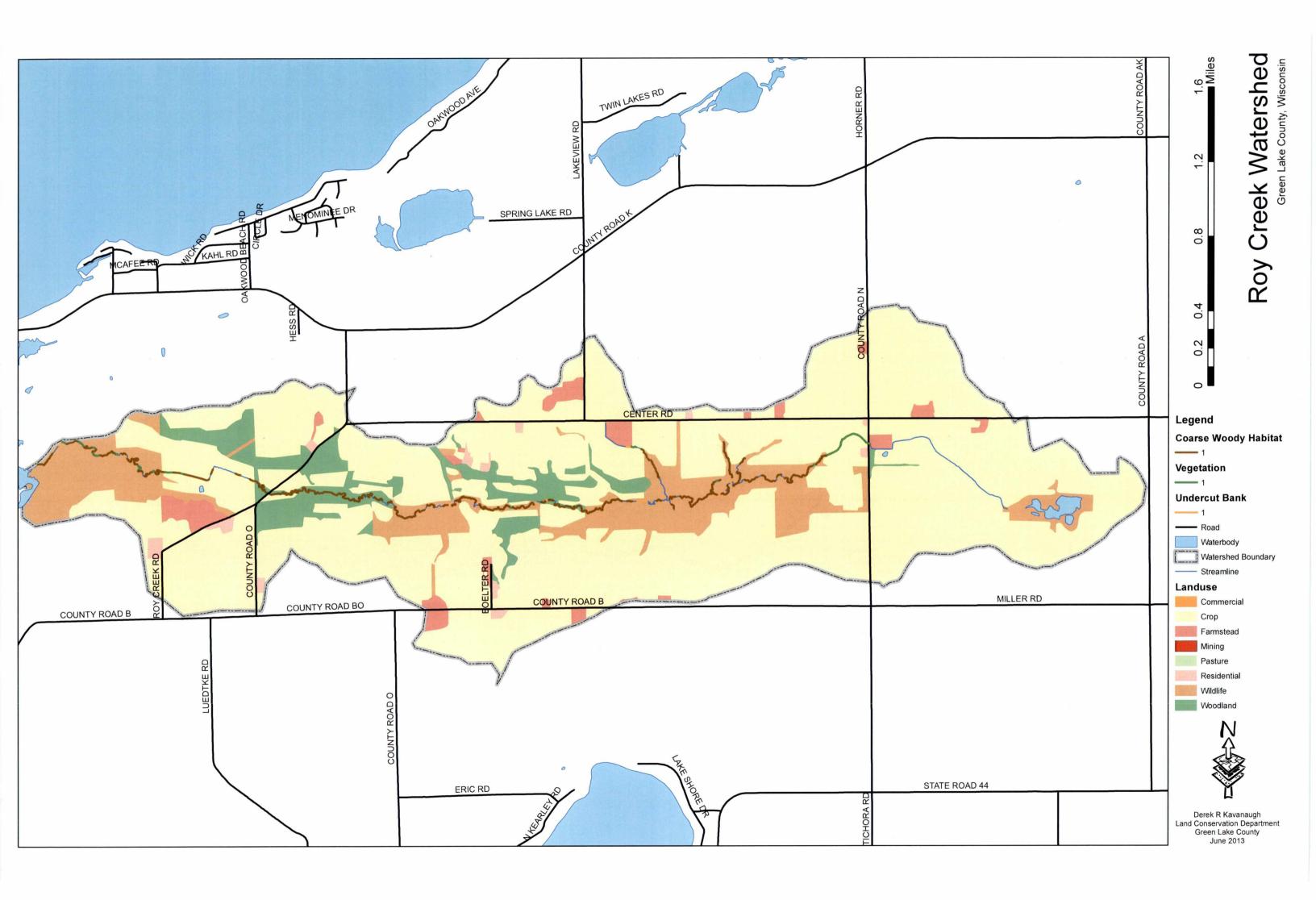


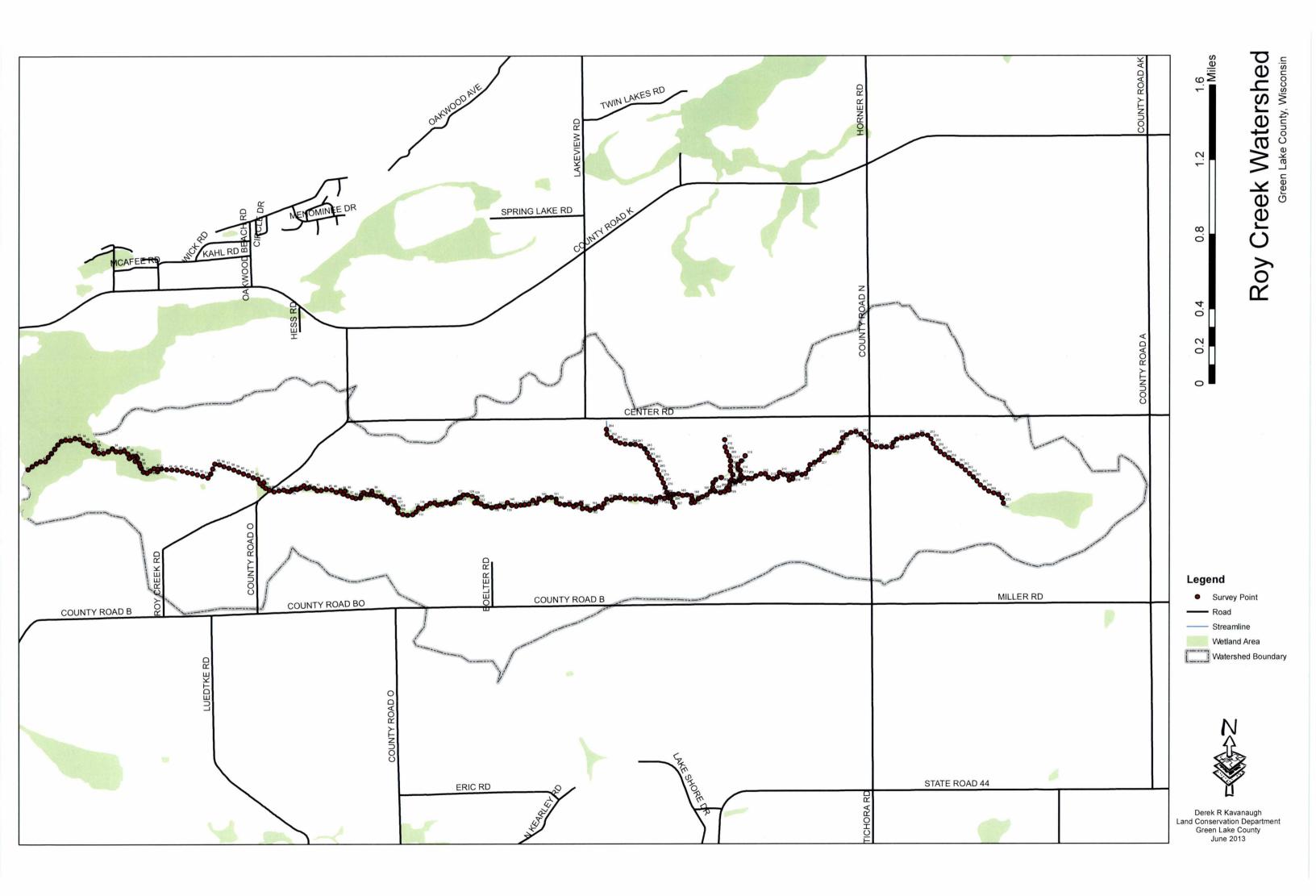




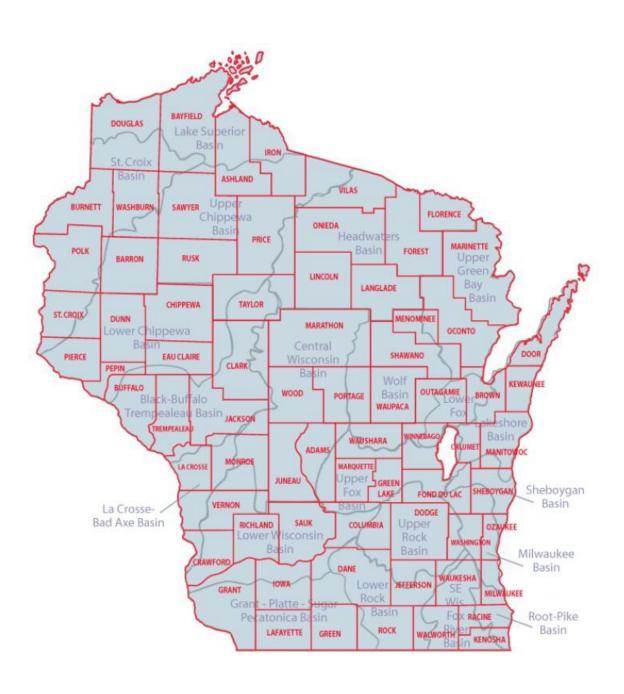








#### **Total Phosphorus Monitoring Annual Reports 2014**

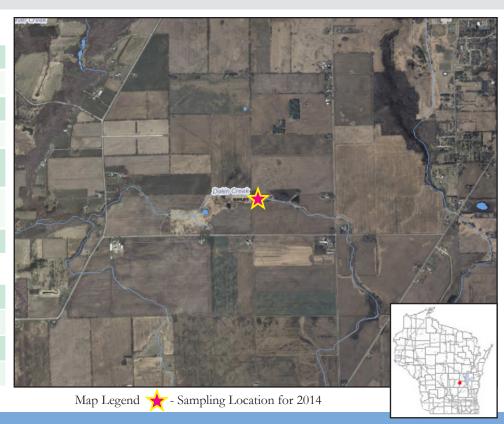


### **Monitoring Site Information**

8			
SWIMS Station ID	10037918		
County	Fond du Lac		
Watershed	Big Green Lake		
Watershed Area	107.31 sq miles		
Total Stream Miles in Watershed	141.08 miles		
Downstream Waterbody	Green Lake		
Volunteer(s)	Eric Godfrey		

### 2014 Monitoring Results

Min TP Value	$0.0351~\mathrm{mg/L}$
Max TP Value	$0.108~\mathrm{mg/L}$
Median TP Value	$0.0465~\mathrm{mg/L}$
No. Samples $> 0.075 \text{ mg/L}$	1



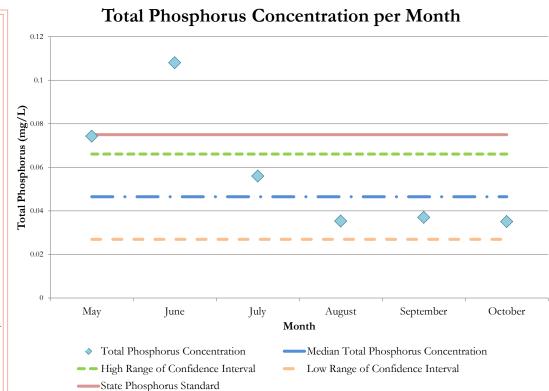


Volunteer Total Phosphorus Monitoring

## Dakin Creek W of Fond du Lac County Line

2014 Monitoring Results

The goal of this project is to assess the water quality condition of both the Silver Creek and the Dakin Creek Watersheds to determine if portions of the watersheds should be placed on the 303(d) Impaired Waters list in 2018. Water Action Volunteers and DNR staff participated in collecting the total phosphorus samples during 2014 and the water quality data will be shared with other interested agencies and groups, such as USGS, Green Lake County LWCD, Green Lake Association, NRCS, and Green Lake Sanitary District.



### Why Phosphorus?

Phosphorus is an essential nutrient responsible for plant growth, but it is also the most visible, widespread water pollutant in Wisconsin lakes. Small increases in phosphorus levels in a lake can bring about substantial increases in aquatic plant and algae growth, which in turn can reduce the recreational use and aquatic biodiversity of said lake. When the excess plants die and are decomposed, oxygen levels in the water drop dramatically which can lead to fish kills.

Additionally, one of the most common impairments in Wisconsin's streams is excess sediments that cover stream bottoms. Since phosphorus moves attached to sediments, it is intimately connected with this source of pollution in our streams. Phosphorus originates naturally from rocks, but its major sources in streams and lakes today are usually associated with human activities: soil erosion, human and animal wastes, septic systems, and runoff from farmland or lawns. Phosphorus-containing contaminants from urban streets and parking lots such as food waste, detergents, and paper products are also potential sources of phosphorus pollution from the surrounding landscape. The impact that phosphorus can have in streams is less apparent than in lakes due to the overall movement of water, but in areas with slow velocity, where sediment can settle and deposit along the bottom substrate, algae blooms can result.







Photo credits to Matt Berg, David Seligman, Linda Warren, and Adrian Konell

### **Volunteer Monitoring Protocol**

To assess in stream phosphorus levels, WAV volunteers collected water samples that were analyzed for total phosphorus (TP) at the State Lab of Hygiene during the growing season (May through October). Following Wisconsin Department of Natural Resources (WDNR) methods, six phosphorus water samples were collected at each monitoring site - one per month for each of the six months during the growing season, The water samples were collected approximately 30 days apart and no samples were collected within 15 days of one another.

Total phosphorus impairment is assessed using the criteria in the table below.



The total phosphorus criteria is exceeded if the lower confidence limit of the sample median exceeds the state total phosphorus criteria of 0.075 mg/L.



The site is classified as Watch Waters if the median total phosphorus concentration falls within the confidence limit.



The total phosphorus criteria is met if the upper limit of the confidence interval does not exceed the state total phosphorus criteria of 0.075 mg/L.



### PROJECT PARTNERS



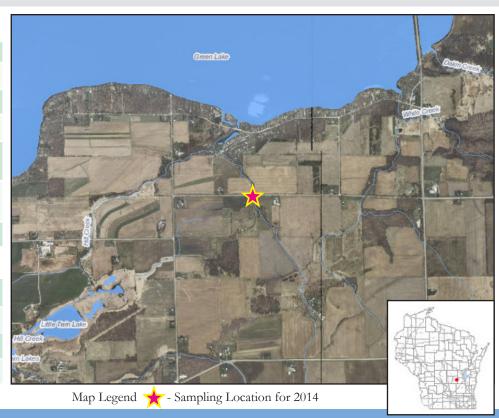


### **Monitoring Site Information**

8				
<b>SWIMS Station ID</b>	10041578			
County	Green Lake			
Watershed	Big Green Lake			
Watershed Area	107.31 sq miles			
Total Stream Miles in Watershed	141.08 miles			
Downstream Waterbody	Hill Creek			
Volunteer(s)	Dave Kummerow			

### 2014 Monitoring Results

Min TP Value	$0.0165~\mathrm{mg/L}$
Max TP Value	$0.0841~\mathrm{mg/L}$
Median TP Value	$0.05985~\mathrm{mg/L}$
No. Samples $\geq 0.075 \text{ mg/L}$	2.



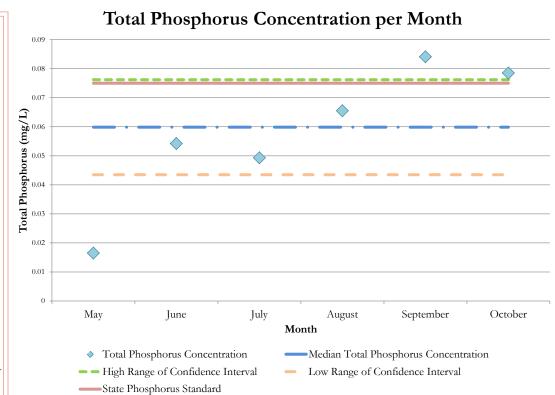


Volunteer Total Phosphorus Monitoring

### Unnamed Trib to Hill Creek US Scott Hill Road

2014 Monitoring Results

This Targeted Watershed Assessment (TWA) will address needs for baseline water quality monitoring of the watershed on the south side of Big Green Lake. This watershed discharges anywhere between 15-30% of the total phosphorus load into Big Green Lake. Considerable stream restoration and watershed BMP work has been done and is projected for the future by multiple agencies, including Green Lake County LWCD and NRCS.



### Why Phosphorus?

Phosphorus is an essential nutrient responsible for plant growth, but it is also the most visible, widespread water pollutant in Wisconsin lakes. Small increases in phosphorus levels in a lake can bring about substantial increases in aquatic plant and algae growth, which in turn can reduce the recreational use and aquatic biodiversity of said lake. When the excess plants die and are decomposed, oxygen levels in the water drop dramatically which can lead to fish kills.

Additionally, one of the most common impairments in Wisconsin's streams is excess sediments that cover stream bottoms. Since phosphorus moves attached to sediments, it is intimately connected with this source of pollution in our streams. Phosphorus originates naturally from rocks, but its major sources in streams and lakes today are usually associated with human activities: soil erosion, human and animal wastes, septic systems, and runoff from farmland or lawns. Phosphorus-containing contaminants from urban streets and parking lots such as food waste, detergents, and paper products are also potential sources of phosphorus pollution from the surrounding landscape. The impact that phosphorus can have in streams is less apparent than in lakes due to the overall movement of water, but in areas with slow velocity, where sediment can settle and deposit along the bottom substrate, algae blooms can result.







Photo credits to Matt Berg, David Seligman, Linda Warren, and Adrian Konell

### **Volunteer Monitoring Protocol**

To assess in stream phosphorus levels, WAV volunteers collected water samples that were analyzed for total phosphorus (TP) at the State Lab of Hygiene during the growing season (May through October). Following Wisconsin Department of Natural Resources (WDNR) methods, six phosphorus water samples were collected at each monitoring site - one per month for each of the six months during the growing season, The water samples were collected approximately 30 days apart and no samples were collected within 15 days of one another.

Total phosphorus impairment is assessed using the criteria in the table below.



The total phosphorus criteria is exceeded if the lower confidence limit of the sample median exceeds the state total phosphorus criteria of 0.075 mg/L.



The site is classified as Watch Waters if the median total phosphorus concentration falls within the confidence limit.



The total phosphorus criteria is met if the upper limit of the confidence interval does not exceed the state total phosphorus criteria of 0.075 mg/L.



### PROJECT PARTNERS



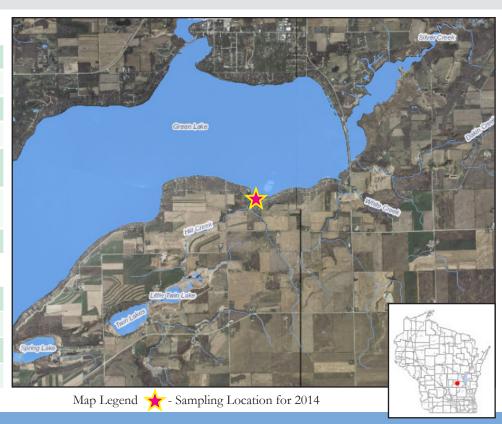


### **Monitoring Site Information**

10033838
Green Lake
Big Green Lake
107.31 sq miles
141.08 miles
Green Lake
Derek Kavanaugh

### 2014 Monitoring Results

Min TP Value	$0.0971~\mathrm{mg/L}$
Max TP Value	$0.2~\mathrm{mg/L}$
Median TP Value	$0.148~\mathrm{mg/L}$
No. Samples $> 0.075 \text{ mg/L}$	6



TP Criteria Exceeded Volunteer Total Phosphorus Monitoring

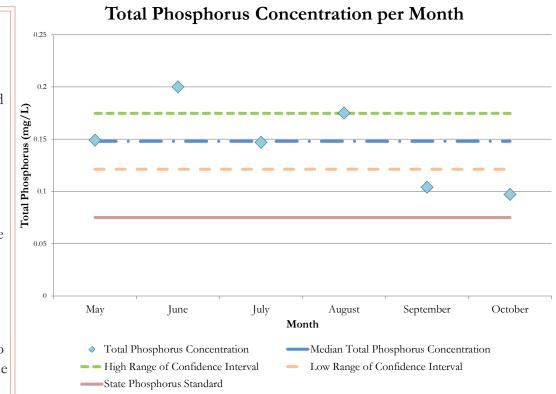
# Hill Creek upstream Spring Grove Road

2014 Monitoring Results

Many of Wisconsin's water quality standards require multiple visits to make an assessment decision. Every year, several stream sites are

Every year, several stream sites are monitored and the field data collected during each visit are used to "flag" problem waters. In the next year, follow up monitoring is carried out on the "flagged" waters where the data suggest there is an impairment, but there are insufficient data to make that determination based on the State's minimum data requirements. In 2014, Water Action Volunteers stream monitors who live nearby the recommended follow up sites were asked to assist in the monitoring process by collecting water samples to be analyzed for total phosphorus at the

Wisconsin State Lab of Hygiene.



### Why Phosphorus?

Phosphorus is an essential nutrient responsible for plant growth, but it is also the most visible, widespread water pollutant in Wisconsin lakes. Small increases in phosphorus levels in a lake can bring about substantial increases in aquatic plant and algae growth, which in turn can reduce the recreational use and aquatic biodiversity of said lake. When the excess plants die and are decomposed, oxygen levels in the water drop dramatically which can lead to fish kills.

Additionally, one of the most common impairments in Wisconsin's streams is excess sediments that cover stream bottoms. Since phosphorus moves attached to sediments, it is intimately connected with this source of pollution in our streams. Phosphorus originates naturally from rocks, but its major sources in streams and lakes today are usually associated with human activities: soil erosion, human and animal wastes, septic systems, and runoff from farmland or lawns. Phosphorus-containing contaminants from urban streets and parking lots such as food waste, detergents, and paper products are also potential sources of phosphorus pollution from the surrounding landscape. The impact that phosphorus can have in streams is less apparent than in lakes due to the overall movement of water, but in areas with slow velocity, where sediment can settle and deposit along the bottom substrate, algae blooms can result.







Photo credits to Matt Berg, David Seligman, Linda Warren, and Adrian Konell

### **Volunteer Monitoring Protocol**

To assess in stream phosphorus levels, WAV volunteers collected water samples that were analyzed for total phosphorus (TP) at the State Lab of Hygiene during the growing season (May through October). Following Wisconsin Department of Natural Resources (WDNR) methods, six phosphorus water samples were collected at each monitoring site - one per month for each of the six months during the growing season, The water samples were collected approximately 30 days apart and no samples were collected within 15 days of one another.

Total phosphorus impairment is assessed using the criteria in the table below.



The total phosphorus criteria is exceeded if the lower confidence limit of the sample median exceeds the state total phosphorus criteria of 0.075 mg/L.



The site is classified as Watch Waters if the median total phosphorus concentration falls within the confidence limit.



The total phosphorus criteria is met if the upper limit of the confidence interval does not exceed the state total phosphorus criteria of 0.075 mg/L.



### **PROJECT PARTNERS**





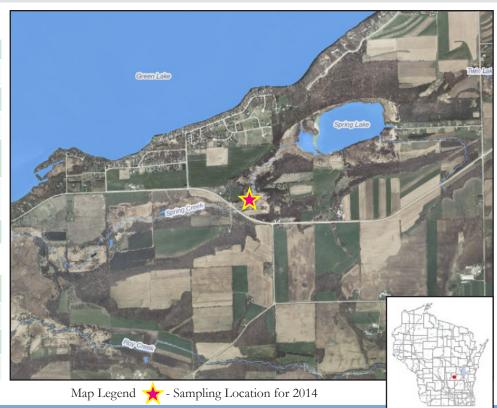


### Monitoring Site Information

manufalling on	
<b>SWIMS Station ID</b>	243026
County	Green Lake
Watershed	Big Green Lake
Watershed Area	107.31 sq miles
Total Stream Miles in Watershed	141.08 miles
Downstream Waterbody	Green Lake
Volunteer(s)	Dave Kummerow

### 2014 Monitoring Results

$\mathcal{O}$	
Min TP Value	$0.0141~\mathrm{mg/L}$
Max TP Value	$0.0253~\mathrm{mg/L}$
Median TP Value	$0.01925~\mathrm{mg/L}$
No. Samples $> 0.075 \text{ mg/L}$	0



Volunteer Total Phosphorus Monitoring

# Spring Creek at County K

2014 Monitoring Results



This Targeted Watershed Assessment (TWA) will address needs for baseline water quality monitoring of the watershed on the south side of Big Green Lake. This watershed discharges anywhere between 15-30% of the total phosphorus load into Big Green Lake. Considerable stream restoration and watershed BMP work has been done and is projected for the future by multiple agencies,

LWCD and NRCS.

### Total Phosphorus Concentration per Month 0.06 Total Phosphorus (mg/L) 0.02 0.01 May June July August September October Month Total Phosphorus Concentration Median Total Phosphorus Concentration including Green Lake County High Range of Confidence Interval Low Range of Confidence Interval ——State Phosphorus Standard

### Why Phosphorus?

Phosphorus is an essential nutrient responsible for plant growth, but it is also the most visible, widespread water pollutant in Wisconsin lakes. Small increases in phosphorus levels in a lake can bring about substantial increases in aquatic plant and algae growth, which in turn can reduce the recreational use and aquatic biodiversity of said lake. When the excess plants die and are decomposed, oxygen levels in the water drop dramatically which can lead to fish kills.

Additionally, one of the most common impairments in Wisconsin's streams is excess sediments that cover stream bottoms. Since phosphorus moves attached to sediments, it is intimately connected with this source of pollution in our streams. Phosphorus originates naturally from rocks, but its major sources in streams and lakes today are usually associated with human activities: soil erosion, human and animal wastes, septic systems, and runoff from farmland or lawns. Phosphorus-containing contaminants from urban streets and parking lots such as food waste, detergents, and paper products are also potential sources of phosphorus pollution from the surrounding landscape. The impact that phosphorus can have in streams is less apparent than in lakes due to the overall movement of water, but in areas with slow velocity, where sediment can settle and deposit along the bottom substrate, algae blooms can result.







Photo credits to Matt Berg, David Seligman, Linda Warren, and Adrian Konell

### **Volunteer Monitoring Protocol**

To assess in stream phosphorus levels, WAV volunteers collected water samples that were analyzed for total phosphorus (TP) at the State Lab of Hygiene during the growing season (May through October). Following Wisconsin Department of Natural Resources (WDNR) methods, six phosphorus water samples were collected at each monitoring site - one per month for each of the six months during the growing season, The water samples were collected approximately 30 days apart and no samples were collected within 15 days of one another.

Total phosphorus impairment is assessed using the criteria in the table below.



The total phosphorus criteria is exceeded if the lower confidence limit of the sample median exceeds the state total phosphorus criteria of 0.075 mg/L.



The site is classified as Watch Waters if the median total phosphorus concentration falls within the confidence limit.



The total phosphorus criteria is met if the upper limit of the confidence interval does not exceed the state total phosphorus criteria of 0.075 mg/L.



### PROJECT PARTNERS



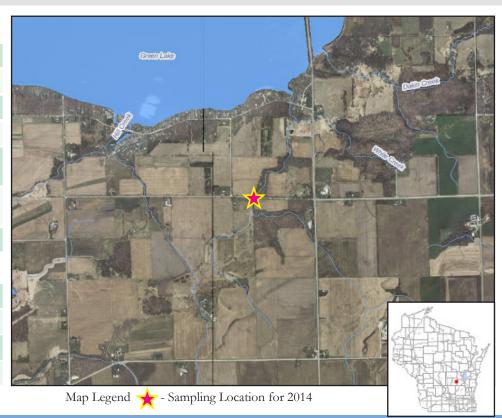


### **Monitoring Site Information**

Monitoring one information				
SWIMS Station ID	10042146			
County	Green Lake			
Watershed	Big Green Lake			
Watershed Area	107.31 sq miles			
Total Stream Miles in Watershed	141.08 miles			
Downstream Waterbody	White Creek			
Volunteer(s)	Dave Kummerow			

### 2014 Monitoring Results

$\mathcal{O}$	
Min TP Value	$0.0589~\mathrm{mg/L}$
Max TP Value	$0.163~\mathrm{mg/L}$
Median TP Value	$0.08055~\mathrm{mg/L}$
No. Samples $> 0.075 \text{ mg/L}$	4



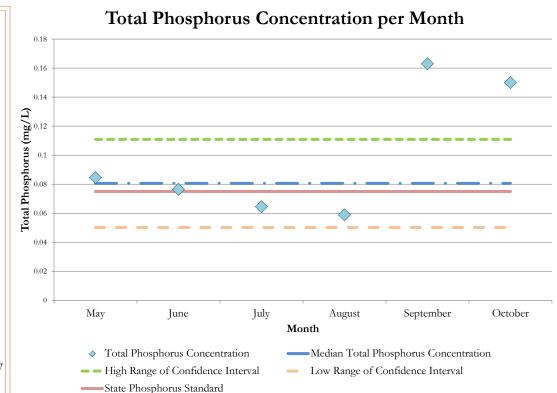


### Volunteer Total Phosphorus Monitoring

### Unnamed Trib to White Creek US Scott Hill Road

2014 Monitoring Results

This Targeted Watershed Assessment (TWA) will address needs for baseline water quality monitoring of the watershed on the south side of Big Green Lake. This watershed discharges anywhere between 15-30% of the total phosphorus load into Big Green Lake. Considerable stream restoration and watershed BMP work has been done and is projected for the future by multiple agencies, including Green Lake County LWCD and NRCS.



### Why Phosphorus?

Phosphorus is an essential nutrient responsible for plant growth, but it is also the most visible, widespread water pollutant in Wisconsin lakes. Small increases in phosphorus levels in a lake can bring about substantial increases in aquatic plant and algae growth, which in turn can reduce the recreational use and aquatic biodiversity of said lake. When the excess plants die and are decomposed, oxygen levels in the water drop dramatically which can lead to fish kills.

Additionally, one of the most common impairments in Wisconsin's streams is excess sediments that cover stream bottoms. Since phosphorus moves attached to sediments, it is intimately connected with this source of pollution in our streams. Phosphorus originates naturally from rocks, but its major sources in streams and lakes today are usually associated with human activities: soil erosion, human and animal wastes, septic systems, and runoff from farmland or lawns. Phosphorus-containing contaminants from urban streets and parking lots such as food waste, detergents, and paper products are also potential sources of phosphorus pollution from the surrounding landscape. The impact that phosphorus can have in streams is less apparent than in lakes due to the overall movement of water, but in areas with slow velocity, where sediment can settle and deposit along the bottom substrate, algae blooms can result.







Photo credits to Matt Berg, David Seligman, Linda Warren, and Adrian Konell

### **Volunteer Monitoring Protocol**

To assess in stream phosphorus levels, WAV volunteers collected water samples that were analyzed for total phosphorus (TP) at the State Lab of Hygiene during the growing season (May through October). Following Wisconsin Department of Natural Resources (WDNR) methods, six phosphorus water samples were collected at each monitoring site - one per month for each of the six months during the growing season, The water samples were collected approximately 30 days apart and no samples were collected within 15 days of one another.

Total phosphorus impairment is assessed using the criteria in the table below.

### Total Phosphorus Exceedence Criteria



The total phosphorus criteria is exceeded if the lower confidence limit of the sample median exceeds the state total phosphorus criteria of 0.075 mg/L.



The site is classified as Watch Waters if the median total phosphorus concentration falls within the confidence limit.



The total phosphorus criteria is met if the upper limit of the confidence interval does not exceed the state total phosphorus criteria of 0.075 mg/L.



PROJECT PARTNERS





# An Assessment of Hill, Roy, Silver, and Wuerches Creeks (303d Impaired Waters)



### **Wisconsin Department of Natural Resources**

 $Ted\ Johnson-Water\ Resources\ Biologist$ 

**Adam Nickel – Water Resources Specialist** 

Eric Evensen – Water Resources Specialist

#### **Background**

The 2008 draft submittal of 303d impaired waters identifies eight waters bodies as being impaired in Green Lake County. Of these eight water bodies four of them including Hill, Roy, Silver, and Wuerches Creeks empty into Big Green Lake. Big Green Lake is also listed as impaired water for PCB pollutant and contaminated fish tissue impairment. Hill and Roy Creek are listed for sediment pollutant and degraded habitat impairment. Silver Creek is listed for sediment pollutant and degraded habitat along with elevated water temperature impairment. Wuerches Creek is listed for sediment and total phosphorus pollutant, degraded habitat, elevated water temperature, and low dissolved oxygen impairment. Implementing plans to improve these water bodies so they meet water quality standards is a priority. Through improving the water quality of these creeks the condition of Big Green Lake will also benefit.

In 1980 Big Green Lake was selected as a priority watershed under the Wisconsin Nonpoint Source Water Pollution Abatement Program. The primary objective of this program is to provide cost-sharing and technical assistance to local agencies for the control of nonpoint source pollution. Following the designation as a priority watershed, a priority watershed plan was written. There were four main goals outlined in the Big Green Lake Priority Watershed Project:

- 1. Reduce the concentrations of bacteria to "acceptable" levels.
- 2. Reduce the nutrient loading levels to streams from nonpoint sources by 40% on a yearly basis.
- 3. Increase the average transparency (secchi disk) readings within Big Green Lake during the open water times.
- 4. Halt the trend of increasing littoral zone establishment as a result of sediment loading to the lake.

The goal most associated with this report is reducing nutrient input into the lake from nonpoint sources by 40%. To reach this goal the installation of best management practices began on agricultural land within the watershed. While many of the projects were successful in limiting the nutrient levels entering Green Lake, much more needs to be done to prevent further runoff and erosion from taking place because these water bodies are still not meeting water quality standards.

To evaluate these 303d water bodies the following sampling protocol was carried out for each of the creeks:

- 1. Fish surveys for community assessment (IBI)
- 2. Macroinvertebrate collection
- 3. Qualitative habitat assessment
- 4. Continuous temperature monitoring
- 5. Phosphorus levels

All sampling was carried out according to Wisconsin DNR protocols.

#### **Results**

#### Hill Creek

Hill Creek measures 2 miles in length. It is an outlet of Little Twin Lake that flows in a northeasterly direction into Green Lake. The stream contains clear, hard water often polluted with agricultural runoff.

#### Fish Survey

Three fish assessments were conducted on Hill Creek during the summer months of 2011. Between the assessments largemouth bass, white sucker, yellow perch, black bullhead and bluegill were found most often. The Coldwater Index of Biotic Integrity (IBI) was calculated for all fish assessments. The assessment conducted on July 20, 2011 generated an IBI of 0 (very poor) while the assessment conducted on July 26, 2011 generated an IBI of 20 (poor). The assessment conducted on August 30, 2011 had an IBI of 10 (poor). These values would indicate that the stream section sampled has likely experienced high levels of environmental degradation.

#### Macroinvertebrate Sampling

Macroinvertebrate sampling from 2008 indicate Hill Creek has a Hilsenhoff Biotic Index (HBI) of 7.7 and a Shannon Diversity Index of 1.18. A HBI of 7.7 indicates poor water quality with very significant organic pollution. The Shannon Diversity Index indicates low diversity of macroinvertebrates. Both indices suggest that the sampled reach has poor water quality most likely due to agricultural practices.

#### Qualitative Habitat

The qualitative habitat rating total score was 58 (Fair-Good). In the sampling reach riparian buffer zones and bank erosion were rated as excellent while the width to depth ratio and habitat diversity had a rating of fair.

#### *Temperature Monitoring*

Water temperature was monitored during 2011 on Hill Creek. Results show that the water temperature rarely got above 70°F with a peak of around 70.5°F on July 20<sup>th</sup>. Water temps in Hill Creek are low enough year round to support a trout population.

#### *Water Quality*

Total phosphorus levels were measured during the summer of 2011. Levels ranged from .06 - .18 mg/L of phosphorus in Hill Creek. Over nine different sampling events Hill Creek had an average phosphorus level of 0.104 mg/L. Wisconsin's water quality standard for small streams is 0.075 mg/L.

Flow rates can be used in conjunction with phosphorus levels to calculate the amount of phosphorus entering a water body over a given time frame. During a non rain event it has been calculated that approximately .8 lbs of phosphorus enter Green Lake every day from Hill Creek. After a rain event that number jumps up to approximately 2.4 lbs of phosphorus entering Green Lake every day. When 2.4 lbs of phosphorus enter Green Lake it has the potential of producing 730 – 1200 lbs of algae.

Total suspended solid (TSS) levels were also measured during this study. TSS levels ranged from 352 - 476 mg/L.

#### Roy Creek

Roy Creek is a tributary to Green Lake containing clear, hard water. The creek measures 8 miles in length from its headwaters till it flows into Green Lake. The stream bank consists mostly of upland hardwoods, farm pasture, and cultivated crops.

#### Fish Survey

A fish assessment was conducted in the summer of 2011 on Roy Creek. Four species were collected which included white sucker, central mudminnow, mottled sculpin and fathead minnow. The coldwater IBI for the creek section was 10, which is categorized as poor. A coldwater IBI rating of poor is usually indicative of major environmental degradation and is a sign that the biotic integrity has been severely reduced.

#### Macroinvertebrate Sampling

Macroinvertebrate sampling from 2009 indicate Roy Creek has a HBI of 4.80 and a Shannon Diversity Index of 1.63. A HBI of 4.8 indicates good water quality while still displaying some organic pollution. The Shannon Diversity Index indicates low diversity of macroinvertebrates.

#### Qualitative Habitat

Roy Creek received a total qualitative fish habitat score of 52. Bank erosion is extensive along the sampled reach and was rated poor. Roy Creek also rated as fair for fine sediments and available cover for fish. The only category where it was rated as excellent was riffle:riffle or bend:bend ratio.

#### *Temperature Monitoring*

Water temperature was monitored on Hill Creek during the summer of 2011. Due to technical difficulties with the temperature probe measurements were only gathered for the months of May and June. The average water temp for this time period was approximately 57.5°F with a peak temp of 69.82 on June 8<sup>th</sup>.

#### Water Quality

Total phosphorus levels were measured during the summer of 2011. Levels ranged from .13 - .44 mg/L of phosphorus in Roy Creek. Over eight different sampling events Hill Creek had an average phosphorus level of 0.104 mg/L. Wisconsin's water quality standard for small streams is 0.075 mg/L.

During a non rain event it was calculated that approximately 1.5-2.5 lbs of phosphorus enter Green Lake every day from Roy Creek. Since Roy Creek is a high gradient stream water can be carried to the lake much faster than other streams in the watershed. We were unable to get accurate measurements for how much phosphorus enters Green Lake after a rain event because of how fast the rain water can pass through the system. One can safely assume that the amount of phosphorus entering Green Lake would be 2-5 times greater during a rain event producing large amounts of algae.

Total suspended solid levels were also measured on Roy Creek. TSS levels in the creek ranged from 456 - 752 mg/L.

#### Silver Creek

Silver Creek begins in northwestern Fond du Lac County where it flows west to Green Lake. Silver Creek drains the largest of Green Lake's sub-watersheds. Much of the watershed is agriculture which is the main reason Silver Creek carries a significant phosphorus load to Green Lake. The watershed also includes the city of Ripon.

#### Fish Survey

A fish assessment was conducted in the summer of 2011 on Silver Creek. White sucker, central mudminnow and largemouth bass were the most common species found. The coldwater IBI was calculated to be 30 (fair). Silver creek received the highest IBI rating of the water bodies sampled. Intolerant and native stenothermal coldwater species were uncommon while tolerant eurythermal species or warmwater species were more abundant.

#### Macroinvertebrate Sampling

Macroinvertebrate sampling conducted in 2007 discovered HBI of 5.5 and a Shannon Diversity Index of 3.00. A HBI of 5.5 indicates good water quality while still displaying some organic pollution. The Shannon Diversity Index indicates relatively high diversity in the macroinvertebrate community.

#### Qualitative Habitat

The total qualitative habitat rating was 62 (Fair-Good) on Silver Creek. Riparian buffers were the only category rated as excellent, while habitat diversity was low and fine sediments were extensive. Gravel substrate was very sparse at this station limiting the chances of trout spawning success.

#### *Temperature Monitoring*

Temperature monitoring was conducted on Silver Creek. The highest temperatures were recorded during the summer months of July and August with a peaking temperature of 86°F. The average temperature from June through August was 71.3°F.

#### Water Quality

Total phosphorus was measured during the summer of 2011 on Silver Creek. Levels ranged from .099 - .56 mg/L of phosphorus. When using flow measurements and phosphorus levels it was calculated that during a non rain event approximately 13 lbs of phosphorus enter Green Lake every day from Silver Creek. After a rain event that number jumps up to 72 lbs of phosphorus entering Green Lake every day. 72 lbs of phosphorus has the ability to produce 21,000 - 37,000 lbs of algae.

Total suspended solids were also measured in Silver Creek. TSS levels in the creek ranged from 466 - 700 mg/L.

#### **Wuerches Creek**

Wuerches creek is an 8 mile continually flowing stream that enters the Green Lake County Park marsh area. A small number of cattle are pastured in the immediate stream area and row cropping occurs near the stream course in some areas.

#### Fish Survey

A fish assessment was conducted on Wuerches Creek during the summer of 2011. Only two species of fish were found, mottled sculpin and central mudminnow. The coldwater IBI for the Wuerches Creek reach was calculated to be 20 (poor), indicating major environmental degradation.

#### Macroinvertebrate Sampling

Macroinvertebrate sampling from 2007 provided an HBI of 5.79 and a Shannon Diversity Index of 3.45. The HBI indicates fair water quality and fairly significant organic pollution. The Shannon Diversity Index indicates relatively high diversity of the macroinvertebrate community.

#### Qualitative Habitat

Wuerches Creek had a qualitative habitat rating total score of 72. The sampling station had no significant bank erosion and habitat diversity was high. The creek is moderately deep and narrow with a width/depth ratio scoring in the fair category.

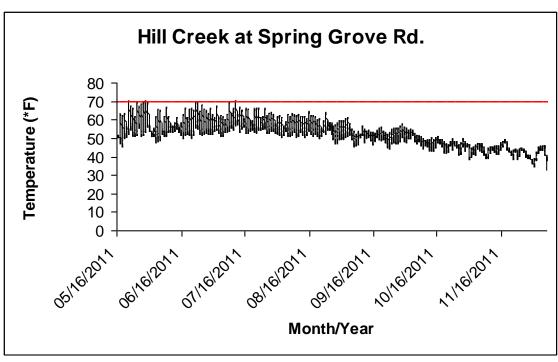
#### *Temperature Monitoring*

Temperature monitoring was conducted on Wuerches Creek in 2011. The highest temperatures were recorded in June and July with a peaking temperature of 73°F.

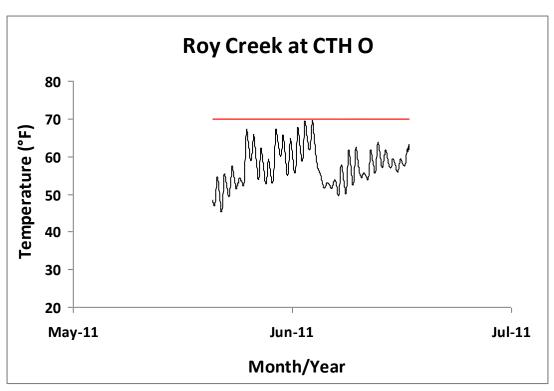
#### Water Quality

Total phosphorus was measured during the summer of 2011 on Wuerches. Levels ranged from .07 - .7 mg/L of phosphorus. During a non rain event approximately 1.1 lbs of phosphorus was calculated to enter Green Lake every day from Wuerches Creek. This creek is a higher gradient system. Water from rain events is able to flow through the system and into Green Lake very quickly. After a rain event we calculated the amount of phosphorus entering the lake to be about 1.5 lbs per day. Due to the high gradient of the creek we believe that we missed the majority of the phosphorus entering the lake after the rain event. This is assumed because the flow of the creek was back down to base flow after receiving 2.95 inches of rain in the previous 48 hours.

Total suspended solid levels were measured during the summer of 2011 as well. Levels ranged from 476 - 524 mg/L on Wuerches Creek.



**Figure 1:** Temperature monitoring on Hill Creek at Spring Grove Road crossing from May to December of 2011.



**Figure 2:** Temperature monitoring on Roy Creek at County Highway O crossing from May to June of 2011.

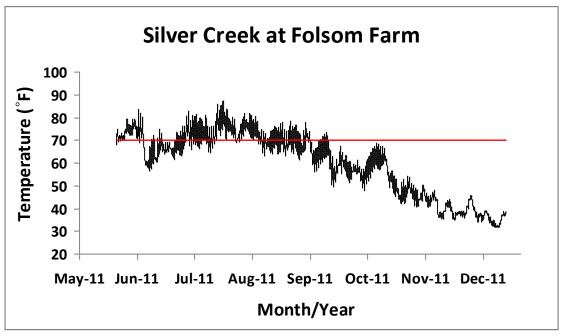
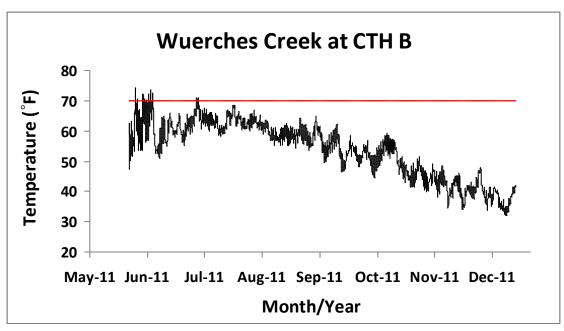


Figure 3: Temperature monitoring on Silver Creek at Folsom Farm from May to December of 2011.



**Figure 4:** Temperature monitoring on Wuerches Creek at County Highway B crossing from May to December of 2011.

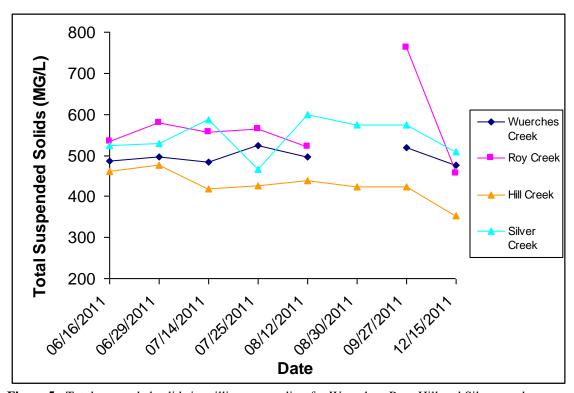


Figure 5: Total suspended solids in milligrams per liter for Wuerches, Roy, Hill and Silver creeks.

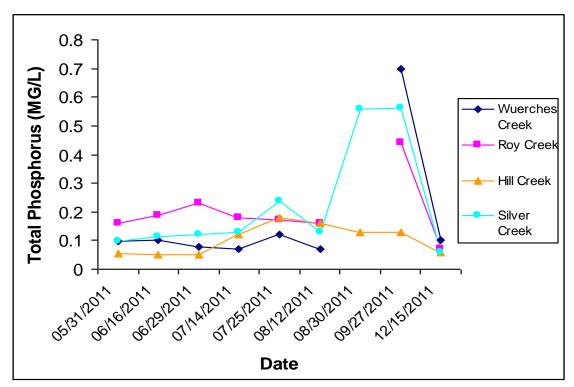


Figure 6: Total phosphorus in milligrams per liter for Wuerches, Roy, Hill and Silver creeks.

Wuerches Creek	05/31/2011	06/16/2011	06/29/2011	07/14/2011	07/25/2011
Temperature ©					
(Field)	20.6		17.6	16.7	17
Conductivity					
(UMHOS/CM)	723	728	758	752	762
Dissolved Oxygen					
(MG/L)	14.3		10	7.7	5.8
DO Percent					
Saturation	159.5		106	80	60
PH	8.1	8.19	7.8	7.7	7.7
Transparency Tube					
(cm)	>120	>120	>120	>120	>120
Alkalinity (MG\L)		301		_	
Flow (m <sup>3</sup> /sec)			0.07487	0.08235	0.06427

**Table 1:** Water quality data collected from Wuerches Creek upstream from the County highway B road crossing.

Roy Creek	05/31/2011	06\16\2011	06/29/2011	07/14/2011	07/25/2011
Temperature ©					
(Field)	17.8		14.5	16	16.9
Conductivity					
(UMHOS/CM)	719	723	736	736	739
Dissolved Oxygen					
(MG/L)	7.4		8.9	8.7	8.5
DO Percent					
Saturation	81.2		87.9	89.1	87.8
PH	8.2	8.27	8.2	8.3	8.3
Transparency Tube					
(cm)	30	100	10	13	15
Alkalinity (MG\L)		285			
Flow (m <sup>3</sup> /sec)			0.065799	0.04182	0.038086

**Table 2:** Water quality data collected from Roy Creek upstream from the County highway O road crossing.

Hill Creek	5/31/11	6/16/11	6/29/11	7/14/11	7/25/11
Temperature ©					
(Field)	15		13.1	22.5	23.1
Conductivity					
(UMHOS/CM)	705	713	717	609	611
Dissolved Oxygen					
(MG/L)	9.8		11.5	9.2	6.8
DO Percent					
Saturation	98.2		109.1	107.1	78.7
PH	8.3	8.39	8.4	8.3	7.9
Transparency Tube					
(cm)	90	90	100	80	40
Alkalinity (MG\L)		255			
Flow (m <sup>3</sup> /sec)				0.03648	0.07093

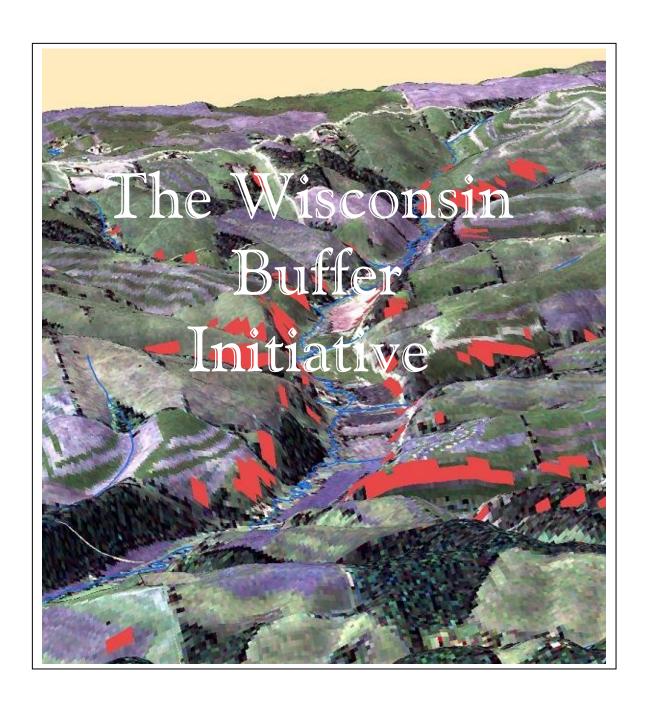
Table 3: Water quality data collected from Hill Creek upstream from the Spring Grove road crossing.

Silver Creek	05/31/2011	06\16\2011	06/29/2011	07/14/2011	07/25/2011
Temperature ©					
(Field)	20.7		22.4	20.8	23.8
Conductivity					
(UMHOS/CM)	830	812	825	910	657
Dissolved Oxygen					
(MG/L)	8.1		8.9	8.4	4.1
DO Percent					
Saturation	95		104.3	105.5	48.4
PH	8	8.35	8.2	8.3	7.9
Transparency Tube					
(cm)	60	60	60	50	35
Alkalinity (MG\L)		295			
Flow (m³/sec)	11 . 16 . 63	G 1.6	0.54851	0.54557	1.591146

 Table 4: Water quality data collected from Silver Creek from Folsom Farm and Zeratsky Farm.

Fish Survey Information	Creek Name						
Species Name	Hill (7/20/11)	Hill (7/26/11)	Hill (8/30/11)	Roy	Silver	Wuerches	
Bluegill (Lepomis macrochirus)			4		9		
Black Bullhead (Ameiurus melas)			4				
White Sucker (Catostomus commersoni)	4		2	12	45		
Yellow Perch (Perca Flavescens)			5		2		
Largemouth Bass (Micropterus salmoides)		3	2		25		
Northern Pike (Esox Lucius)			2		5		
Creek Chub (Semotilus atromaculatus)			1		7		
Brown Trout (Salmo trutta)			2				
Green Sunfish (Lepomis cyanellus)		1	2		1		
Mottled Sculpin (Cottus bairdi)				11		30	
Central Mudminnow (Umbra limi)				7	16	10	
Fathead Minnow (Pimephales promelas)				6			
Smallmouth Bass (Micropterus dolomieu)					2		
Muskellunge (Esox masquinongy)					4		
Yellow Bullhead (Ameiurus natalis)					2		
Log Perch (Percina caprodes)					5		
Black Crappie (Pomoxis nigromaculatus)					2		
Johnny Darter (Ethostoma nigrum)					2		
Hognose Sucker (Hypentelium nigricans)		-			1		
Golden Shinner (Notemigonus crysoleucas)					2		

**Table 5:** Fish collected during backpack electrofishing surveys conducted on Hill, Roy, Silver and Wuerches Creeks.



A Report to the Natural Resources Board of the Wisconsin Department of Natural Resources by the University of Wisconsin-Madison, College of Agricultural and Life Sciences.

December 22, 2005





December 29, 2005

To: Secretary P. Scott Hassett, Wisconsin Department of Natural Resources

Members of the Natural Resources Board

From: David Hogg, Interim Dean

Re: Report of the Wisconsin Buffer Initiative

In the spring of 2002 you requested that faculty in the UW-Madison College of Agricultural and Life Sciences (CALS) review the research literature relating to riparian buffer effectiveness as part of your deliberations on the revision of NR 151. The University responded by convening a group of university, agency, environmental and agricultural and natural resources representatives that reviewed the scientific literature and recommended an adaptive management approach to more precisely design and locate buffers in Wisconsin. On the basis of that interim report, the NR Board passed a resolution directing CALS to conduct the necessary research that would help determine effectiveness of buffers under various Wisconsin conditions with a final report delivered to you on or before December 31, 2005.

Over the last 30 months scientists in CALS have collaborated with colleagues in the UW-Madison Center for Limnology to address your charge. The following report reflects the research and analysis that was conducted to examine the potential role of riparian buffers as part of Wisconsin NR 151.

A critical dimension of the work conducted over the last 30 months was the extensive collaboration with agricultural, environmental, natural resource professional associations, and our state and federal agencies. Our faculty worked closely with all these diverse groups to help specify what research needed to be conducted, how this research should be carried out, and the implications of the data that emerged from this research. There is a general agreement by all these diverse groups and interests that the recommendations contained in this report are the best way to use buffers to address the water quality concerns in Wisconsin.

Finally I believe that this report reflects the willingness of UW-Madison to work with state agencies in a cooperative partnership that serves to advance our joint mission of better serving the citizens of Wisconsin.

### **Table of Contents**

1. GENESIS OF THE WISCONSIN BUFFER INITIATIVE (WBI)
2. ESTABLISHING THE COURSE FOR THE WBI
3. RESEARCH QUESTIONS ON CONSERVATION SYSTEMS AND RIPARIAN BUFFERS IN WISCONSIN
4. ADAPTIVE MANAGEMENT AS A BASIS FOR NATURAL RESOURCE MANAGEMENT PROGRAMS
5: TARGETING BUFFER IMPLEMENTATION ON A STATEWIDE SCALE 13
6. PLANNING AND IMPLEMENTATION TOOLS THAT CAN BE USED AT THE LOCAL LEVEL
7. PLACEMENT AND DESIGN OF CONSERVATION SYSTEMS AND RIPARIAN BUFFERS
8. PILOT STUDY OF THE PROPOSED WBI IMPLEMENTATION PROCESSES
9. ASSESSING THE ECONOMIC IMPACTS OF ALTERNATIVE MANAGEMENT PRACTICES ON SELECTED WISCONSIN FARMS 4:
10. RECOMMENDATIONS OF THE WBI ADVISORY COMMITTEE 5
APPENDIX A: RECOMMENDATIONS FOR MONITORING 55
APPENDIX B: WBI LIST OF RANKED WATERSHEDS 5
APPENDIX C: ECONOMICS OF MANAGEMENT RECOMMENDATIONS 9

### **Acknowledgements**

As Chair of the Wisconsin Buffer Initiative I write this acknowledge page with a sense of both humility and optimism. Humility because literally hundreds of Wisconsin citizens contributed to this report and need to be acknowledged. The initial goals, and then the progress toward those goals, were brought to the citizens of Wisconsin in the town halls, schools and other meeting places across Wisconsin. The insights, comments and suggestions of these citizens were then brought back to the Wisconsin Buffer Initiative Advisory Committee. Here citizen and scientist worked side-by-side as equals in trying to determine how to address a complex and contentious issue. In the end, the title of Advisory Committee proved to be a misnomer. This group of citizens, scientists and agency staff provided the leadership and courage to chart a new course for natural resource management in Wisconsin. Representatives from the environmental and agricultural communities of Wisconsin, while acknowledging their differences, also found a higher, common ground. Staff from federal and state agencies, as well as representatives from professional conservation associations, were receptive and supportive of seeking innovative approaches consistent with a challenging fiscal climate. Scientists and staff from the University of Wisconsin-Madison proved through their actions that the core philosophy of the "Wisconsin Idea" is still a vibrant and an invigorating a concept as it was when originally conceived over a hundred years ago. To each and every member of this group I offer my sincere thanks for your commitment to proving the ideal that by working together we can find solutions that were beyond the grasp on any one of us. You have given me a sense of optimism relative to the future of resource management in Wisconsin.

A special acknowledgement is also due to Senator Herb Kohl who helped secure federal funds that allowed this citizen-driven experiment in natural resource management policy to flourish. His faith in creating an opportunity for Wisconsin's Land Grant University to work with the citizens of Wisconsin in charting their own future is especially appreciated.

An effort such as the Wisconsin Buffer Initiative does not occur without the explicit support from the leaders of the University of Wisconsin, agencies and organizations who have been represented in this effort. You have my thanks for allowing your staff to be bold and creative in their thinking and actions. I believe this report demonstrates that your support was justified.

Sincerely,

Telen Nowak

Pete Nowak, Chair, Wisconsin Buffer Initiative, December 21, 2005

### **Executive Summary**

The Wisconsin reformers have accomplished the extraordinary results for which the whole nation owes them so much...as soon as they decided that a certain object was desirable they at once set to work practically to study how to develop the constructive machinery through which it could be achieved....That state has become literally a laboratory for wise experimental legislation aiming to secure the social and political betterment of the people as a whole.—Teddy Roosevelt 1912

Wisconsin's water resources are among its most critical economic, ecological, and cultural assets. Although human activities have degraded many of our freshwater resources, progress has been made in addressing many sources of degradation. Agricultural nonpoint-source pollution, however, remains a persistent problem, in part because of its diffuse nature and in part because of the high economic value of agriculture in Wisconsin. Recently, the Wisconsin Natural Resources (NR) Board identified a need for new and innovative strategies for reducing agricultural pollution. The Wisconsin Buffer Initiative (WBI) was formed in response to this need. The goal of the WBI was designing a buffer implementation program to achieve water quality improvements for Wisconsin in the most cost-effective and efficient manner.

The WBI approach differs dramatically from other natural resource management approaches in many ways. The hallmark of WBI is the use of a science-based approach to carefully target conservation efforts, including buffer locations and configurations, so as to maximize water quality improvements. This approach is both effective and efficient. In addition, the WBI advocates for the use of watershed-level adaptive management. Watersheds are the units of program implementation, and water quality is monitored in response to implementation. On-the-ground projects are viewed as opportunities for learning, and are intended to greatly accelerate improvements in policy.

This report is organized into ten chapters. Chapters 1–3 discuss the **genesis and direction of the WBI**. Revisions to the state administrative rules on nonpoint-source pollution (NR151) resulted in an impasse over the issue of mandatory implementation of riparian buffers. In May 2002, the Wisconsin NR Board directed the Wisconsin Department of Natural Resources to collaborate with the University of Wisconsin College of Agricultural and Life Sciences on the development of a scientifically-based agricultural buffer standard. The WBI was formed to facilitate the enhancement and synthesis of the best available research on riparian buffers. A civic science process involving a wide range of stakeholders was driven by addressing the following questions:

- 1. Where are buffers needed most across the diverse landscapes and land use in Wisconsin?
- 2. What types of buffers are needed in these specific locations?
- 3. What are the consequences when buffers are installed in these locations?
- 4. What will be needed to get these buffers into these specific locations?

The remainder of the report describes how the WBI addressed each of these questions.

Chapter 4 describes the rationale for using an **adaptive management approach** to reducing nonpoint-source pollution. A major goal of the WBI was to *design an efficient and cost-effective statewide program for achieving water quality improvements*. Consequently, the WBI advocates an adaptive management approach whereby policy is

designed to learn from implementation experience and is modified on the basis of new knowledge. Management is based on the best available science, but is structured so as to learn from uncertainty and surprises.

Chapter 5 presents the **statewide prioritization of watersheds**. Small watersheds are most likely to provide measurable improvements in stream water quality. Scientists at the University of Wisconsin-Madison ranked 1598 watersheds throughout the state of Wisconsin on their potential to meet three management goals:

- Improve stream water quality
- Protect and enhance aquatic biological communities
- Sustain lake water quality

These management goals were developed by the WBI Advisory Committee and used to guide the statewide analysis. Data from state and federal agencies were used to build statistical models to identify watersheds that are most likely to respond to reductions of phosphorus and sediment through the implementation of conservation practices. The analyses were then used to compile a ranked list of watersheds to be targeted for more intensive conservation efforts. A **poster** (enclosed) displays a map of the ranked watersheds, and also summarizes the WBI targeting approach.

Chapter 6 describes procedures to aid local conservation staff in **planning and implementing conservation systems** in selected WBI watersheds. Local knowledge of watershed conditions can verify that selected watersheds have the capacity to respond effectively to conservation practices. A simple computer model of soil loss can be used to identify subwatershed areas that are most likely to be contributing to water quality impairment. On individual farms, SNAP-Plus software (developed by UW-Madison scientists) can be used to conduct field-level management need analysis. This software can also simulate changes in soil and nutrient losses under different management scenarios, such as changes in crop rotations, tillage, manure application, and buffer implementation.

If buffers are found to be an effective and favorable management practice for an individual field, the **placement and design of riparian buffers** (Chapter 7) should be determined by a contributing area analysis. In this approach, sections of buffer that receive runoff from larger drainage areas are wider than sections with small drainage areas. Allowing for contributing area more effectively removes soil and nutrients from runoff than constant width buffers, and it reduces the amount of land taken out of production.

A **pilot study** (Chapter 8) was conducted in two Wisconsin watersheds to examine the feasibility of the targeting and implementation process proposed by the WBI and to receive feedback from local conservation staff. Farm management and soil information was fed into SNAP-Plus to evaluate the need for management changes and to compare the projected **economic consequences** (Chapter 9) of various management options. Most of the fields had predicted soil and phosphorus losses below current Natural Resources Conservation Service (NRCS) 590 standards. Of those fields that exceeded thresholds, feasible management changes, such as tillage changes, would correct problems. In many instances, these changes would result in increased farm profitability.

The work described in this report is the foundation for the WBI final **recommendations** (Chapter 10) for a statewide program for riparian buffers. Our

program carefully targets conservation efforts, thus maximizing efficiency, while simultaneously providing the environmental improvements that the public demands. The WBI feels that adoption of this approach for nonpoint source pollution will elevate Wisconsin to its former position as a national leader in innovative policy and natural resource management. Perhaps more importantly, the civic science approach of the WBI exemplifies the Wisconsin Idea. Our hope is that these efforts will serve as the foundation for far-sighted and innovative policy that will contribute to the betterment of the citizens and the natural resources of the State of Wisconsin.

### 1. Genesis of the Wisconsin Buffer Initiative (WBI)

Wisconsin has a rich and diverse array of water resources; over 15,000 inland lakes are interconnected with 32,000 miles of perennial streams and rivers, which are complemented by another 23,000 miles of intermittent streams. These waters drain into the Mississippi River, Lake Superior, or Lake Michigan while recharging over two quadrillion gallons of groundwater (Wisconsin Academy of Sciences, Arts and Letters 2003). Wisconsin also has a rich and diverse agricultural economy, which is represented by 76,500 farms operating across 15.5 million acres that generate almost \$7 billion dollars in receipts (USDA National Agricultural Statistical Service 2005). Besides being the number one producer of cheese in the United States, our agricultural diversity is represented by the fact that Wisconsin is among the top five producers of oats, potatoes, cranberries, tart cherries, carrots, snap beans, sweet corn, and, as would be expected with the dairy industry, corn for silage. Given Wisconsin's diverse agricultural systems and their importance to the state's economy, it is inevitable that farming in some situations and at some times would impair our rich water resources.

Protecting our water resources is a major priority for all the citizens of Wisconsin—urban, suburban, rural, and farm. The Wisconsin Buffer Initiative (WBI) believes that Wisconsin can have both a viable agricultural system and quality water resources. This belief also served as a guiding principle during the genesis of the WBI. The WBI began as a scientific review of riparian buffers in support of potential rule changes for Wisconsin's regulation of nonpoint source pollution. Since its inception, the effort has evolved into a demonstration of how diverse interests can work together to chart a course for protecting the state's waters. Understanding the history that led to the formation of the WBI is necessary to appreciate the recommendations that have emerged from this diverse coalition.

### Wisconsin's Nonpoint-Source Pollution Programs

In 1977, the Wisconsin Department of Natural Resources (DNR) created a grant program, then known as NR 418 or the DNR Priority Watershed and Priority Lake Program, to address the issue of nonpoint-source water pollution. From the inception of this program through June 2004, over \$187 million was spent on local assistance and cost share grants to protect the state's waters from nonpoint source pollution. On the basis of area-wide water quality plans developed under the requirements of the Federal Water Pollution Control Act, the DNR identified watersheds and lakes in which the need for nonpoint pollution control was most critical. High- or medium-priority watersheds became eligible for funding based on an analysis of DNR district workload, county ability to manage a project, and projected landowner participation. The Wisconsin Legislature in 1997 directed the DNR to re-rank all watersheds in the state based on the level of impairment. Current priority watershed projects (n = 62) that were active were to be terminated under Wisconsin Act 27 unless they were designated a priority by the

<sup>&</sup>lt;sup>1</sup> The term "buffer" in this report refers to a riparian (along a stream or river) buffer. The WBI acknowledges that the US Department of Agriculture Natural Resources Conservation Service defines at least nine different types of buffers appropriate to Wisconsin. The term "buffer program" in this report refers to an organized effort to implement agricultural conservation practices, including the installation of buffers, for the purpose of improving water quality.

Wisconsin Land and Water Conservation Board. The Land and Water Board subsequently re-designated all 62 projects as priorities. Then, in the 1999–2001 biannual state budget, the Wisconsin Legislature made a number of major modifications to the nonpoint-source pollution-abatement program. The Legislature also revamped the various administrative rules that govern nonpoint source pollution in both rural and urban settings. These revisions included NR 151 through NR 155 and ACTP 50. These changes in the administrative code and administration of the nonpoint program are often referred to as the Redesign of the Nonpoint Pollution Program.

#### The Role of Riparian Buffers in Nonpoint-Source Pollution Control

As noted, provisions by the Wisconsin Legislature, Act 27 in 1997 and Act 9 in 1999, directed the DNR to develop performance standards to control polluted runoff from non–agricultural activities, develop performance standards and prohibitions for agricultural activities in cooperation with the Department of Agriculture, Trade and Consumer Protection (DATCP) including four manure management prohibitions developed through a previous advisory committee effort, and to make other changes to address polluted runoff problems from rural and urban sources.

In response to these Legislative directives, a series of technical recommendations and suggested administrative code language were developed. An agency-appointed advisory committee then solicited and generated feedback and public comments as part of the process of developing specific implementation recommendations. Although a broad spectrum of citizens and vested interests participated in this process, consensus was not reached on all issues. In particular, one remaining controversial aspect was associated with the issue of mandatory riparian buffers.

Specifically, the work group developing rules for the implementation of NR 151 recommended a ten-foot vegetated buffer with 50% crop residue on the next ninety feet of cropland adjacent to the riparian area, a twenty-foot vegetated buffer in the riparian zone with 30% residue on the next thirty feet on adjacent cropland, or a thirty-five-foot vegetated buffer in the agricultural riparian areas of the state. Mandating this combination of riparian buffers and conservation tillage in water-quality management areas (i.e., those areas proximate to a river, stream, or lake) proved to be a contentious recommendation. Environmental interests viewed this recommendation as necessary to protect the state's waters, while agricultural interests viewed it as imposing a hardship on the state's agricultural producers.

What emerged in this debate over mandated riparian buffers was the question of scientific justification for this position: both sides, pro and con, appeared to use scientific justifications to support their position. In response to this impasse, DNR Secretary Darrell Bazzell sent a request to Elton Aberle, Dean of the University of Wisconsin-Madison College of Agricultural and Life Sciences (CALS) to review the science on the functioning of riparian buffers.

In response to this request, an ad hoc committee was formed of UW-Madison scientists with expertise in this area. In addition to the scientists, every effort was made to include all of the vested interests that had been involved in the larger discussions with the Wisconsin Legislature and the DNR. Operating under a sixty-day deadline, this group issued a report on April 26, 2002, (*Filter Strips and Buffers on Wisconsin's Private Lands: An Opportunity for Adaptive Management*) and an 890-item scientific

bibliography on this topic (Correll 2003). This report was brought before the NR Board in May 2002.

There were several key elements in the ad hoc committee report that had a major influence on the subsequent activities of the WBI. In particular, these elements included:

- Implementing NR 151 relative to riparian buffers should be based on an adaptive management approach. (Chapter 4 of this report discusses this in depth.)
- Research reviewed in this process clearly specified that riparian buffers can have many potential benefits, but the nature and extent to which any of these benefits are achieved is very site specific. Thus, the issue is not whether buffers are good, but where across Wisconsin's diverse landscape can riparian buffers achieve the most benefit for water quality in accord with the intent of NR 151.
- The ad hoc report also noted that riparian buffers by themselves would be unlikely to induce significant changes in the quality of the state's waters. Riparian buffers need to be part of a larger conservation system.
- Due to the potential to take land out of production as part of the process of installing conservation systems and riparian buffers, the ad hoc report emphasized the importance of the private landowner on whose land riparian buffers are needed being an integral part of the overall process.

The NR Board accepted this ad hoc report and asked the UW-Madison CALS to carry out the necessary research and activities to address the recommendations contained in the April 2002 report. The UW-Madison CALS was to submit a final report containing their research and activities to the NR Board on or before December 31, 2005.

#### The Charge of the Wisconsin Buffer Initiative

The WBI was officially developed in response to this charge from the NR Board:

Based on the best available science, where across the
diverse Wisconsin agricultural landscape would
conservation systems and riparian buffers enhance the
quality of the state's waters?

The Dean of UW-Madison CALS appointed a committee of scientists to work with the agencies, organizations, and citizens of Wisconsin. This group formed the WBI and was made up of an executive committee and an advisory committee. The WBI Executive Committee consisted of representatives from state and federal natural resource agencies and UW scientists. The WBI Advisory Committee operated under an open door policy in that any vested interest was welcome to participate. Meetings of the WBI Advisory Committee have been held approximately every quarter since its formation in 2002. The WBI Executive Committee met during the first year of the WBI to establish the direction and decision protocols but has been relatively inactive since that time. It is important to note that although UW scientists composed a significant part of the WBI Advisory Committee, the WBI was not intended as a top-down process. Rather than beginning with scientific facts and asking for commentary on those facts, the WBI process purposively began with a blank agenda other than the initial charge from the NR Board. This civic science strategy allowed for an initial, open exchange between all vested interests on what research questions needed to be addressed and what type of research would have credibility for all WBI participants and the landowners of Wisconsin.

#### References

Correll, D. 2003. Vegetated Stream Riparian Zones: Their Effects on Stream Nutrients, Sediments, and Toxic Substances (Thirteenth Edition). Available at http://www.unl.edu/nac/ripzone03.htm.

USDA National Agricultural Statistical Service. 2005. Wisconsin Annual Statistical Bulletin. Available at: http://www.nass.usda.gov/Statistics\_by\_State/Wisconsin/Publications/Annual\_Statistical\_Bulletin/index.asp

Wisconsin Academy of Sciences, Arts and Letters. 2003. Wisconsin's Waters: A Confluence of Perspectives. *Transactions* 30.

### 2. Establishing the Course for the WBI

The WBI used the process of civic science to meet the charge of the NR Board. Civic science occurs when citizen and scientist work together, with neither viewed as superior or subordinate, to develop an agenda, determine appropriate methodology, and agree on a desired outcome to this agenda. Both citizen and scientist are recognized as having valuable contributions to make to meet the mutually desirable outcome. Each group gains from participation in this process. Scientists begin to understand both the concerns and knowledge of participating citizens. In turn, these citizens begin to understand the complexity and challenges associated with conducting robust and valid research.

### WBI Participants

In addition to the UW scientists, representatives from the River Alliance of Wisconsin, Trout Unlimited, Wisconsin's Environmental Decade (Clean Wisconsin), The Nature Conservancy, Farm Bureau Federation, Wisconsin Corn Growers Association, and Professional Dairy Producers of Wisconsin participated in this process. Professional associations such as the Wisconsin Association of Land Conservation Employees (WALCE) and the Wisconsin Land and Water Conservation Association (WLWCA) were also active participants, as were representatives from both state and federal conservation agencies that included the DNR, DATCP, US Department of Agriculture Natural Resources Conservation Service (NRCS), US Department of Agriculture Farm Services Agency (FSA), and US Geological Survey (USGS). Input was also sought from constituencies from outside the WBI on a regular basis; several dozen presentations were made each year at meetings ranging from local town meetings to statewide association conferences. Ideas and comments from these meetings were then brought before the WBI Advisory Committee for consideration.

### Addressing the Question of Buffers

Many of the initial WBI Advisory Committee meetings were spent gauging positions and commitment from those participants who had been involved in the earlier Redesign of the Nonpoint Pollution Program debates. Early discussions at these meetings vacillated between "buffers are good" arguments to "protect private property rights" positions. Rather than debating the merits of these two contrasting positions, both of which have merit, the discussion was refocused on the type of research necessary to begin to address underlying commonalities. Consensus was finally reached on four basic questions that needed to be addressed by the WBI.

- Where are buffers most needed across the diverse landscapes and land uses found in Wisconsin?
- What *types* of buffers are needed in these specific locations?
- What are the consequences when buffers are installed in these specific locations?
- What will be needed to get these buffers into these specific locations? Subcommittees were formed to address each of these core questions, with the intention that each subcommittee would meet independently to develop recommendations to submit to the full WBI Advisory Committee.

With these core questions established and a committee structure developed, the WBI discussion turned to what is known about buffers from both a scientific and programmatic viewpoint. Many of the participants in the WBI went on a two-day fact-finding trip to central Iowa. Iowa is often cited as a leader in the Midwest relative to the amount of riparian buffers it has installed during the last decade using federal and state conservation programs. This trip gave WBI participants the opportunity to view some of Iowa's buffer efforts and to discuss buffer programs with local, state, and federal program managers.

As acknowledged in the 2002 ad hoc committee report, buffers can have benefits relative to water quality, wildlife, endangered species, and aesthetic values, among others. However, in keeping with the charge from the NR Board, a decision was made early in the WBI process to focus on the water quality benefits of riparian buffers as part of NR 151.

#### WBI Decisions on the Role of Riparian Buffers in Intervention Efforts

The outcome of early discussions—recognizing that buffers need to be part of a larger conservation system and that this system needs to be designed to address water quality concerns—became the foundation for the development of the specific research challenges for UW scientists. This decision implied shifting the intervention effort from trying to stop the water at the stream edge (the focus of the original controversy) to an intervention effort that used an array of conservation practices further up on the landscape to minimize the amount of water reaching the riparian area. Riparian buffers were viewed in subsequent WBI discussions as a measure that would be recommended only if these changes further up in the landscape could not adequately address the water quality concerns.

The WBI further decided that water quality should determine whether assessment of intervention efforts is necessary; if water quality indicates that intervention is needed, then the intervention should begin on the landscape draining into the riparian area of concern. A systems approach should be used in which conservation practices in the upland areas minimize the transport of pollutants into the riparian area.

Much of the scientific literature on buffers examines the design and composition of buffers relative to their effectiveness in isolation or an experimental setting. These buffers were then assessed regarding their ability to remove sediments and nutrients. Yet the systems approach adopted by the WBI asked the question, why allow these sediments and nutrients to move to the riparian area in the first place? Would it not be more effective to retard the movement of sediments and nutrients in the upland area, thereby minimizing the need for large buffers in the riparian area? This broader perspective that buffers are part of conservation systems means that WBI recommendations are cognizant of and compatible with other soil conservation and water quality programs and regulations.

Both the agricultural and environmental interests in the WBI process were interested in protecting water quality. All of those involved recognized that taking a significant amount of land out of production in riparian areas would be both controversial and expensive. Consequently, a conservation systems approach appealed to all sides because water quality would be protected while the amount of land taken out of production would be minimized.

#### Using the Best Available Science to Address the Buffer Question

The complementary role of conservation systems and riparian buffers in the WBI recommendations is nothing new as conservationists have been advocating this approach for decades. What is new is how the best available science was used to advance this perspective regarding the four questions underlying the WBI process. In particular, the first question (*Where are buffers most needed across the diverse landscapes and land uses found in Wisconsin?*) is where the WBI scientists made significant advances. As will be explained in the next chapters of this report, two critical qualifications were added to this basic question.

First, contributions of WBI scientists advanced the first question from simply asking where to one that asked where is the greatest probability of getting a meaningful water quality response to the implementation of conservation systems including riparian buffers. This rephrased question implied the need to rank the diversity of Wisconsin's agricultural landscapes using a common set of criteria. After a significant amount of discussion, the WBI Advisory Committee agreed that the watersheds in the Wisconsin landscape should be ranked on the ability to (1) improve stream water quality, (2) protect and restore aquatic biological communities, and (3) sustain lake water quality. As addressed later in the report, a critical question to this ranking process was the delineation and selection of the watersheds being ranked.

The second, science-based contribution of the WBI was to specify what should happen within the ranked watersheds. Participants agreed that simply identifying priority areas was insufficient, so procedures were developed to identify the specific portions of the landscape that had the greatest probability of contributing to water quality degradation. The conservation systems and riparian buffer approach would initially be advanced in these areas in order to maximize the probability of positive water quality responses.

Ultimately, the WBI began by looking at the state and ended by focusing on specific fields. This focus was achieved by using the best available science to craft a strategic decision that resulted in identifying specific areas where local activities would have the greatest probability of achieving intended water quality outcomes. Besides the inherent effectiveness and efficiency associated with this approach, the WBI collaboration found acceptance to this strategy across a wide spectrum of different interests.

#### Conclusion

No specific recommendation will be made regarding the civic science process followed by the WBI participants, but it is hoped that those who read this report will recognize that this process should be considered when addressing other natural resource management issues. The civic science philosophy embedded in the WBI bears repeating in other settings because it was much more than simply facilitating public participation. The diverse interests represented in the WBI were equal participants in a process that guided and assessed the scientific contributions of UW-Madison researchers. The contributions of these diverse groups were indispensable to the WBI process, and their efforts have fundamentally strengthened the WBI's final product. The close collaboration between these diverse interests and the University that comprised the WBI process

exemplifies the Wisconsin Idea—that our state academic institutions should contribute directly to the betterment of the citizens of Wisconsin.

# 3. Research Questions on Conservation Systems and Riparian Buffers in Wisconsin

The previous chapter described how the WBI used the civic process to focus on specific questions that would become part of the final recommendations. Although the NR Board acknowledged the diversity of the agricultural landscape in its charge to the WBI, it did not specify which watersheds to study. The agricultural land within these watershed units can vary significantly. Moreover, there can be important variation within farms located in these watersheds. Finally, if strategic placement of a conservation system, possibly including riparian buffers, needs to occur, then this implementation may include the subfield, the farm, and possibly portions of neighboring farms. Thus, in response to the earlier "where in Wisconsin" question, it was agreed that WBI research needed to address questions that range from the entire state of Wisconsin down to the subfield. Consequently, the original four basic questions (p.6) evolved into the following four specific questions:

- 1. How do we develop a buffer implementation strategy based on adaptive management?
- 2. How do we identify the watersheds that have the greatest probability of showing demonstrable improvements with investment in conservation and buffer systems?
- 3. What types of tools can be developed that can be employed at the local level to assess watersheds and fields where conservation and buffer systems have the greatest probability of addressing water quality degradation?
- 4. How do we develop techniques for determining the optimal placement and configuration of conservation and buffer systems on designated landscapes? Each of these questions was used to solicit additional research from UW scientists that generated information that was then used to develop WBI recommendations. Much of this research was based on cooperative efforts with ongoing research. This included research being conducted by the Wisconsin Agricultural Stewardship Initiative on Discovery Farms, farm systems research being conducted at the UW-Platteville Pioneer Farm, UW-Madison Center for Limnology, and UW-Madison CALS work on related nutrient management and conservation research as part of the Wisconsin Agricultural Experiment Station research agenda. This ability to leverage WBI funds with ongoing research allowed the WBI to develop the impressive and rigorous set of research findings that underlie the WBI recommendations. It should be emphasized that the recommendations developed from this research are based on the best available science at this time in accord with the charge from the NR Board.

# 4. Adaptive Management as a Basis for Natural Resource Management Programs

Solutions to problems cannot be commanded. They must be discovered.

-Lee 1993

A major goal of the WBI is to set a course for achieving *measurable and substantial* improvements in water quality in the most efficient and cost-effective manner. An early outcome of the WBI process was recognition of the uncertainty associated with predicting the effectiveness of measures intended to protect and improve water quality. There was also recognition that cause-and-effect relationships are often not well-known, and that current knowledge does not provide easy answers as to how to best address Wisconsin's water quality concerns. Finally, unexpected outcomes (surprises) are surprisingly common in natural resource management. WBI participants maintained

diverse ideas about the most effective approach for making improvements in water quality. Rather than reaching a stalemate regarding this lack of consensus, the WBI came to realize that this represented an opportunity to embrace these diverse views and employ a resource management approach called *adaptive management*.

What is the WBI's vision of adaptive management? The starting point is the WBI watershed (Chapter 5) as the scientifically appropriate "management unit." The WBI watersheds are large and complex systems. Environmental data are limited and science does not provide adequate theory to allow precise prediction of how in-stream water quality will respond to implementation of the WBI recommendations. Uncertainty comes in two forms: (1) how much and how quickly in-stream water quality will respond to the implementation of buffers, and (2) what are the most effective ways to go about implementing WBI recommendations. The first involves exclusively the response of the natural ecosystem to implementation, and the second involves the broader socio-ecological system.

Implementation of our recommendations in WBI watersheds (Chapter 5) should be viewed as ecosystem experiments and should

Natural Resource Management Approaches

One way to understand adaptive management is to contrast it with other natural resource management approaches. Each has different ways of dealing with the uncertainty and unexpected outcomes.

- Externally prescribed: "thou shalt" regulations from a regulatory agency. This approach largely ignores uncertainty and unanticipated outcomes, resulting in episodic changes to the regulations.
- Error and no trial: based on theories such as how markets work, how people make decisions, or the appropriate role of government. While the theory often attempts to account for uncertainty, unanticipated circumstances often cause error and inefficiency.
- Trial and error: policy is tested on a small scale to assess whether large-scale implementation is feasible and effective.
   This is dependent on whether conditions of the small-scale test can be extended to the larger area.
- Adaptive management: policy is designed to learn from implementation experience and is adapted on the basis of new knowledge. One goes forward based on current knowledge, but structures the process so as to learn from unexpected outcomes.

serve as the primary vehicle for learning and generating new knowledge. Adaptive

management is a formal process for continually improving management policies and practices by learning from the outcomes of implementation. Surprises in ecosystem response are not viewed as failures but, instead, as a source for learning better ways of accomplishing water quality goals. There are four central pillars to the WBI adaptive management approach:

- 1. Use of reference and treatment watersheds
- 2. Replication of watersheds
- 3. Environmental monitoring
- 4. Adapting the program in response to new knowledge

Use of reference watersheds ensure that changes are not driven by external factors such as weather. Replication is needed because for knowledge to be reliable, it needs to be shown to work on more than one ecosystem. Monitoring is designed to detect ecosystem-level changes, such as sediment-sensitive indicator species and loads of nitrogen, phosphorus, and suspended sediments. These ecosystem experiments accelerate the rate of learning, and new knowledge feeds back into on-the-ground management and allows for constant improvement. Furthermore, the effectiveness of competing or alternative management approaches can be compared within the WBI adaptive management framework.

### Monitoring in Adaptive Management

As noted above, monitoring is central to the adaptive management process. Adaptive management is based on learning from current management efforts. The WBI Advisory Committee has developed a set of recommendations for monitoring the implementation of conservation systems and riparian buffers (Appendix A). The recommendations were developed around a series of core questions that emerged from WBI Advisory Committee discussions and are designed to address these core questions:

- 1. Baseline: What was the status of the watershed before implementation efforts?
- 2. Implementation: Is the plan being implemented as intended, and is it consistent with the county Land and Water Resource Management Plan?
- 3. Effectiveness: To what extent is the implementation effort having the desired effect relative to the water quality objectives?
- 4. Efficiency: What is the cost per unit gain toward the desired water quality objective?
- 5. Scientific Validation: Are predicted water quality responses to implementation observed?
- 6. Ownership: Has the management process increased natural resource stewardship? It is critical to emphasize that monitoring must provide timely feedback concerning program success. This feedback is necessary for agency administrators to make improvements in the implementation process. A relatively small investment in an effective and well-designed monitoring program can vastly improve the effectiveness and efficiency of natural resource management programs. Chapter 1 of this report pointed out that the State of Wisconsin invested \$187 million in state funds in nonpoint pollution efforts. One reason cited for abandoning that program was that the effectiveness of this program could not be determined. An adaptive management program will allow the most judicious use of limited funds dedicated to natural resource management. Although the WBI does not prescribe specific water-quality management targets (for example, a 50%

percent reduction in stream phosphorus concentrations), adaptive management provides a roadmap for achieving specific targets and generates the knowledge required to make improvements in water quality. The WBI Advisory Committee encourages our elected leaders and decision makers to take advantage of the benefits of adaptive management.

There is no doubt that adaptive management will require a profound paradigm shift for decision makers, administrators, scientists, and technicians. Integrating adaptive management into Wisconsin's natural resource management efforts will not be a simple task and goes beyond the charge to the WBI. This challenge will likely be taken up by the Wisconsin Legislature, agency administrators, and the citizen boards that advise these agencies. The WBI recommends that the opportunity of specifying the role of riparian buffers within NR 151 be used to begin the process of integrating an adaptive management approach into natural resource management in the state of Wisconsin.

#### Reference

Lee, K. N. 1993. Compass and Gyroscope: Integrating Science and Politics for the Environment. Washington, D.C.: Island Press.

# 5: Targeting Buffer Implementation on a Statewide Scale

The WBI approach differs dramatically from previous buffer implementation programs in that the cornerstone of the WBI process is to install buffers at *carefully* selected sites where potential benefits are greatest. Most programs are offered on a firstcome, first-served basis, resulting in a highly dispersed distribution of buffers on the landscape. As a result, the extent of buffers in any one watershed tends to be low, and studies at the watershed scale have often indicated marginal effectiveness (Parkyn et al. 2003, Moerke and Lamberti 2003, Nerbonne and Vondracek 2001, Wolf 1995). Although buffers may adequately protect the stream reach closest to where they are implemented, the benefits may be masked by unbuffered areas elsewhere in the watershed. All this indicates that water quality improvements may not be detectable at the watershed scale without careful and extensive placement of conservation and buffer systems within the watershed. Aggregated implementation of buffers is needed to counter this masking effect and create improvements that endure downstream. The water quality benefits of this approach would achieve improvements in water quality at the outlet of the watershed that exceed those of a geographically dispersed, available-to-all approach (Figure 5.1). Benefits of buffers can be further amplified by targeting watersheds based on their potential responsiveness to buffers.

The WBI strategy is designed to result in a successful nonpoint-source pollution-reduction program that leads to improvements that are measurable and noticed by the public. This approach also uses public funds more efficiently than other programs because directing funds to areas that are most likely to respond to buffers will result in more benefit per dollar spent. All of these benefits would bolster public support for future expenditures.

In this chapter, we explain how we derived a set of watersheds that were used in an analysis of their suitability for conservation system and riparian buffer implementation. We then developed a system for scoring and ranking these watersheds according to the environmental benefits derived from conservation system and riparian buffer implementation. Finally, we combined these rankings to produce a composite list reflective of the potential water quality benefits of buffers.

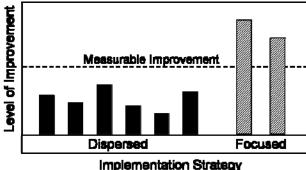


Figure 5.1. The difference between hypothetical focused and dispersed buffer implementation strategies. In both strategies, total improvement (represented by the sum of the heights of the bars) is equal and is the result of equal effort. In the dispersed strategy, a few highly polluting fields in each of six watersheds are buffered, but none of the watersheds improve measurably. In a focused strategy, all highly polluting fields in two watersheds are buffered and both watersheds improve measurably.

### The Watershed Approach

The watershed is defined as the area of land that drains to a river or other aquatic system. The watershed is a central concept to the WBI and is the appropriate unit for implementing and monitoring this program for the following reasons:

- Water quality in a stream is a function of *upstream* land use activities.
- The watershed explicitly includes all streams within the watershed area, including small headwater streams. These small streams are critical connections between land use and water quality (Meyer et al. 2003) but are among the least protected of natural resources (Peterson et al. 2001).
- Watersheds are convenient geographic units for implementation because they are stable over time (Bohn and Kershner 2002).
- Changes in water quality at the outlet of the watershed can be easily monitored over time (Bohn and Kershner 2002, McNitt and Kepford 1999), which facilitates the adaptive management process.

### Choosing an Appropriate Watershed Size

Choosing an appropriate watershed size is important to the success of a nonpoint-source pollution-reduction program because the costs and benefits of the program are affected by

the size of the watershed. As watershed size increases. cumulative environmental benefits increase, but cost is also higher and evaluating the entire area is more difficult and complex. As watershed size decreases, cumulative benefits are smaller; however, on the plus side, costs are lower and evaluating changes induced by an implementation effort is less difficult. Therefore, an intermediate-sized watershed area may be ideal because it allows for improvements in water quality while keeping costs down by concentrating program resources in a limited area (Figure 5.2).

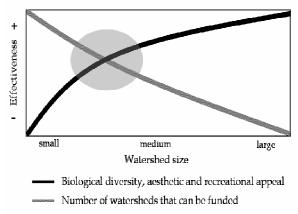


Figure 5.2. Conceptual illustration of tradeoffs in various aspects of program effectiveness across a range of watershed sizes. The gray oval indicates the size that maximizes the average effectiveness across criteria.

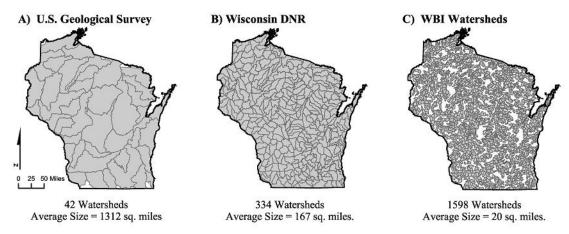


Figure 5.3. Three Wisconsin watershed delineations.

We examined two previous Wisconsin watershed delineation efforts by the USGS and the Wisconsin DNR used in the Nonpoint Priority Watershed Program to see if those watersheds would be appropriate for the WBI. After careful consideration, we chose to define a new set because the watersheds delineated by these agencies were too large and were hydrologically incomplete. Figure 5.3 compares these previous efforts with the WBI watersheds. The WBI watersheds were derived from a geographic information system (GIS)-based analysis of stream hydrography and digital elevation models. As such, they should be hydrologically consistent and true watersheds based on ridgeline divides.

The WBI watersheds range from 7.7–38.6 square miles (4,900–24,700 acres, 20–100 square kilometers) in area. Each of these watersheds is a candidate for implementation of riparian buffers and upland management technologies. It is important to note that the WBI watersheds do not include 100% of the land area in Wisconsin; any areas that drain more or less than the targeted size range were not included in the WBI watersheds (shown in white in Figure 5.3).

The best available science indicates that the scale of these WBI watersheds is optimal for identifying and ranking contributions to nonpoint pollution, cost-effective for implementation efforts, and small enough to be meaningful to local conservation staff and landowners.

# Assessing Watershed Responsiveness to Buffers

According to the scientific literature (e.g., Wenger 1999) and natural resource management guidelines (Iowa State University 1997, USDA 1998), agricultural conservation practices, including the installation of riparian buffers, are capable of many functions, including filtering sediments and removing nutrients and pesticides from runoff, stabilizing stream banks, promoting biological diversity, regulating stream temperatures, promoting native plant restoration, and serving as dispersal corridors for both aquatic and terrestrial organisms. The WBI Advisory Committee focused on a subset of these functions to develop management goals that can specifically address agricultural nonpoint pollution, can respond meaningfully on the WBI watershed scale, and can be assessed statewide using readily available data.

Following significant discussion and debate, the WBI Advisory Committee reached consensus that the desired management goals should be (in no particular order) (1) improve stream water quality, (2) protect and restore aquatic biological communities, and (3) sustain lake water quality. Each of the 1598 WBI watersheds was rated on its potential responsiveness toward each of these management goals. The following three sections describe the conceptual development for each goal while summarizing the technical procedures used to assign responsiveness ratings to each watershed. The last section describes the combination of individual goal ratings into an overall composite rating.

### Goal 1: Improve Stream Water Quality

Excessive amounts of nutrients and sediment are delivered to streams and downstream water bodies from agricultural land (Carpenter et al. 1998). This pollution impairs the use of these waters by humans for recreation and drinking water and negatively affects aquatic organisms (USEPA 2000). Reducing nutrient and sediment delivery to surface waters through conservation systems and riparian buffers may mitigate these effects.

Research by the USGS (Corsi et al. 1997) reveals considerable variability in the amount of sediments and nutrients delivered by watersheds across the state. This variability is useful from a targeting standpoint because it suggests that changes in land management may produce a much larger benefit in some areas than in others. To quantify this benefit, we built a simple model to estimate the nutrient and sediment reduction potential in each of the 1598 WBI watersheds.

For each constituent (sediment, phosphorus, nitrogen), load reduction potential is equal to the current load minus the sum of unbufferable sources. For the purposes of this analysis, unbufferable sources are defined as those that cannot be attenuated by riparian buffers. These include point sources (e.g., discharges from sewage treatment plants), fine-textured soils that release sediments that are not trapped by buffers, and stream bank erosion. The WBI management goal is to maximize the reduction potential, i.e., to select watersheds that have high loads of sediment and nutrients from agricultural activities, most of which can be substantially attenuated using conservation systems, including riparian buffers.

### **Current Loads**

To estimate current constituent loads for the WBI watersheds, we first constructed multiple linear regression models. These models relate measured sediment and nutrient loads in streams across Wisconsin (USGS unpublished data) to those streams' watershed characteristics. Average annual loads with at least three years of records were available for 116 sites. We used a GIS and widely available spatial data to calculate characteristics of the streams' watersheds that we expected would influence sediment and nutrient levels. These characteristics included measures of land cover, precipitation, soils, slope, and the stream network. We then used stepwise variable selection to choose models with high explanatory power and whose structure was consistent with our knowledge of landscape processes. In the statistical models, variables measuring land cover in stream riparian zones (Figure 5.4) explained most of the variability in sediment and phosphorus loads. Conversely, total nitrogen and nitrate were primarily driven by the percentage of agriculture in the overall watershed. These results suggest that riparian restoration coupled with upland conservation systems in agricultural watersheds will likely result in

greater reductions in phosphorus and sediment than in nitrogen. The ability of conservation systems and riparian buffers to attenuate nitrogen delivery to streams is governed largely by different factors (see sidebar).

We then used the regression equations for each water quality constituent and GIS-derived watershed characteristics to predict annual loads for each of the WBI watersheds. Since annual loads have been measured in only a few of these watersheds, these predictions are the best available estimates of current statewide patterns of nutrient and sediment loads.

### Unbufferable Sources

If riparian buffers were capable of eliminating all of the sediment and nutrients that would otherwise be transported to a stream, then rating watersheds for goal one would only require an estimate of the current load. However, some sources are unbufferable. We categorized unbufferable sources of phosphorus and sediment into the following types: meander-belt (stream bank) erosion, fine-textured soils, point-source discharges, and urban stormwater.

*Meander-belt erosion*. The annual sediment load of a stream derives from both upland and channel sources. Channel sources may be divided into two components: that which originates from devegetated banks and that

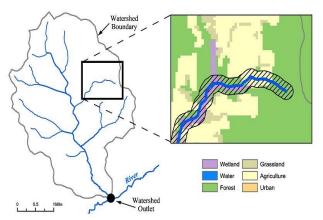


Figure 5.4. An example watershed showing the stream network (blue lines). The inset shows the riparian zone (cross-hatched area) for one stream. To build the models that predict current phosphorus and sediment loads, we analyzed the land use within the riparian zone as well as the land use within the entire watershed. We used a very detailed stream map; in some areas of Wisconsin the smallest streams are intermittent.

### Buffers and Nitrogen

Unlike sediment and phosphorus, whose transport from fields to streams is largely driven by surface runoff, nitrogen is transported primarily by subsurface flow. For nitrogen to be removed from groundwater before it reaches surface water, the groundwater must move relatively slowly through the root zone of riparian plants (Mitsch et al. 2001). Therefore, nitrogen load reduction potential is largely driven by local variation in subsurface hydrology. Baker and colleagues (2001) developed a GISbased model that links nutrient dynamics to riparian hydrology across Michigan. These methods could be applied to Wisconsin in the future to identify watersheds with high nitrogen reduction potential.

which originates from meander-belt migration. Sediment losses from devegetated banks can be reduced using buffers (Zaimes et al. 2004), particularly if the cause is from cattle grazing and cattle are excluded from the buffer zone. In contrast, meander-belt migration is a natural process whereby streams erode sediments from their floodplains as they adjust their course. In some areas of Wisconsin, the combined effects of poor past land management, steep slopes, and erodible soils have created conditions in which meander-belt erosion is accelerated. This sediment source is for the most part not reducible using buffers (Trimble 1993). We based the estimate of meander-belt erosion on a measure of land form and the amount of agriculture in the stream riparian zone. This estimate was

converted to a proportion of the total annual sediment load based on measurements at Coon Creek, Wisconsin, where a detailed study was conducted (J.C. Knox, UW-Madison Geography Dept., unpublished data). This equation predicts values ranging from 5% to 36% for the WBI watersheds. In some Lake Superior tributaries in northern Wisconsin, a severe form of meander-belt erosion contributes as much as 90% of the sediment load to streams (Fitzpatrick et al. 2004).

Fine-textured soils. Strategically located conservation systems and riparian buffers are capable of trapping a large portion of sediment derived from upland sources. However, in areas with fine-textured soils, such as clay, a larger portion of this sediment remains suspended in runoff as it moves through the buffer and is delivered to the stream. Phosphorus attached to these soil particles is also more likely to pass through the buffer. Therefore, regional variation in soil texture influences the potential for buffers to reduce sediment and phosphorus delivery to streams. We estimated the proportion of sediment and phosphorus that will pass through a buffer as one-fourth of the average proportion of clay in the surface soils within a watershed. Predicted values of soil-texture-related unbufferable phosphorus and sediment range from 0% to 11% for WBI watersheds. **Point-source discharges.** In some streams, part of the nutrient and sediment load comes from point sources, such as sanitary sewage treatment and industrial facilities. Since these discharges deliver pollutants directly to streams through pipes, buffers are not capable of mitigating against their effects. We estimated the annual load of phosphorus and total suspended solids from all point sources in the Wisconsin Pollutant Discharge Elimination System database using records from 2004. We summed these loads within each WBI watershed to estimate the fraction of the total derived from point sources. Values for suspended solids were generally insignificant relative to nonpoint source estimates, but point sources of phosphorus in a few watersheds exceeded the nonpoint source total. Urban stormwater. Much of the runoff from urban areas is carried directly to receiving waters by stormwater conveyances such as pipes and concrete channels. A recent study in the Chicago area found that buffers of natural vegetation did little to mitigate against the effects of urban areas, presumably because of hydrologic alterations that bypass riparian areas (Fitzpatrick et al. *in press*). We considered the proportion of urban land in each WBI watershed (0% to 82%) to be unbufferable.

### Sediment and Phosphorus Reduction Potential

The statewide pattern of sediment and phosphorus reduction potential was similar. Therefore, we combined these ratings into a composite reduction potential, weighing sediment and phosphorus equally (Figure 5.5). Buffer implementation in watersheds that score highly for load reduction potential will contribute most to the goal of improving stream water quality by reducing stream sediment and phosphorus loads. The statewide pattern of load reduction potential is largely driven by variation in current loads. Most watershed ratings were not significantly affected by accounting for unbufferable sources. However, load reduction potential in some watersheds was greatly reduced because of large contributions from one or more unbufferable sources. Accounting for the presence of these sources is important in screening out watersheds where buffers cannot be as effective at improving water quality.

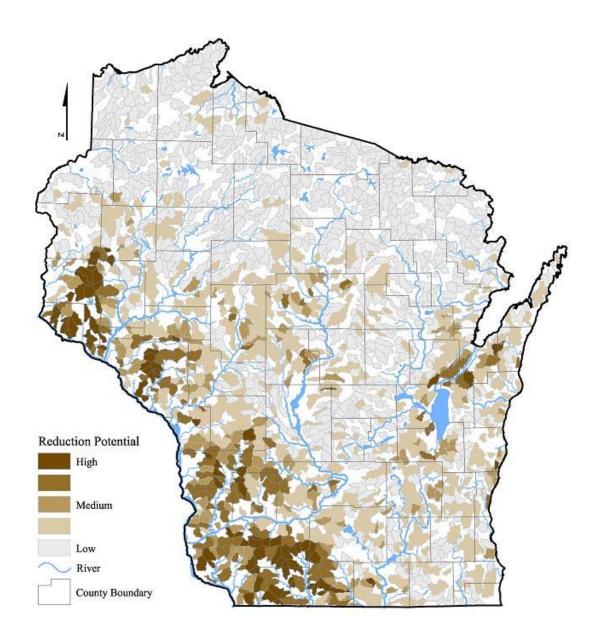


Figure 5.5. Potential for reduction of phosphorus and sediment in WBI watersheds.

# Goal 2: Protect and Restore Aquatic Biological Communities

Aquatic life has been negatively impacted by agriculture (Wang et al 1997). Improving water quality through the methods outlined in nutrient and sediment reduction will certainly help reverse some of these effects. Overall condition of the aquatic biological community has been acknowledged as a useful indicator of stream health and can be even more informative than direct measures of water quality (Wang and Lyons 2003). In this section, we address the following questions: Can conservation systems and riparian buffers correct the most critical, limiting effects of agriculture on aquatic organisms? Will some organisms respond more strongly to improvements? And most importantly, from a geographic targeting standpoint, in what settings can we expect to see

the strongest overall biological response to conservation systems and riparian buffer implementation?

Aquatic biological communities typically consist of fish, insects, mussels, crustaceans, and plants. Of these, fish are by far the best known by science in terms of their distributions and habitat needs. They are also useful as an indicator of overall biological conditions because they depend on other organisms for food. Finally, fishing is important recreationally and economically (Wisconsin DNR 2005). For these reasons, we chose to focus on fish as a surrogate for assessing biological community status.

Excessive suspended sediment and siltation have been documented (Wood and Armitage 1997) as having widespread and serious effects on many fish species. Wisconsin fish species vary greatly in their sensitivity to sediment (Becker 1983). Modes of sensitivity include spawning requirements for coarse substrate, sight-dependent feeding, and feeding on other organisms that require coarse substrate. Buffers have been shown to be particularly effective at reducing sediment inputs from runoff, with removal efficiencies as high as 95% in some studies (Peterjohn and Correll 1984).

From among the 150 fish species in Wisconsin, we chose eighteen (Table 5.1) that are particularly sensitive to sediment, live in small- to medium-sized streams, and are common enough to assess their habitat preferences. All of these species are native to Wisconsin except brown trout, which is

naturalized. Six fish families are represented. Most of the species are not game fish.

Using fish occurrence and environmental data from the USGS's Aquatic Gap Analysis Project database, we developed habitat models using logistic regression for each of these species and simulated the effect of current human land use. The model predicted the likelihood of a species' presence, which depended both on the sensitivity of the fish to deleterious land uses and on the amount of these land uses present in the watershed. Streams where the likelihood of a species' presence improved significantly under the land-use change simulation and where the potential stream condition was good or excellent were considered good candidates for buffer implementation (red in Figure 5.6). Streams that did not change (green in Figure 5.6) or only improved from poor to marginal (yellow in Figure 5.6) in the simulation were considered poor candidates.

Table 5.1. Wisconsin Stream Fish Species Used in Development of Biological Responsiveness Potential.

Common name	Family	
Chestnut lamprey	Lamprey	
Northern brook lamprey	Lamprey	
Silver lamprey	Lamprey	
American brook lamprey	Lamprey	
Redside dace	Minnow	
Blacknose shiner	Minnow	
Longnose dace	Minnow	
Northern hog sucker	Sucker	
Silver redhorse	Sucker	
River redhorse	Sucker	
Golden redhorse	Sucker	
Greater redhorse	Sucker	
Brown trout	Trout	
Brook trout	Trout	
Mottled sculpin	Sculpin	
Rainbow darter	Perch	
Banded darter	Perch	
Logperch	Perch	

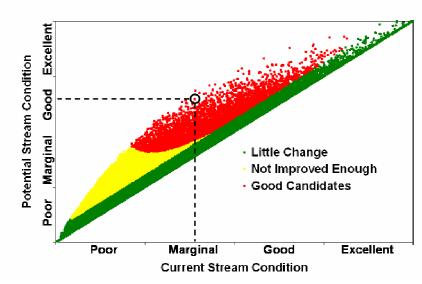


Figure 5.6. Responsiveness scores for one of eighteen sediment-sensitive fish species. Each point on the figure represents a stream (solid areas are tightly clumped points). The stream represented by the circled point is an example of a good candidate for buffer implementation because the predicted habitat improvement (difference between current and potential condition) is large and the potential condition is good.

We repeated this procedure for all eighteen sediment-sensitive species, which produced a responsiveness score for each species for every stream in a given WBI watershed. Next, we took the highest score for each species from among the streams in the watershed and summed those for the biological responsiveness score. This total biological responsiveness score (Figure 5.7) indicates the potential degree of improvement in populations of sediment-sensitive fishes. All of the highly rated watersheds are in areas with significant agricultural land use. However, in some agricultural areas, other environmental characteristics, such as slope and soil type, create stream conditions that are not favorable for most of these species. In these particular watersheds, reductions in sediment would not be as effective at improving conditions for fish because other factors would limit the response.

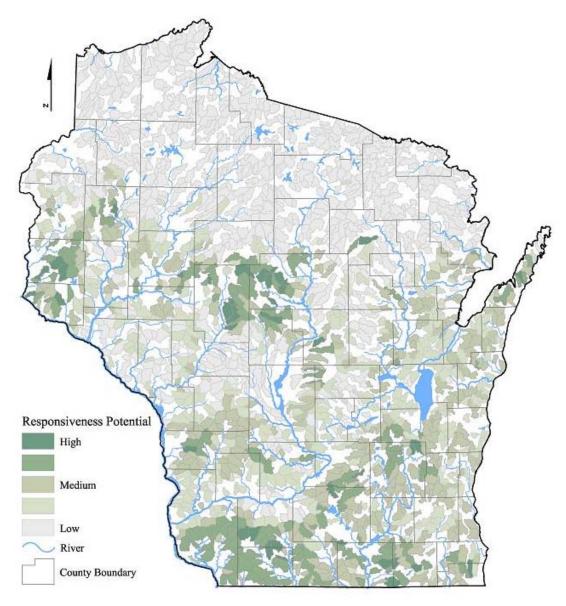


Figure 5.7. Biological responsiveness indicates the potential for improvement in populations of sediment-sensitive fish species.

# Goal 3: Sustain Lake Water Quality

Lakes are one of Wisconsin's most prized natural resources. In addition to providing habitat to a large range of species, they are places for people to recreate. Wisconsin's lakes are a major reason why Wisconsin generated more than \$8.4 billion in recreation and tourism revenue in 2004 (Wisconsin Department of Tourism 2005). Clean and clear *oligotrophic* lakes are generally more desirable than *eutrophic* lakes that have poor water clarity due to the presence of weeds and frequent algal blooms.

Weeds and algae demonstrate that a lake has become more "productive" due to an increase in nutrients in the lake water. This increase in production is a process called *eutrophication*. Human activities, however, have greatly increased the rate at which

nutrients are added to lakes and, as a result, the water quality in lakes has degraded. Agriculture is one of the activities that contribute excessive nutrients into Wisconsin lakes. When runoff from farmland goes into streams during snowmelt or rainfall, it carries phosphorus into Wisconsin's lakes, where it fuels algal blooms.

Conservation systems and riparian buffers can reduce the amount of phosphorus that enters streams and degrades the quality of Wisconsin lakes. The following describes our effort to identify the upstream watersheds where buffers are most likely to help attenuate eutrophication and sustain good lake water quality.

### Assessing Phosphorus Reduction Potential for Lakes

We targeted lakes that were most likely to respond to the implementation of conservation systems and riparian buffers in the WBI watersheds. Specifically, we used the following criteria to select the lakes:

- Drainage lakes: Only some of the lakes in Wisconsin are fed by surface water. Since conservation systems and riparian buffers are primarily intended to remove phosphorus from surface streams, we eliminated seepage and spring lakes from the WBI analysis.
- Watershed area: In order to preserve our efforts to provide the best targeting possible, the lakes used in our analysis are either located within one of the WBI watersheds or downstream from three or fewer WBI watersheds. In addition, we disregarded lakes with less than 75% of their watershed in WBI candidate watersheds.
- Current Trophic State: Shifting a lake from a eutrophic state to an oligotrophic state is far more difficult than slowing the rate of cultural eutrophication. As a result, this analysis only considered lakes that are oligotrophic or marginally eutrophic (i.e., mesotrophic) state. All lakes in this analysis have a Trophic State Index value of 55 or less, as calculated by the Satellite Lake Observatory Initiative (SLOI) (Chipman et al. 2004). In addition, only lakes that are large enough to provide a suitable spectral signature were included in the SLOI data set. This limitation eliminates many of Wisconsin's smaller water bodies (e.g., farm ponds) from this analysis.

Three hundred fifty-five lakes satisfied these screening criteria. The phosphorus reduction potential for these lakes was calculated based on the reduction potential in the upstream WBI watersheds. Figure 5.8 shows the final results of this analysis.

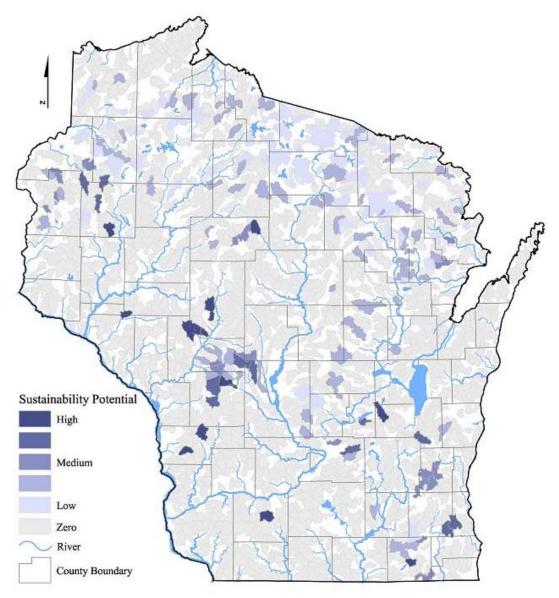


Figure 5.8. The dark watersheds in this map have the most potential to sustain lake water quality through the reduction of phosphorus inputs. The lightest gray WBI watersheds are those that were not included in the lake analysis.

# A Composite Responsiveness Score

The three goals described above were used to identify watersheds that are likely to respond most strongly to implementation of conservation systems and riparian buffers. Each goal focuses on one aspect of what people value in streams and lakes and identifies places where conservation systems and riparian buffers have the greatest probability of having a measurable, positive impact. We used the responsiveness scores from each goal to rank all the 1598 WBI watersheds. If the distribution of these ranks between the three criteria corresponded well with each other (i.e., watersheds that ranked high on one goal also ranked high on the other two goals), then ranking their overall or composite responsiveness would be easy. However, while there is moderate overlap of high ranking

watersheds in goals one and two, there is relatively poor overlap of these with high ranking watersheds in goal three. This outcome means that the best places to use conservation systems and riparian buffers to improve stream water quality and to protect and restore aquatic biological communities are not necessarily the best places for sustaining lake water quality.

Since the WBI Advisory Committee agreed that all three of these goals are important, we designed a strategy that (1) selects the highest ranked watersheds in each goal to maximize the likelihood that measurable progress will be made toward each goal; (2) selects the highest composite-ranked watersheds to maximize efficiency by contributing to more than one goal; and (3) can accommodate different levels of program resources (i.e., all goals will be addressed in comparable proportions regardless of the number of watersheds that can be funded).

We used this strategy to create a master ranked list of watersheds (Appendix B) using the following process:

- 1. Select the watershed with the highest ranking for goal one.
- 2. Select the watershed with the highest ranking for goal two.
- 3. Select the watershed with the highest average ranking for the first two goals (i.e., goal one + goal two).
- 4. Select the watershed (or group of watersheds) with the highest ranking for goal three.
- 5. Repeat steps one through four, selecting watersheds from those not already chosen, until all watersheds had been added to the list.

This procedure places watersheds into grouped rankings (Appendix B). Each group contains a watershed selected for each one of the first two goals, one selected for its high ranking in both goals one and goal two, and one to three watersheds selected for its high ranking on the goal three. This means that the 1598 watersheds were ranked in a manner that resulted in approximately 350 groups composed of four to six watersheds each (Figure 5.9). When watersheds are selected from this list for NR 151 implementation, they should be selected in these groups so that all of the goals of the WBI Advisory Committee are addressed in comparable proportions.

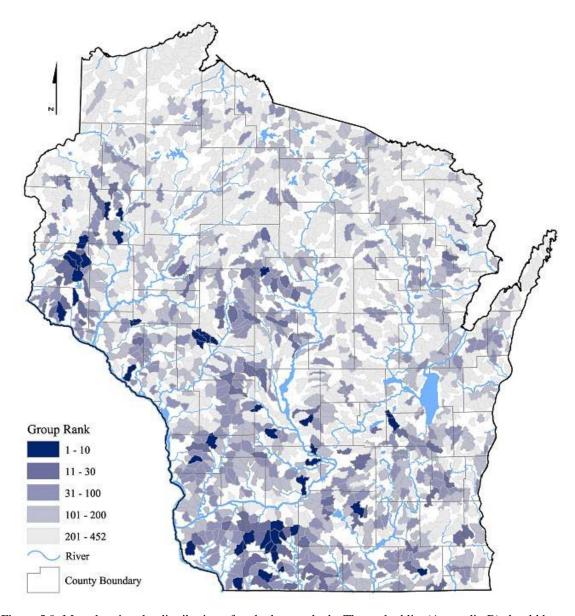


Figure 5.9. Map showing the distribution of ranked watersheds. The ranked list (Appendix B) should be used to select groups of four watersheds for inclusion in the program.

### References

Baker, M. E., M. J. Wiley, and P. W. Seelbach. 2001. GIS-based hydrologic modeling of riparian areas: Implications for stream water quality. *Journal of the American Water Resources Association* 37:1615–1628.

Becker, G. C. 1983. Fishes of Wisconsin. Madison, WI: University of Wisconsin Press.

Bohn, B. A., and J. L. Kershner. 2002. Establishing aquatic restoration priorities using a watershed approach. *Journal of Environmental Management* 64:355–363.

Carpenter, S. R., N. F. Caraco, et al. 1998. Nonpoint pollution of surface waters with phosphorus and nitrogen. *Ecological Applications* 8(3): 559-568.

Chipman, J. W., T. M. Lillesand, J. E. Schmaltz, J. E. Leale, and M. J. Nordheim. 2004. Mapping lake water clarity with Landsat images in Wisconsin, U.S.A. *Canadian Journal of Remote Sensing* 30:1–7.

Corsi, S. R., D. J. Graczyk, D. W. Owens, and R. Bannerman. 1997. Unit-area loads of suspended sediment, suspended soils, and total phosphorus from small watersheds in Wisconsin. *U.S. Geological Survey Fact Sheet* FS-195-97.

Fitzpatrick, F. A., M. W. Diebel, M. A. Lutz, T. L. Arnold, and M.A. Harris. In press. Hydrology, geomorphology, and habitat of urbanizing streams, Chicago area, Illinois. *American Fisheries Society Special Publication*.

Fitzpatrick, F. A., M. C. Peppler, H. E. Schwar, J. A. Hoopes, M. W. Diebel. 2004. Monitoring channel morphology and bluff erosion at two installations of flow-deflecting vanes, North Fish Creek, Wisconsin, 2000-03. *U.S. Geological Survey Scientific Investigations Report* 2004:5272.

Iowa State University. 1997. Stewards of our Streams: Riparian buffer systems. *Iowa State University Extension Publication* Pm-1626a.

Meyer, J. L., et al. 2003. Where rivers are born: The scientific imperative for defending small streams and wetlands. *American Rivers and Sierra Club*.

McNitt, J., and R. Kepford. 1999. Developing a new regulatory paradigm to address the impacts of diffuse pollution attributable to agriculture. *Water Science and Technology* 39:299–305.

Mitsch, W. J., J. W. Day Jr., J. W. Gilliam, P. M. Groffman, D. L. Hey, G.W. Randall, and N. Wang. 2001. Reducing nitrogen loading to the Gulf of Mexico from the Mississippi River Basin: Strategies to counter a persistent ecological problem. *Bioscience* 51:373–387.

Moerke, A. H., and G. A. Lamberti. 2003. Responses in fish community structure to restoration of two Indiana streams. *North American Journal of Fisheries Management* 23:748–759.

Nerbonne, B. A., and B. Vondracek. 2001. Effects of local land use on physical habitat, benthic macroinvertebrates, and fish in the Whitewater River, Minnesota, USA. *Environmental Management* 28:87–99.

Parkyn S. M., R. J. Davies-Colley, N. J. Halliday, K. J. Costley, and G. F. Croker. 2003. Planted riparian buffer zones in New Zealand: Do they live up to expectations? *Restoration Ecology* 11:436–447.

Peterjohn, W. T., and D. L. Correll. 1984. Nutrient dynamics in an agricultural watershed: observations on the role of a riparian forest. *Ecology* 65:1466–1475.

Peterson, B. J., et al. 2001. Control of nitrogen export from watersheds by headwater streams. *Science* 292: 86–90.

Trimble, S. W. 1993. The distributed sediment budget model and watershed management in the paleozoic plateau of the upper midwestern United States. *Physical Geography* 14:285–303.

USDA. 1998. Practice code 391: Riparian forest buffer. U. S. Department of Agriculture, Natural Resources Conservation Service.

USEPA. 2000. National Water Quality Inventory Report. Office of Water. EPA 841/R-02/001.

Wang, L. Z., and J. Lyons. 2003. Fish and benthic macroinvertebrate assemblages as indicators of stream degradation in urbanizing watersheds. In *Biological Signature Responses: Indicating Patterns Using Aquatic Communities*, ed. T. P. Simon. Boca Raton, FL: CRC Press.

Wang, L. Z., J. Lyons, P. Kanehl, and R. Gatti. 1997. Influences of watershed land use on habitat quality and biotic integrity in Wisconsin streams. *Fisheries* 22:6–12.

Wenger, S. 1999. A review of the scientific literature on riparian buffer width, extent and vegetation. Athens, GA: University of Georgia Institute of Ecology, Office of Public Service & Outreach

Wisconsin DNR. 2005. Wisconsin Fishing Report. PUB-FH-506 2005.

Wisconsin Department of Tourism. 2005. The economic impact of expenditures by travelers on Wisconsin: calendar year 2004.

Wolf, A. T. 1995. Rural nonpoint source pollution control in Wisconsin: the limits of a voluntary program? *Water Resources Bulletin* 31:1009–1022.

Wood, P. J., and P. D. Armitage. 1997. Biological effects of fine sediment in the lotic environment. *Environmental Management* 21:203–217.

Zaimes, G. N., R. C. Schultz, and T. M. Isenhart. 2004. Stream bank erosion adjacent to riparian forest buffers, row-crop fields, and continuously-grazed pastures along Bear Creek in central Iowa. *Journal of Soil and Water Conservation* 59:19–27.

# 6. Planning and Implementation Tools That Can Be Used at the Local Level

How conservation systems and riparian buffers are implemented in response to NR 151 and other programs is a task that will ultimately fall to local conservation staff. County land conservation staff and their local Natural Resources Conservation Service (NRCS) counterparts will need to work with individual farmers and land managers to help them understand what kinds of practices are appropriate, where these are needed, and what is involved in implementation. Because of the variety of agricultural production systems and the complexity of landscapes, conservationists' work will be considerably easier with tools that help assess the situation and convey information in a readily understandable form.

Tools to assist in conservation decisions range from complex computer-based models to simple look-up tables or charts. Regardless of the sophistication, they should share these several common properties:

- Be based on sound science and be representative of local situations.
- Produce understandable and unbiased information.
- Incorporate local knowledge and build on local expertise and experience.
- Be compatible with ongoing conservation and nutrient management efforts.
- Be useable by field staff without extensive training in modeling, GIS, etc.

The WBI developed and evaluated tools to assist in the implementation process. The first tool is actually a decision-making rubric (Figure 6.1) for reviewing the output of information from the watershed ranking process. The second tool includes two approaches for identifying vulnerable areas within a selected watershed, i.e., narrowing the area that will be subjected to more detailed analyses including soil testing and the Soil Nutrient Application Planning (SNAP)-Plus computer program. The evaluation of digital elevation models (DEM), and their role in analyses at multiple scales, is explained in a sidebar on page 31. The last part of this chapter summarizes the development and validation of the SNAP-Plus tool and describes how it can be used to evaluate potential nutrient and sediment losses from individual fields.

# Local Screening of Statewide Analysis

Chapter 5 describes a process for ranking the suitability of watersheds across Wisconsin in terms of their potential responsiveness to conservation systems and riparian buffers. Because this is a statewide analysis, it relies on somewhat coarse data. Thus, the analysis does not have the specificity to identify individual farm fields and surrounding land cover or small reaches of streams. It is also based on available data, which may be out-of-date in some areas. Because of these limitations, we recommend that local knowledge be used in the selection of watersheds participating in any state-supported buffer initiative. In Figure 6.1, this is the diamond called *Incorporate local knowledge*.

From the statewide analysis, a selected number of watersheds will be recommended for participation. Local conservation staff in selected areas will be provided the opportunity to review and accept or reject these recommendations on the basis of local knowledge and conditions. Criteria for local review include evaluating the statewide results in light of more detailed or up-to-date local data, existence of other conservation programs and local activities that would influence or interact with a buffer program, likelihood of significant engagement in a program

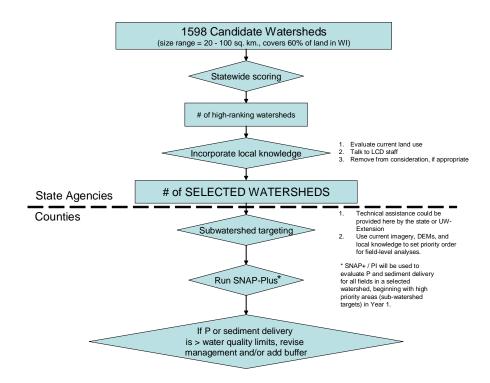


Figure 6.1. Process for evaluating watersheds, subwatershed areas, and fields.

by stakeholders, and knowledge of local conditions that may influence the likelihood of success of a buffer program. Questions related to these criteria can be put into an evaluation form to provide the basis for dialog between local and state staff about the final selection of watersheds.

### Subwatershed Area Priorities

Once a watershed has been selected for participation in a buffer program, it will be necessary to set priorities for further analysis and implementation. Although the watersheds are relatively small (20–100 square kilometers [7.7–38.6 square miles]), they still typically contain hundreds of farm fields, making it impossible in a single year to do detailed soil testing and SNAP-Plus analyses on every field.

Watershed-wide approaches based on the Universal Soil Loss Equation (USLE) can be used to identify vulnerable areas within watersheds—areas that should receive more immediate attention because they are likely to be contributing disproportionately to water quality impairment. This analysis could be done by GIS staff at a state agency as a technical support function or by local conservation staff.

Numerous watershed-scale models of erosion and nonpoint source pollution exist. Most start with a model of upland erosion based on the USLE or variants. Some models also include delivery, flow routing, and other evaluations of how much and how fast nutrients and sediments end up in receiving bodies of water. In all cases, data about

topography, soils, and cropping practices are important factors in detecting areas most likely to be contributing to water quality impairment.

For the purpose of initial identification of areas containing fields where more definitive evaluation should be conducted, it does not matter which model is used as long as the model uses a reliable representation of topography, soils, and cropping practices. (For example, as discussed in Chapter 8, the original, simple version of the USLE was chosen for the WBI pilot study.) With 30 meter-resolution digital elevation models (DEMs), county soils data (obtainable from the Soil Survey Geographic [SSURGO] Database from the US NRCS), and up-todate land cover data, a GIS-based analysis can show areas where biophysical conditions are likely to be conducive to excessively high erosion rates if appropriate practices are not in place. Individual hotspots can be aggregated to target areas, either by visual inspection of products or through GIS algorithms that identify clusters. An example of this type of analysis is illustrated in Figure 6.2.

Local conservation staff should be most familiar with local conditions and data sets. Some land conservation departments have GIS capabilities in-house or work closely with other county departments that have staff expertise in this area. The preferred option would be for these counties to do the local screening (as discussed earlier in this chapter) followed by a watershed-scale USLE-type analysis to identify vulnerable areas and set priorities for more detailed evaluation. If the time or resources for this do not exist, the analysis could be done by a state agency or UW staff and provided to the county as a map or GIS database.

# Topographic Data from Digital Elevation Models

A potential weak link in watershed-scale analysis is topographic data. The only data available throughout the state are 30-meter digital elevation models, a term that refers to data that contain a spot elevation at each point in a grid with 30-meter horizontal spacing and a vertical accuracy specified as plus or minus six feet. Through our case study and evaluation on Discovery Farms sites, we have concluded that these products are adequate for initial screening to indicate high priority areas at a watershed scale in most areas, although in areas with flat or gently undulating terrain more precise information is necessary. Our site-scale research indicates that in all but steep and regular terrain, 30-meter DEMs do not provide reliable information for field-level modeling (e.g., SNAP-Plus) or for determining the locations of concentrated flow with enough accuracy to be of assistance in buffer design and layout. For the SNAP-Plus software, slope information derived from other sources such as soils or topographic maps or from in-field observations will be necessary.

Fine resolution DEMs are necessary for delineation of flow paths and would be quite useful in determining exactly where buffers would be most effective. Unfortunately, production of DEMs with vertical accuracy in the two-foot range may be prohibitively expensive if done for a single purpose such as agricultural land management. However, Wisconsin counties are increasingly investing in such products for multiple purposes, and these products will be increasingly (though spottily) available in the future. At the scale of an individual field, it is also possible to create very accurate elevation data from GPS observations. Receivers can be mounted on four-wheelers and rapidly acquired, although processing the data to generate useful information is technically complex.

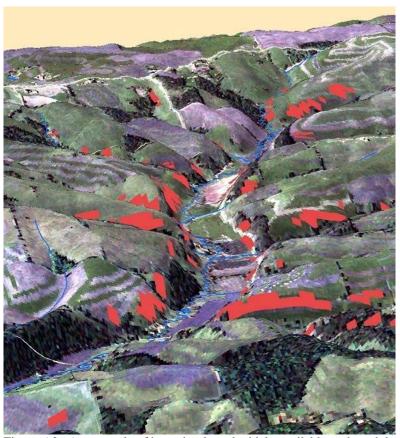


Figure 6.2. An example of how simple and widely available tools and data can be used to show areas (in red) where biophysical conditions are likely to be conducive to excessively high erosion rates if appropriate practices are not in place. (Note: Topography exaggerated for illustration purposes.)

### Evaluation of Fields In and Around Vulnerable Areas

Once an area of concern has been established in a watershed, field-level targeting is required to determine where management changes are needed to reduce sediment and nutrient losses. If changes are needed, then appropriate field-specific management options to reduce losses must be found. One tool that is available to assist this process is the SNAP-Plus software program.

SNAP-Plus was originally designed to allow agricultural producers and consultants to prepare nutrient management plans that meet the requirements of the Wisconsin NRCS Field Office Technical Guide (FOTG) Nutrient Management Standard 590 (Wisconsin NRCS 2005). A nutrient management plan indicates the rate, timing, and method of application of crop nutrients, both manure and fertilizer, to a field. The 590 standard requires that producers prepare a nutrient management plan following guidelines intended to protect groundwater and/or surface water. As a result of the limits on soil erosion and runoff phosphorus losses mandated by the 590 standard, the SNAP-Plus software includes both the current national level NRCS erosion calculation program—Revised Universal Soil Loss Equation 2 (RUSLE2)—and an agricultural phosphorus (P) runoff risk estimator—the P Index.

RUSLE2 uses crop management information and readily available soil and topographic data to produce field-specific estimates of erosion in tons per acre per year. For soil conservation planning, the estimated erosion rate is required to be below the tolerable soil loss (T). Values for T have been established by the NRCS for all mapped soils to indicate how much soil can erode from a field without degrading its ability to continue to produce crops. In RUSLE2, a field is considered to be a series of homogeneous planes, each with a specified slope, length, and management. Unlike previous NRCS erosion estimation tools, it can account for the effect of within-field deposition in addition to erosion, so it can provide an edge-of-field sediment delivery estimate. Using RUSLE2, it is possible to assess the effects of changes in type and direction of tillage operations, crop rotations, manure applications, in-field grass buffer strips, and edge-of-field filter strips on sediment delivery to the edge of a field.

The P Index uses routine cropland soil tests and other information to estimate the risk of phosphorus delivery to surface water from a given field with specified management and fertilizer and manure applications. It estimates an annual edge-of-field phosphorus loss taking into account RUSLE2 sediment delivery, rainfall and snowmelt runoff volumes, soil characteristics, soil phosphorus concentration, and manure and fertilizer phosphorus additions. This edge-of-field phosphorus loss is then multiplied by a total phosphorus delivery factor that accounts for the proportion of phosphorus leaving the field in runoff that is actually transported to a stream. The equations used to calculate this factor assume that runoff leaving the field travels to the nearest stream, pond, or lake in a concentrated flow channel, such as a grassed waterway, ditch, or gully.

The P Index in SNAP-Plus was designed specifically for Wisconsin conditions using results from laboratory and field experiments on Wisconsin soils. It can be used to evaluate the effect of varying field management practices on phosphorus delivery. That the P Index can be used to indicate the relative effects of field conditions and

management practices on phosphorus loss risks has been verified through infield runoff monitoring. Annual edge-offield P Index values correspond well to annual measured runoff phosphorus loads from cropped fields throughout Wisconsin with a range of crops, field characteristics, slopes, tillage types, and manure and fertilizer application practices (Figure 6.3).

The SNAP-Plus program's capabilities to assess field-level sediment and phosphorus runoff potential using locally available information make it a suitable tool for identifying fields in which management changes are needed to reduce runoff losses. Ideally, the maximum allowable sediment delivery or P Index values will be set at levels that address watershed

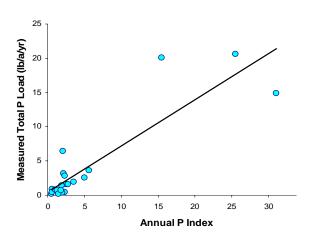


Figure 6.3. The relationship between annual edge-of-field P Index values and measured annual runoff phosphorus losses from 18 cropped fields.

water quality goals and will be modifiable through the adaptive management process. We anticipate that, initially, these maximum levels will correspond to those mandated by the Wisconsin NRCS Nutrient Management Standard 590. Updated guidelines can then be developed on a watershed basis if deemed necessary.

For fields exceeding the established levels, site-appropriate management options will be evaluated with SNAP-Plus to determine their effectiveness in reducing estimated losses below the specified levels. Additional capabilities are being added to SNAP-Plus to allow it to assess the effectiveness of the full suite of management options. It already accounts for those conservation practices that prevent erosion and increase infiltration in the field, such as reduced tillage. In-field and edge-of-field grass buffer strips are being added as management options. The P Index's field-to-stream total phosphorus delivery factor calculations are also being modified to account for other buffers between the edge of the field and the riparian area. The results of in-field monitoring for the WBI project are being used as the basis for developing equations to estimate the effectiveness of grass filter strips for removing phosphorus, as well as sediment, from runoff.

### Using Buffers to Reduce Sediment and Phosphorus Losses

Research for the WBI has helped to identify the conditions where buffers are appropriate management options and where they may not be as beneficial. We monitored parts of two cropped fields with 45-foot grass buffers. Runoff was collected above and below the buffer year-round for two years and analyzed for sediment, dissolved phosphorus, and total phosphorus. The first year of monitoring included extremely large spring and early summer storms at both sites; in the second year there was comparatively

little spring and summer runoff, with most of the runoff occurring during the winter. Field A had a sandy loam soil on a steep (10%) slope and was in corn silage for most of the study period. Field B had a clay loam soil on a half as steep (5%) slope and was in corn for grain for most of the study period. Both sites were chisel-plowed in the fall.

Soil phosphorus concentrations were lower on field B than field A. Table 6.1 shows that field A lost more sediment and phosphorus than field B in both years, which

Table 6.1. Annual Sediment and Phosphorus Runoff Losses From Two Cropped Fields Both With and Without a Buffer.

	Sediment (T/a)	Dissolved P (lb/a)	Total P (lb/a)
Field A			
Year 1			
No buffer	11.3	1.3	20.1
With buffer	3.5	0.7	17.1
Year 2			
No buffer	0.5	0.9	1.3
With buffer	0.1	0.2	0.4
Total			
No buffer	11.8	2.2	21.4
With buffer	3.6	0.9	17.5
Field B Year 1			
No buffer	0.7	< 0.1	0.5
With buffer	< 0.1	< 0.1	0.2
Year 2			
No buffer	< 0.1	< 0.1	< 0.1
With buffer	0.1	0.1	0.3
Total			
No buffer	0.7	< 0.1	0.5
With buffer	0.2	0.1	0.5

Abbreviations: P = phosphorus, T = tons

was expected due to field conditions that were more conducive to erosion, along with higher soil phosphorus. The buffer on field A greatly reduced per acre sediment loads; although even with a buffer, the losses were very high, and soil loss exceeded T for that soil. In contrast, runoff losses from field B were low—the buffer was established here for research rather than management purposes. Nonetheless, in year one, the buffer in field B did reduce runoff sediment and phosphorus loads. In the winter of year two, however, there was more runoff from the buffer than from the field itself. On the field, snowmelt was held in depressions resulting from chisel-plowing the clayey soil, while it ran off the comparatively smooth surface of the unplowed buffer, which may have been partially frozen. Consequently, this resulted in more phosphorus and sediment loss from the buffer than the field. In year two, the buffer of field B became a source, rather than a sink, for phosphorus. Over the two-year period, the buffer on field A captured 69% of the sediment but only 18% of phosphorus, while the buffer on field B captured 72% of sediment but only 7% of phosphorus.

SNAP-Plus will be useful for comparing the potential effectiveness of buffers to that of other management options for reducing runoff phosphorus and sediment loads. It cannot, however, be used to design the buffer (i.e., determine the width required at any point) because both RUSLE2 and the P Index assume that fields are homogenous planes without any in-field concentration or channelization of runoff flow. If a buffer is chosen as the preferred management option, its placement and design must take into account runoff flow patterns within the landscape as described in Chapter 7.

#### Reference

Wisconsin NRCS. 2005. Field Office Technical Guide Conservation Practice Standard Nutrient Management Code 590. Available at:

http://efotg.nrcs.usda.gov/treemenuFS.aspx?Fips=55025&MenuName=menuWI.zip.

# 7. Placement and Design of Conservation Systems and Riparian Buffers

A buffer is just one component of the entire management system that will be required to meet the state's water quality objectives. Upland practices such as minimal tillage, residue cover, cover crops, and terraces must be in place in order to reduce sediment and water flow to a level that can be handled by the buffer in an environmentally friendly manner. We know that a vegetated buffer's effectiveness depends on the vegetation density and the buffer width. Erosion research (Jin and Römkins 2001) indicates that standing stem density is one of the main measures of the effectiveness of the buffer—many stiff stems slow the water flow, which results in sediment deposition conditions. Current Wisconsin NRCS recommended grass-based buffer mixes have sufficient stem density to produce an effective buffer. For a given vegetation density, the width of the buffer determines the amount of sediment that can be removed without overwhelming, or "blowing out," the buffer (Magette et al. 1989). Therefore, it is important to design the buffer width based on the volume of water that will flow through the buffer cross-section at any point along the riparian buffer.

The design of riparian buffers was one of the main questions posed at the initiation of the WBI. As described in earlier chapters, the need for a riparian buffer is dependent upon a conservation systems approach that includes the ranking of watersheds on their probability of responding to three specific water quality goals, the identification of areas within those watersheds that have the greatest probability of needing conservation systems and riparian buffers, and a decision-making process to determine if upland management changes will mitigate the need for riparian buffers in these areas. If it has been determined that a riparian buffer is necessary, the WBI recommends that, rather than using a uniform width as is currently the recommendation according to Wisconsin NRCS guidelines, the buffer be designed relative to the contributing area.

The science of designing buffers relative to the contributing area has been validated in the emerging scientific area of precision conservation (Delgado et al. 2005). In their article "Establishing Conservation Buffers Using Precision Information," Dosskey et al. (2005) document the importance of considering the topography in designing a riparian buffer (Figure 7.1). They note that "runoff [is] commonly nonuniform, converging on some parts of the field margin and diverging from others because of uneven topography and patterns of soil conditions and farming practices" (Dosskey et al. 2005, 349).

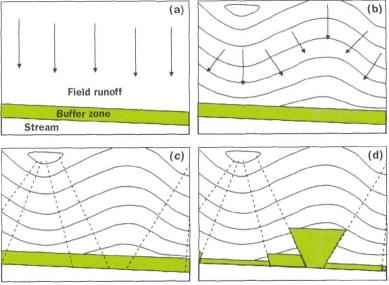


Figure 7.1 Diagrams of crop-field runoff patterns, topographic contours, and alternative buffer designs: (a) uniform runoff flow to a uniform-width buffer; (b) non-uniform runoff flow to a uniform-width buffer, (c) non-uniform runoff areas and the corresponding uniform-width buffer areas to which they flow; (d) non-uniform runoff areas and the corresponding variable-width buffer areas to which they flow. Both (a) and (d) yield an approximately constant level of pollutant filtering along the entire length of the buffer. (Dosskey et al. 2005)

# **Contributing Area**

Relative water-flow volume can be estimated using contributing area calculations and estimates. The contributing area is the area of land from which runoff will flow to any common point. An example of the application of this concept to buffer design follows: if 20% of a ten-acre field drains through a thirty-foot section of field edge while only 5% of the field drains through the adjacent thirty-foot section of field edge, the section of buffer handling the larger contributing area should be wider than the section that handles the smaller contributing area (all else being equal).

When management decisions for agricultural fields are made, it is common to consider the field to be a plane with a single slope or a sequence of planes with one or two slope breaks. In this sort of geometry, each point at the bottom of the field has the same contributing area and the same runoff. However, natural landscapes in Wisconsin are rarely like this. Variations in topography perpendicular to the average slope cause areas of runoff convergence and divergence, resulting in large differences in contributing areas along the bottom of the field. Analyses of several DEMs with contributing area algorithms and a distributed runoff model (Precision Agricultural-Landscape Modeling System or PALMS [Molling et al. 2005]) show a factor of 10 to a factor of over 100 difference between the amount of runoff that would be expected to flow through different sections of an edge-of-field buffer on fields in Wisconsin.

The WBI recommends a modification to the current Wisconsin NRCS Filter Strip Standard 393 (Wisconsin NRCS 2001) that includes contributing area. Current Wisconsin NRCS Standard 393 does not consider contributing area when calculating the width of buffers; however other factors, such as slope and soil characteristics, are part of the design. The proposed modification would not change the current recommendations for buffer dimensions for a field that is a uniform plane with no convergence. Only when

convergence occurs would the shape of the buffer change with the modified code. Our recommended changes are in Table 7.1. To incorporate contributing area into buffer design, a unit contributing area factor (UCAF) is used. First the field is assigned a point score according to Table 1 in the Wisconsin NRCS Standard 393 (Wisconsin NRCS 2001). Then the UCAF corresponding to the 393 point score is selected from Table 7.1. The contributing area (CA in square feet/foot) at each location along the edge of the field is then divided by the UCAF to produce the buffer width. For example, if the number of points from the Standard 393 Table 1 is 35 and the contributing area at the edge of the field is 240 square feet /foot from TauDEM (an extension for the ArcGIS computer program), then the buffer-width parallel to the direction of flow should be 40 feet, i.e. CA/UCAF(points = 35) = 240/6 = 40 feet. (See Figure 7.2 for an example of this process computed along an entire field boundary.) If estimated buffer widths exceed 300 feet, then buffers are not likely to be appropriate in those situations without modifications to the landscape; too much contributing area is emptying into too small an area to be effectively aided by a buffer. If most of the runoff is leaving a field as concentrated flow, which is frequently the case in Wisconsin, then the buffer should be integrated with grassed waterways or other structures that are designed to retain maximum sediment.

In our experience with the calculation of contributing area, we have found two software packages that have performed consistently well (TauDEM and Autodesk Land

Desktop for

AutoCAD) and two software packages that have produced inconsistent results (TOPAZ and the ARC INFO Flow Accumulation Function). To calculate the points in Table 7.1, GIS can be used to determine the average slopes along

Unit Contributing Area **Total Point Range From** Current 393 Standard Current 393 Standard Buffer Width (ft)<sup>a</sup> Factor (UCAF) 0-10 20 15 15-20 10 30 25–30 40 7.5

Table 7.1. Modified Table to Determine Minimum Buffer Width.

35 50 6 40 5 60 45 4.3 70 50 80 3.8 >50 100 3

<sup>a</sup>The buffer width from the current Wisc. NCRS Standard 393 is shown for reference.

the incoming flow direction.

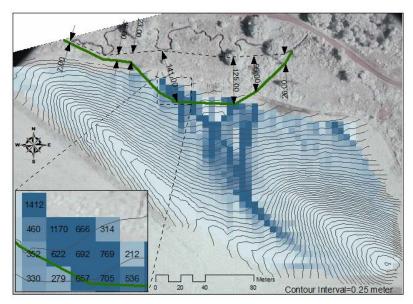


Figure 7.2. A variable-width buffer in a Wisconsin field that was evaluated for appropriate buffer width taking contributing area into account. The background of the figure is an aerial photo of the site, in which a stream appears as a meandering dark line along the top of the image. A natural buffer exists along the stream north of the field. Topographic contour lines (in black) are overlaid on the figure in areas in which a digital elevation model (DEM) was available. The grid in shades of blue is the contributing area; light blue denotes small contributing areas, and dark blue denotes large contributing areas. The contributing area was calculated with TauDEM. The inset shows the actual contributing area values in square feet/foot for a small portion of the DEM. Contributing area values were divided by a unit contributing area factor (UCAF) of 10, which corresponds to the Wisc. NRCS Standard 393 Table 1 point range (15–20) for this location. The green line denotes the minimum buffer width as calculated using contributing area and the UCAF. The buffer was measured from the smoothed stream bank line (dotted line). The numbers on the arrows are the buffer widths (feet) along the direction of overland flow. As this figure shows, the natural buffer is sufficiently wide in most places, so additional buffer would only be required in the center.

We recognize that the minimum buffer widths calculated from Table 7.1 will be very wide in some areas and narrow in other areas nearby. This may cause difficulty for equipment operations, so some smoothing or squaring off will probably need to be done. How this is done is entirely up to the landowner or land manager, although the resulting buffer width may not be narrower than the value calculated from the tables.

#### References

Delgado, J. A., C. A. Cox, F. J. Pierce, and M. G. Dosskey. 2005. Precision conservation in North America. *Journal of Soil and Water Conservation* 60:340.

Dosskey, M. G., D. E. Eisenhauer, and M. J. Helmers. 2005. Establishing conservation buffers using precision information. *Journal of Soil and Water Conservation* 60:349–353.

Jin, C.-X., and M. J. M. Römkens. 2001. Experimental studies of factors in determining sediment trapping in vegetative filter strips. *Transactions of the ASAE* 44: 277–288.

Magette, W. L., R. B. Brinsfield, R. E. Palmer, and J. D. Wood. 1989. Nutrient and sediment removal by vegetated filter strips. *Transactions of the ASAE* 32: 663–667.

Molling, C. C., J. C. Strikwerda, J. M. Norman, C. A. Rodgers, R. Wayne, C. L. S. Morgan, G. R. Diak, and J. R. Mecikalski.In press. Distributed runoff formulation designed for a precision agricultural-landscape modeling system. *Journal of the American Water Resources Association*.

Wisconsin NRCS. 2001. Field Office Technical Guide. Section IV Conservation Practices. Filter Strip (393). Available at: http://efotg.nrcs.usda.gov/treemenuFS.aspx?Fips=55025&MenuName=menuWI.zip.

# 8. Pilot Study of the Proposed WBI Implementation Processes

A pilot study of the proposed WBI implementation process was conducted in two watersheds—Hefty Creek in Green County and the Sheboygan River watershed in Fond Du Lac County—during the spring and summer of 2005. Following watershed identification, we followed the implementation steps shown in Figure 6.1: subwatershed targeting, field-scale sediment loss and phosphorus-delivery potential assessment in vulnerable areas using SNAP-Plus, and finally, identifying management alternatives to reduce phosphorus delivery risks from fields exceeding the established limits.

### Step 1: Subwatershed Area Targeting

In early May 2005, we met with the Land Conservation Departments of both Green and Fond Du Lac Counties to discuss the usefulness of combining the GIS tool with local knowledge to determine where conservation and buffer practices should be implemented. We explained that the tool is designed to provide a quick, meaningful, and objective evaluation of the landscape within a selected watershed that would assist local staff with their outreach activities. The results of the analysis would then be used as a guide, supplementing—but not replacing—the extensive knowledge that local conservation staff possess.

The GIS-based subwatershed targeting tool uses the USLE to estimate soil erosion within the watershed (Desmet and Govers 1996, Fernandez et. al. 2003) The data required for the tool, which are available to the public free-of-charge, include: 30-meter DEMs, SSURGO Database, Wisconsin Initiative for Statewide Cooperation on Landscape Analysis and Data (WISCLAND), and USGS runoff estimates.

Analysis using the GIS-based subwatershed tool was conducted for both watersheds and yielded results that were, for the most part, consistent with information provided by the local staff. In general, the tool showed that agricultural areas near streams with steep slopes and erodible soils were most likely to lose soil to surface waters. However, local staff suggested that erosion from within the channel was not adequately represented with this tool. This should be taken into consideration when evaluating the entire watershed.

Time constraints dictated that only one watershed could be looked at in great detail, so Green County was chosen as the primary study area because of its closer proximity to UW-Madison. By late May, the Land Conservation Department of Green County had chosen two subregions of the Hefty Creek watershed to be of the most interest. Combined, these areas are 15% of the watershed and include all or part of six farms.

# Step 2: Identifying High Risk Fields

First, we collected the information needed to run SNAP-Plus to produce rotation average soil loss estimates and P Index values for all fields. The team met with each of the farmers in the study area. Farmers were asked about their current manure management, fertilizer applications, tillage practices, crop rotations, typical yields, herd

sizes, and field names and acreages. Soil and landscape information was either observed in the field or obtained from NRCS soil survey maps provided by the county. All cropped or grazed fields were sampled according to the standard procedure for routine agronomic soil tests. We followed the UW-Extension recommendation of at least one sample per field and no more than 5 acres per sample. Within the pilot area, 274 routine soil samples (six-inch depth) for phosphorus, potassium, pH and organic matter were taken on 1019 acres. This represents an average of 3.7 acres per sample.

The Hefty Creek watershed pilot areas are characterized by very steep slopes and highly erodible soils. The northwestern pilot area has steeply sloping fields (9% and greater) located within 300 feet of all stream stretches. About a quarter mile of the stream in this area runs through grazed pasture. In the southeastern study area, the fields adjacent to the stream are comparatively broad lowland areas with slopes less than 4%. In this area, fields adjacent to the

### Time and Costs Associated With Snap-Plus

Collecting the information and running SNAP-Plus requires time and resources. For areas similar to the pilot study area, we estimate that sampling a 200-acre farm would take forty hours. The actual data entry and analysis in SNAP-Plus for this farm would require roughly five hours, but field layouts in the study area are so complicated, with as many as 100 or more fields per farm, that matching the information with the field names used by the farmer will make the process take an additional twenty-five hours. The average cost of routine soil samples is \$7.00.

stream are planted in corn and soybean rotations. Throughout both study areas, all of the fields with slopes of 4% or greater are farmed on the contour and many are in contour strips. Crop rotations on these upland fields were six to eight years with two to three years of row crops and three to four years of established alfalfa hay. Most of the upland fields are hydrologically connected to the lowland areas adjacent to the stream via grassed waterways or tree-lined ravines mapped as intermittent streams.

Soil test P values on the southeastern study area lowland corn and soybean fields ranged from 46 ppm to 344 ppm with the majority above 100 ppm. These values are all "excessively high" meaning that they are well above levels where additional phosphorus is required for crop growth. In contrast, the soil test P range for the upland soils with rotations that included alfalfa hay was 9 ppm to 103 ppm, with the majority of fields below 50 ppm. Although currently little to no manure is applied to the lowland fields, their high soil test P values likely reflect high rates of past manure applications to the flatter, relatively accessible land. The drop in manure applications is a result of a dramatic drop in animal numbers in this area within the last decade.

The current Wisconsin NRCS Nutrient Management Standard 590 target maximum for rotation-average P Index values is 6. Overall, of the 973 cropped acres examined, 19% had P Index values greater than 6, 21% were between 4 and 6, 45% were between 2 and 4, and 15% were less than 2. Erosion and movement of sediment-bound P was the greatest contributor to P loss estimates; all of the fields with P Index values above 6 were upland fields and, except for 1.8 acres, also had estimated soil loss values above the NRCS designated tolerable soil loss value or T. We should note that the percentage (19%) of cropped acres with high P Index values reported above should be considered a maximum. This is because one of the farmers with fields with high P Index values resulting from high soil loss estimates told us that the farm varied between using

some tillage and using no-till in parts of the rotations. Where this variation occurred, we used the more erosive management with tillage in our analysis. The majority of the cropped fields adjacent to the creek had P Index values lower than 4 (with one exception of 4.8) despite having comparatively high soil test P values. This is a result of the very low risk of runoff and erosion from these fields. Of the 31 acres of nonrotationally grazed pasture within the two study areas, all had P Index values greater than 6. Fifteen acres were grazed on a monthly rotation, of which 6 acres had P Index values higher than 6.

### Step 3: Identifying Management Alternatives

After our initial analysis to identify fields with unacceptable soil loss or P Index values, we reran SNAP-Plus for those fields to evaluate a variety of alternative managements. We found that the P Index values for the cropped fields could always be brought below 6 through adjustments in rotation, tillage, or, on a few fields, shifting the timing of manure applications from winter to spring. In most cases, the necessary adjustments could be made using rotations and managements that the farmers were already using on other fields on their farm. The exceptions were fields that were spring chisel plowed and had with two years of corn silage in the rotation. Switching to no-till would bring the risk of soil loss and phosphorus delivery from almost all of these fields to acceptable levels (Figure 8.1). Some fields had acceptable P Index values but had estimated soil loss levels in excess of T. Figure 8.2 shows an example of a tract with one field that could be brought below T by removing soybeans from the rotation on that field.

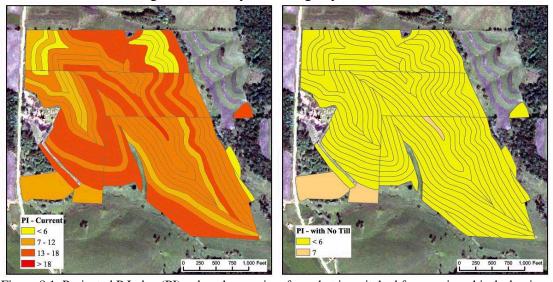


Figure 8.1. Projected P Index (PI) value changes in a farm that is switched from spring chisel-plowing to no-till. The map on the left shows field PI values using the present tillage system, and the one on the right shows that most of the PI values would be brought to acceptable levels (below 6) if a no-till system was adopted without any other management changes.



Figure 8.2 Tract with one field that is losing soil at a rate that exceeds the tolerable soil loss (T), shown in orange on the left. Planting corn instead of soybeans in the rotation would reduce the soil loss to below T (right).

Prior to discussing reducing the P Index values on the pasture lands, we must note that the P Index validation with runoff monitoring data described in Chapter 6 and shown in Figure 6.3 was on cropland. At present, we have no data verifying that the P Index algorithms are appropriate for grazed lands, although some monitoring projects are underway in Wisconsin that should allow us to address this in the future. Most of the pasture in the pilot study area is on very steep and highly erodible land not well-suited to cropland. Using the present P Index algorithms, we found that an ungrazed grass filter strip between the grazed field and waterways appears to be an appropriate way to reduce P Index values below 6. Within the study area, there are also unvegetated paddocks that receive high rates of manure. Since the surface of these paddocks is a permanent covering of manure, not soil, the P Index is not a suitable tool to estimate paddock runoff P loss risks. However, the paddocks are hydrologically connected to the stream via a grassed waterway and probably have a high potential to contribute dissolved phosphorus, if not particulate phosphorus, to the stream.

#### References

Desmet, P. J. J., and G. Govers. 1996. A GIS procedure for automatically calculating the USLE LS factor on topographically complex landscape units. *Journal of Soil and Water Conservation* 51:427–433.

Fernandez, C., J. Q. Wu, D. K. McCool, and C. O. Stöckle. 2003. Estimating water erosion and sediment yield with GIS, RUSLE, and SEDD. *Journal of Soil and Water Conservation* 58:128–136.

# 9. Assessing the Economic Impacts of Alternative Management Practices on Selected Wisconsin Farms

This chapter summarizes the results of an economic evaluation of alternative management practices for fields with a high risk of excessive phosphorus and sediment loss. Because the WBI recommendation uses a conservation systems approach to address the contributing area relative to a riparian area, upland management may be required to adopt various remedial practices or best management practices (BMPs). Whether BMPs alone or BMPs in conjunction with a riparian buffer will be required will be determined as part of the assessment process by local conservation staff. Of the initial four questions developed by the WBI Advisory Committee, two addressed this area: what will be required to get conservation and buffer practices in place, and what are the consequences of implementing conservation and buffer practices? Both questions have economic dimensions to them.

In the longer term, the objectives of this economic research are to develop analytical tools (e.g., EXCEL spreadsheets) for whole-farm economic optimization subject to nutrient balance and environmental constraints, such as the NRCS 590 standard. These analytical tools will provide a framework to better assess economic and environmental performance trade-offs over a range of BMPs.

With respect to the WBI project, the specific objectives are to assess the economic impacts of farm/field level adjustments that will meet the NRCS 590 standard. NRCS 590 is used because it is the current standard, but this analysis could be performed using alternative thresholds. While the farm/field adjustments presented here are not the optimal adjustments in that they are not necessarily the profit maximizing and/or cost minimizing adjustments required to meet the 590 standard, they do provide a measure of the cost of compliance. Comparison of alternative BMPs (e.g., tillage practices, changing corn grain for corn silage in a noncompliant field) provides a basis for better understanding the potential economic costs and environmental benefits of alternative management practices on Wisconsin farms.

#### Simulating Field/Farm Level Conditions for Analysis

In our research approach, representative farms in WBI-targeted watersheds were simulated under alternative farm/field level conditions using SNAP-Plus. The conditions included tillage practices (spring/fall moldboard versus chisel versus no-till) and crop rotations (switching out corn silage for corn grain, alfalfa/brome versus alfalfa). The economic costs of alternative farm/field cropping activities were assessed using standardized crop budgets for corn grain, corn silage, soybeans, alfalfa, wheat, oats, and other small grains. These basic data were obtained from the average production costs/acre observed on several Discovery Farms. These data provide reasonable, ballpark cost of production estimates without revealing the actual costs of production for particular farms. It must be noted, however, that the actual cost of compliance will likely be very situation specific.

The economic assumptions used in the analyses are summarized in Table 9.1. Costs of production/acre were computed for all crops and tillage methods using standard machinery and labor costs and custom rates for particular field operations. Changing tillage, crop rotation, field-level nutrient management, feed/rations (which may change Table 9.1. Wisconsin Buffer Initiative Economic Assumptions Cost by Crop and Tillage.

							5 Yr	
							Avg	
	Fall	Fall			Spring	Spring	FSA	Unit of
Crop	Chisel	Plow	None	No-Till	Chisel	Plow	Prices	Measure
Alfalfa (Hay/Haylage)	\$319.00	\$319.00	\$0.00	\$319.00	\$319.00	\$319.00	\$74.00	Ton
Corn Grain	\$330.64	\$338.80	\$0.00	\$311.36	\$332.63	\$340.80	\$2.12	Bushel
Corn Silage	\$409.28	\$417.44	\$0.00	\$390.00	\$411.27	\$419.44	\$19.08	Ton
Oats w/ Alfalfa Seeding								
Spring	\$214.28	\$222.44	\$0.00	\$195.00	\$216.27	\$224.44	\$1.38	Bushel
Oats w/ Alfalfa/Brome								
Seeding Spring	\$214.28	\$222.44	\$0.00	\$195.00	\$216.27	\$224.44	\$1.38	Bushel
Oatlage w/ Alfalfa								
Seeding Spring	\$214.28	\$222.44	\$0.00	\$195.00	\$216.27	\$224.44	\$48.10	Ton
Pasture (not rotational),								
Grass	\$0.00	\$0.00	\$0.00	\$0.00	\$0.00	\$0.00	\$48.10	Ton
Pasture Rotational,								
Grass	\$0.00	\$0.00	\$0.00	\$0.00	\$0.00	\$0.00	\$48.10	Ton
Pasture (not rotational),								
unimproved	\$0.00	\$0.00	\$0.00	\$0.00	\$0.00	\$0.00	\$48.10	Ton
Soybeans	\$249.58	\$269.05	\$0.00	\$241.61	\$249.58	\$271.05	\$5.35	Bushel
Winter Wheat	\$252.55	\$260.72	\$0.00	\$233.28	\$254.55	\$262.71	\$3.50	Bushel

Abbreviation: FSA = Farm Service Agency

manure composition), among other BMPs, to meet the 590 standard have the potential to change cost of production.

While there is an active debate as to the yield implication of alternative tillage practices, this analysis assumes that crop yields are identical across tillage practices (although the analytical tools allow for this assumption to be changed). In particular, spring or fall chisel plowing is assumed to generate the same yields as no-till. In contrast, changing tillage method or timing can have large impacts on soil loss and the P Index. Crop yields were obtained from participating farms as basic input data to the Snap Plus program. All crops are valued at Farm Service Agency five-year average commodity prices (see Table 9.1).

#### Effectiveness of Alternative BMPs

Comparing the reductions in the P Index with the associated changes in economic costs provides a partial budgeting tool that may be used by landowners to assess the cost effectiveness of alternative BMPs. Although particular fields are the focus of the Snap-Plus 590 standard compliance simulations, the analysis tracks changes in crop acreage, yields, production, costs, revenues, and profits across *all* fields over the full rotation of the farm being analyzed. This provides an economic assessment that parallels the 590 standard computation of the P Index over the full rotation.

Some BMPs will be more effective (i.e., generate more P Index reductions per dollar of profit and/or cost) compared with others. However, some of these more effective BMPs may not be compatible with the farm's current cropping and/or livestock system. For example, substituting corn grain for corn silage may leave a dairy operation with an on-farm forage deficit, thereby necessitating buying forage. Therefore, in order to better match local conditions and specific farm constraints, it may be necessary to explore a variety of BMPs and their P Index reduction effectiveness when designing 590 standard compliant cropping and livestock systems.

# Analysis of Three Southwest and South Central Wisconsin Farms

In all analyses, the BASE simulation is defined as existing management.

#### Case 1: WBI-1

This farm uses a dominant rotation comprised of three years alfalfa (plus an establishment year with oatlage followed by two years of corn silage for a six-year rotation on about 140 acres. The dominant tillage is spring chisel. There are multiple fields with winter spreading restrictions under the NRCS 590 standard, due mainly to steep slopes and nearness to Hefty Creek.

The BASE simulation (see Appendix C, Tables1, 2 and 3) for this farm reveals that every field but one is noncompliant with the NRCS 590 standard (i.e., P Index > 6) over the full rotation. The whole-farm acreage weighted-average P Index for the BASE simulation of this farm is 16.6.

Introducing no-till on all fields, while keeping the BASE rotation and winter spreading, generates 590 standard compliance in all but four fields. The whole-farm acreage weighted-average P Index for this simulation is 3.8, a radical reduction in overall average farm P Index due to a change in tillage from spring chisel to no-till. Additional elimination of winter spreading on these four fields generates full 590 standard compliance on all fields over the full six-year rotation. The whole-farm acreage weighted-average P Index for this simulation is 3.5. It should be noted, however, that the elimination of winter spreading on this farm would necessitate additional winter manure-storage strategies. The cost of these strategies is not addressed here.

Changing the BASE rotation to two years of corn grain rather than corn silage, while keeping tillage and winter spreading the same as in the BASE simulation, reduces the P Index in most fields compared to the BASE, but only fourteen fields are 590 compliant. The whole-farm acreage weighted-average P Index for this simulation is 7.1. Introducing no-till while maintaining BASE simulation winter spreading makes all fields 590 compliant and reduces the whole-farm acreage weighted-average P Index to 2.8. Additionally, eliminating winter spreading further reduces the whole-farm acreage weighted-average P Index to 2.5.

These results indicate that three of the simulations would enable this farm to meet 590 P Index standards on all fields: corn silage, no-till, no winter spreading; corn grain, no-till, winter spreading; and corn grain, no-till, no winter spreading. Determining the "best" way to meet the 590 standard is likely to be somewhat subjective because of certain practical crop and livestock considerations and economic and environmental trade-offs. For example, corn grain, as modeled here, leaves more residue than corn silage and generates bigger P Index reductions on this farm's hilly ground, particularly when

combined with no-till; however, corn silage better meets the livestock/forage needs of the farm. The economic and environmental trade-offs involved with 590 standard compliance are summarized in Appendix C, Table 3 as the change in profits (i.e., total revenues minus total costs, aggregated to the farm level by field by year, over the six-year rotation) for each BMP scenario compared with the BASE simulation. Since yields were assumed to be constant across tillages, the profit gains due to no-till are mainly due to reduced crop costs/acre. In the case of this farm, no-till corn silage with no winter spreading generates the largest economic return per unit of P Index reduction, whereas no-till corn grain with no winter spreading generates the largest P Index reduction. If managed using no-till corn silage with no winter spreading, this farm could improve profitability by \$7,359 over the six-year rotation while becoming 590 standard compliant. Profitability would increase by only \$354 with the implementation of no-till corn grain with no winter spreading. Here, the economics reinforce the livestock and forage benefits of no-till corn silage with no winter spreading. Again, however, the potential costs of additional winter manure-storage strategies required by these 590 standard compliant simulations are not addressed here.

#### Case 2: On Farmer's Ground (OFG)-14

This farm uses eight different rotations and two tillages (fall chisel and no-till) on 600 plus acres. The longest rotation is six years. The BASE simulation (see Appendix C, Tables 4 and 5) indicates that one 6-acre field (less than 1% of the total farm acreage) is noncompliant with the NRCS 590 standard with a P Index of 10.6. The whole-farm acreage weighted-average P Index for the BASE simulation is 2.8. The noncompliant field has a 7% slope and is in a fall chisel, corn silage/no-till, soybean rotation. Simply swapping this field from corn silage to corn grain with a similar sized field that would go from corn grain to corn silage at a similar point in the rotation generates 590 standard compliance with a field level P Index of 4.5, a 6.1 unit decrease from the BASE simulation. The P Index for the other field involved in this corn silage/corn grain crop swap increases from 2.7 to 3.9, hence remains 590 compliant. Due to the differences in soils and yields between these two fields, this generates \$1,172 in profits over the six years of the full farm rotation. The whole-farm acreage weighted-average P Index for this full 590 standard compliance simulation is 2.8, which is a reduction of 0.1 from the BASE simulation.

Another alternative for this noncompliant field is switching the corn silage tillage from fall chisel to no-till, which would reduce tillage costs and, assuming identical yields, increase profits by \$347 over the full six-year farm rotation. Of course, this assumes labor and machinery or timely custom hire are available for this shift in tillage practice. In this simulation, the field-level P Index decreases to 4.0, while the whole-farm acreage weighted-average P Index for the simulation is 2.8, an identical P Index to the corn silage/corn grain rotation simulation. The economics in this case suggest that changing the crop rotation would be a better way to go as it generates more farm profits (\$1,172 versus \$347 over six years) and a bigger return (\$3.13 versus \$0.85) per unit of P Index reduction.

#### Case 3: OFG-16

This farm uses four different seven-year rotations on about 190 acres, all with three years of alfalfa or alfalfa/brome and an establishment year of oats or oatlage. Three of these rotations use three years of continuous corn grain following the alfalfa, and the

fourth uses three years of corn silage. Fall chisel is the only tillage practice used. Ten of the twenty-seven fields (32 acres or 16 % of the total farm acreage) are in grazed pasture. Appendix C, Tables 6 and 7 summarize the SNAP-Plus simulations for this farm.

Four of the twenty-seven fields are noncompliant in the BASE simulation. Two of these fields are in crop rotations with P Indexes of 7.1 and 9.8, and two are pastures with P Indexes of 8.5 and 6.4. The two noncompliant crop fields, totaling 15 acres or 8% of the total farm acreage, are in the Oa-A-A-A-Csl-Csl-Csl rotation with fall chisel tillage. The noncompliant pastures total 7 acres or 3.6% of the total farm acreage and about 22% of total pasture acreage. The whole-farm acreage weighted-average P Index for this BASE simulation is 3.5.

Two obvious choices for reducing the P Index in the noncompliant crop fields are to switch to corn grain from corn silage or to switch to no-till corn silage. Switching to corn grain will leave more crop residue, hence reducing the P Index, but could generate livestock forage shortages in several years. This switch reduces field level P Indexes from 7.1 and 9.8 to 3.8 and 3.0, respectively. In addition, the whole-farm acreage weighted-average P Index for this partially compliant—the pasture fields are still not addressed—is 3.1, a reduction of 0.4 units of P Index. The economic analysis indicates that this change will generate an extra \$5,256 in profits over the seven-year full-farm rotation.

Addressing the two noncompliant pastures with rotational grazing is assumed to increase harvested forage yield. This increase forage yield, with no assumed increases in costs generates an additional \$4,700 over the seven-year full-farm rotation. (Pro-rated fixed costs for paddock establishment and variable labor costs to rotate the cattle are not computed here.) This translates to around \$35 per pastured acre per year increase in farm profits due to rotational grazing. This switch also reduces field-level P Indexes from 8.5 and 6.4 to 5.6 and 2.9, respectively. The whole-farm acreage weighted-average P Index for this farm simulation is 3.0, a reduction of 0.4 units of P Index.

# Summary

This research provides a snapshot on the economics of meeting NRCS 590 standard compliance for three farms in southwest and south central Wisconsin. Although the particular economics of 590 standard compliance are likely to be highly situation specific due to the local environmental, agronomic, and farm specific (e.g., machinery, labor, management) constraints, these analyses provide a ballpark economic assessment using standardized crop production costs and Farm Service Agency five-year average commodity prices.

In general, for the farms studied here, adoption of NRCS 590 standards would require straightforward management changes on noncompliant fields such as tillage practices (e.g., from chisel to no-till), rotation changes (substituting corn grain for corn silage), and the adoption of managed rotational versus continuous grazing. In many instances, these 590 standard compliance-induced changes were found to be profitable under the economic assumptions used for the analysis. However, several potential costs, such as additional winter manure-storage strategies and the costs of rotational grazing, were not included in this analysis.

In addition, it should be noted that individual farms may be unable to adapt alternative tillage techniques for a variety of reasons. For example, the costs of crop and

forage production may be quite different from those assumed here, or perhaps custom hire is more expensive than assumed here or is not available in a timely fashion. Furthermore, while small changes in rotations like swapping corn grain for corn silage on similar sized fields were generally not found to be problematic, these types of changes should be modeled in conjunction with the farm's livestock feed/ forage needs where applicable.

Finally, the economics of 590 standard compliance evaluated here are not optimized. That is, the simulations are not the result of maximizing farm profits subject to field-level environmental constraints. Viewed from a whole farm, nutrient balance, cropping and livestock systems perspective, this economic and environmental optimization context could provide profitable opportunities for rethinking farm management. This systems perspective allows for a better integration and synergy of component farm enterprises: cash grain/forage crops versus on-farm use versus purchased feeds/forage; planting and/or harvesting operations provided on-farm versus custom hire; nutrient management plans to minimize off-farm purchases and maximize returns to on-farm nutrient sources (e.g., manures and legumes); better feed/ration management to better control off-farm nutrient sources; and better management of the number and types of livestock, their feed/forage needs, and their manure volume and composition. Rethinking farm management in this economic and environmental optimization context will require analytical tools to facilitate and quantify the relevant trade-offs. Building on the tools developed for the WBI, further research will be focused in this direction.

## 10. Recommendations of the WBI Advisory Committee

Before listing the individual recommendations, it is important to address the context of the WBI within larger resource management issues. The status and level of understanding of the science underlying the natural resource concerns in the state of Wisconsin is constantly improving. This ever-emerging new knowledge requires us to constantly reexamine the adequacy of yesterday's recommendations for tomorrow's changing policy expectations. The collaboration between UW-Madison and the state and federal agencies, as well as all the other participants in the WBI process, was productive while establishing a framework for future efforts. All involved hope there will be opportunities in the near future when there will be a call for the "best available science" that will facilitate UW-Madison scientists to work with the citizens of Wisconsin in addressing our natural resource management challenges.

Wisconsin does not exist in a natural resource policy vacuum. Landowners, local staff, state agencies, and federal partners are all responding to policy changes made at various levels of government. The WBI cannot be expected to anticipate all the complexities occurring in the natural resource management arena.

Perhaps the most important recommendation emerging from the WBI is the need for an adaptive management approach. Implementation of our recommendations in WBI watersheds should be viewed as opportunities for learning. Adaptive management is designed to foster continual improvement in management practices. Surprises in ecosystem response are not viewed as failures but, instead, as a source for learning better ways of accomplishing water quality goals.

Adaptive management promises more than improved effectiveness and efficiency in resource management. It offers a new paradigm for designing and implementing resource management programs while accelerating our understanding of how to most effectively solve resource management problems.

# Areas of Agreement

The WBI was asked to conduct research to determine where riparian buffers are needed to enhance water quality in Wisconsin relative to agricultural runoff. We began by agreeing that riparian buffers by themselves would not allow us to achieve our water quality goals. Instead, a conservation systems perspective is recommended in which riparian buffers are one potential component in this system. Viewing riparian buffers as an integral part of a larger conservation system, however, does not address the "where" question.

There was agreement that the implementation of these conservation systems should occur first in areas where there is the greatest probability of a positive water quality response. To locate these areas, WBI scientists identified and ranked 1598 watersheds in Wisconsin based on criteria agreed to by the entire WBI Advisory Committee. The WBI makes no recommendation on how far down this ranked list any resulting implementation efforts should occur. We recognize that this is a funding decision that needs to be agreed upon by elected officials in the Wisconsin Legislature and agency administrators.

There was also general agreement among WBI participants that simply identifying the boundaries of a set of watersheds would not be sufficient to achieve our

water quality goals. The identification of high-priority areas within those watersheds is also important, and spatial analytical tools and widely available digital data, may be used to provide an initial assessment of those subwatershed priority areas. WBI participants also agreed that local staff should have the ability to revise subwatershed priority areas based on additional data or experience. For areas designated as priorities within targeted watersheds, the SNAP-Plus tool can be used to help landowners assess alternative management options for a conservation system. Following this conservation systems approach may result in situations where changes in upland management practices may reduce the need for a riparian buffer.

#### Recommendations

The following recommendations have to be interpreted in the adaptive management context. Removing the recommendations from this context lessens their value and defeats the very essence of the adaptive management philosophy. Moreover, the recommendations will not solve all our current water quality problems. However, they do focus limited resources on those problems that are causing a disproportionate share of degradation. They also focus efforts on those situations that have the highest probability of responding to remedial efforts. There was general agreement by the participants in the WBI Advisory Committee on the following recommendations.

- 1. It is recommended that the DNR and UW have the lead responsibility for conducting the data acquisition and analysis necessary to establish and maintain a list of WBI watersheds based on the three agreed upon criteria (1) improve stream water quality, (2) protect and restore aquatic biological communities, and (3) sustain lake water quality), and make this information available to appropriate staffs of the DNR and DATCP on an on-going basis. This data acquisition and analysis activity will also include identifying areas within the WBI watersheds that are especially vulnerable based on soils, topography, land use, and any water quality data available. This activity is to be conducted based on a memorandum of agreement between the DNR and UW-Madison.
- 2. It is recommended that WBI watersheds receive special consideration for new state funding based on a tiered approach where funds are allocated from the highest ranked to the lowest ranked, with the number funded at any one time based on fiscal considerations. It is also recommended that conservation agencies and organizations in Wisconsin coordinate existing programs to address high-ranked WBI watersheds.
- 3. A WBI watershed targeted for remediation will be informed of this fact by a letter from the DNR to the appropriate Land Conservation Committee(s) (LCC). The DNR will provide the county with a preliminary map or list of fields/locations that should be subject to priority implementation of the nonpoint and buffer standards. The LCC will review this information in conjunction with their Land and Water Resource Plan and any additional data the county may have to determine which agricultural fields should receive priority treatment based on the estimation that these fields are most likely to yield the greatest water quality benefit in the watershed. Once the nonpoint issues are addressed on the priority fields, the LCC will engage landowners to address other fields that may need appropriate treatment to achieve water quality objectives. The LCC may also

review other nonpoint pollution sources within the watershed. Where these additional sources are found to substantially contribute to water quality problems within the selected watershed, the LCC will work with the DNR and other agencies to address these problems as well. The LCC coordination with other agencies will be especially important when federal programs are targeted as major funding sources.

- 4. When working on implementation with a landowner in a vulnerable area within a funded WBI watershed, LCCs should formulate a plan based on a conservation systems approach. Each field's contributing area will be identified, and the US Department of Agriculture NRCS conservation planning model will guide the process of determining appropriate upland practices and riparian buffer options. The expectation is that conservation and appropriate management practices will be installed in this contributing area to reduce the impact of concentrated flow areas and runoff of nutrients and sediments moving to the riparian area. If riparian buffers are required, then these will be designed to specifically address the upland contributing area (see recommendation #7). Where sheet and rill erosion are the cause of water quality impairment, appropriate buffers will be installed to achieve water quality goals. As part of this process, it is expected that local conservation professionals will use the US Department of Agriculture NRCS conservation planning model as a basis for making decisions to prohibit agricultural activities from encroaching on the stream. Each LCC will determine if local conservation staff has or should have the capability to employ the SNAP-Plus model. The SNAP-Plus model may be used to determine management options for assessing remedial practices within the contributing area. This determination would be based on the phosphorous (P) Index or soil erosion values, which are components of the SNAP-Plus model. The DNR will work with the UW to ensure that the SNAP-Plus model remains consistent with state nonpoint performance standards and that training and updates to this planning tool occur on a regular basis for those that chose to use it.
- 5. Local staff will be responsible for coordinating the monitoring of targeted watersheds and reporting those results to the DNR in accordance with a predetermined process. The DNR will work with UW and other salient agencies or organizations to coordinate monitoring efforts in targeted watersheds to the extent feasible (see Appendix A). The DNR is also encouraged to work with UW and any other salient organizations or agencies in interpreting these monitoring results in order to determine what type of changes or adaptation, if any, is needed in the funded WBI watersheds.
- 6. It is recommended that the need for riparian buffers in any targeted location be determined using a two-step process. First, using the SNAP-Plus model, the landowner will be encouraged to adopt various management practices as part of a larger conservation system that results in meeting existing soil and water conservation standards. If the efficacy of the various conservation and management practices is not capable of meeting these standards, then the second step will be to compliment these upland treatments with a riparian buffer. This riparian buffer will be engineered based on elevation contours so as to specifically address the upland contributing area. If a landowner wants to install a riparian buffer without making any changes to upland

practices, and the buffer is designed to be sufficient to meet the standard as determined by SNAP-Plus calculations, then a buffer alone should be permissible.

- 7. It is recommended that the US Department of Agriculture NRCS buffer standards be updated to incorporate the knowledge gained through the WBI and Discovery Farm research on buffers in Wisconsin over the past three years. In particular, the NRCS standards should recognize contributing drainage area, in-field soil erosion rates and variations in buffer designs and landscape conditions.
- 8. It is recommended that the senior staff in the DNR and DATCP work with the Wisconsin Legislature to investigate the type of institutional arrangements that would be required to allow adaptive management to be the foundation of our natural resource management programs.

# **Appendix A**

## Recommendations for Monitoring the Effects of WBI Implementation on Nonpoint-Source Pollution Levels

In order to collect the data required for the adaptive management process, it will be necessary to allocate adequate financial resources for the purpose of monitoring nonpoint-source pollution levels in Wisconsin. The following recommendations have been developed by the WBI-affiliated scientists for monitoring progress towards the three WBI management goals (1) improve stream water quality, (2) protect and restore aquatic biological communities, and (3) sustain lake water quality. All of these recommendations should complement the monitoring of the actual program implementation.

- 1. Monitoring requires a long-term commitment of *at least* ten years in selected watersheds.
- 2. The monitoring strategy proposed here is a tiered approach consisting of three levels. These monitoring options were developed based upon the experience of WBI scientists working on these types of projects; however, each situation needs to be independently evaluated and the monitoring may need to be adjusted on the basis of site-specific conditions.
  - Level 1: A small number of watersheds should be monitored with continuousflow gages and flow-rate-metered automated samplers located at the
    watershed outlets. Scheduled biweekly and storm-event driven water
    sampling should occur at these locations. In addition, event-driven
    samples of soil and nutrient movement should be collected at field
    outlets or edges where management changes are made. Annual
    monitoring of stream biota should also be conducted in these watersheds
    above and below fields of interest. In the drainage basins of these Level
    1 sites, detailed, field-level land cover and management inventories that
    include cropland and non-cropland should be conducted throughout the
    monitoring period. Observations on cropland should include crops,
    tillage, fertilizer, and manure management practices.
  - Level 2: A moderate number of watersheds should be monitored with biweekly water sampling, including continuous stream flow monitoring at the outlet of the watershed. Biological monitoring should also be conducted in these watersheds. After an initial monitoring period, a comparative analysis will be conducted between the Level 2 and Level 1 sites to evaluate the ability of the Level 2 sampling method to detect water quality changes resulting from changes in known upstream land management. Initial estimates suggest that the cost of implementation and operation of the Level 2 sites is approximately one-fourth the cost of a Level 1 site over a ten-year period.
  - Level 3: In all other selected watersheds, biological monitoring should be conducted. This can be accomplished by seasonal staff and/or citizens volunteering through the Water Action Volunteers program.

- 3. Paired watersheds should be monitored where opportunities exist. Reference basins can be monitored with the Level 2 monitoring and volunteers can defray the costs. Reference gages will be located near the gages for a selected watershed, making the additional travel time needed to monitor its paired watershed minimal.
- 4. Baseline monitoring should begin as soon as possible. We recommend installing water-quality sampling stations and sampling the biota in the top ranked groups of watersheds. To the extent feasible, existing monitoring programs and one-time stream studies should be incorporated into the information base.
- 5. The Wisconsin DNR should collaborate with the UW system to carry out this monitoring and develop procedures for data integration, analysis, and dissemination, with the goal of creating an open information resource relevant to the ongoing implementation and funding of state nonpoint-source pollution rules.

We encourage the careful monitoring of management changes within the watershed, landowner participation rates, and the time necessary for staff to complete analyses recommended by the WBI process. This information can be collected through surveying landowners and local staff in addition to the use of available remotely-sensed imagery. We also encourage using multiple approaches to program implementation where feasible, as long as these approaches are capable of generating information about the efficacy of nonpoint-source control measures. For example, one variation might be the use of an auction to allocate program dollars within a subwatershed. Finally, we recommend incorporating the research conducted at the Discovery Farms and Pioneer Farm when program changes are considered. All of this information should be open and easily accessible through Web-based resources.

# **Appendix B**

Appendix B: Ranked list of WBI watersheds. In the "ID" field, the number indicates the group rank and the letter indicates the goal(s) in which that watershed ranked the highest (W, improve stream water quality; F, protect and restore aquatic biological communities; C, composite of W and F; L, sustain lake water quality). The "Description" field contains the name of the stream at the outlet of the watershed (Unnamed streams were named "Trib. to [first named stream downstream]"). The "County" field is the county the contains the outlet of the watershed (Note that many watersheds cross county boundaries). The "Area" field is the area of the watershed in square kilometers.

<u>ID</u>	Description	County	Area (sq. km.)
1-W	Tiffany Creek	Dunn	89
1-F	Trib. to Lemonweir River	Juneau	44
1-C	Willow River	St Croix	96
1-L	Brownlee Creek	Buffalo	48
2-W	Mineral Point Branch	Iowa	86
2-F	Vermillion River	Barron	65
2-C	Eau Galle River	St Croix	99
2-L	Mill Creek	Iowa	94
3-W	Lovett Creek	Lafayette	21
3-F	Leech Creek	Columbia	60
3-C	Ames Branch	Lafayette	92
3-L	Upper Pine Creek	Barron	81
4-W	Dodge Branch	Iowa	95
4-F	Campbell Creek	Adams	72
4-C	Dry Run	St Croix	74
4-L	Silver Creek	Fond du Lac	72
5-W	East Branch Pecatonica River	Iowa	65
5-F	Otter Creek	Sauk	96
5-C	South Fork Willow River	St Croix	95
5-L	South Fork Bad Axe River	Vernon	60
6-W	Furnace Creek	Lafayette	54
6-F	Onemile Creek	Juneau	81
6-C	Isabelle Creek	Pierce	99
6-L	West Fork Kickapoo River	Vernon	98
7-W	Conley Lewis Creek	Iowa	38
7-F	East Branch Big Eau Plein River	Marathon	56
7-C	Pecatonica River	Iowa	99
7-L	Arnold Creek	Clark	33
8-W	Cave Creek	Pierce	57
8-F	Trib. to Wisconsin River	Columbia	44
8-C	Rattlesnake Creek	Grant	90
8-L	Jackson Creek	Walworth	47
9-W	Eagle Creek	Buffalo	80
9-F	Koshkonong Creek	Dane	98
9-C	Otter Creek	Lafayette	79
9-L	Halls Creek	Jackson	70
9-L	South Fork Halls Creek	Jackson	35
9-L	East Fork Halls Creek	Jackson	71

<u>ID</u> 10-W	<u>Description</u> Sinsinawa River	<u>County</u> Grant	Area (sq. km.)
10-VV 10-F	Barker Creek	Barron	44
10-1 10-C	Little Platte River	Grant	86
10-C	Trib. to Silver Creek	Fond du Lac	31
10-L 11-W	Wind River	Pierce	24
11-vv 11-F	Koshkonong Creek	Dane	26
11-C	North Fork Bad Axe River	Vernon	85
11-C	Mill Creek	Monroe	66
11-L 11-L	Brandy Creek	Monroe	23
11-L 12-W	•	Dunn	60
12-vv 12-F	Cady Creek	Walworth	86
	North Branch Nippersink Creek		
12-C	Yellowstone River	Lafayette	75 05
12-L	Big Rib River	Taylor	85
13-W	Silver Creek	Barron	20
13-F	Scotch Creek	Marathon	69
13-C	Shullsburg Branch	Lafayette	85
13-L	Wedges Creek	Clark	88
13-L	Meadows Creek	Clark	31
14-W	Lost Creek	Pierce	67
14-F	Freeman Creek	Marathon	71
14-C	Galena River	Lafayette	88
14-L	Roy Creek	Green Lake	40
15-W	Parker Creek	St Croix	39
15-F	Steel Brook	Jefferson	71
15-C	Rush River	St Croix	98
15-L	Trib. to Fox River	Waukesha	90
16-W	Little Trimbelle Creek	Pierce	51
16-F	Cawley Creek	Clark	99
16-C	Fennimore Fork	Grant	61
16-L	Beaver Creek	Juneau	56
17-W	South Fork Elk Creek	Buffalo	66
17-F	Kinnickinnic River	St Croix	99
17-C	Sudah Branch	Iowa	89
17-L	Little Yellow River	Juneau	99
18-W	Otter Creek	Iowa	50
18-F	West Branch Rock River	Fond du Lac	80
18-C	Blake Fork	Grant	89
18-L	West Branch Beaver Creek	Jackson	51
18-L	West Branch Beaver Creek	Jackson	21
19-W	Plum Creek	Pierce	84
19-F	Turtle Creek	Walworth	81
19-C	Little Baraboo River	Sauk	66
19-L	Sand Creek	Burnett	90
20-W	Smock Creek	Green	20
20-F	Mill Creek	Wood	95
20-C	Spafford Creek	Lafayette	59
20-L	Root River	Milwaukee	99
21-W	Kittleson Valley Creek	Iowa	85
21-F	Hay River	Barron	86
21-C	Moccasin Creek	Wood	75

<u><b>ID</b></u> 21-L	<u>Description</u> North Branch Duck Creek	<u>County</u> Columbia	Area (sq. km.)
21-L	Middle Branch Duck Creek	Columbia	59
22-W	Trib. to Willow River	St Croix	28
22-F	South Branch Beaver Creek	Marinette	78
22-C	Bonner Branch	Lafayette	89
22-L	Ashippun River	Waukesha	91
23-C	Apple River	Lafayette	28
23-W	Big Rock Branch	Grant	26
23-F	Door Creek	Dane	81
23-L	Trib. to Yellow River	Barron	21
23-L	Boyer Creek	Washburn	33
23-L	Trib. to Boyer Creek	Washburn	24
24-W	Jordan Creek	Green	42
24-F	South Branch Yellow River	Wood	93
24-C	Little Richard Creek	Green	73
24-L	Spring Creek	Columbia	46
25-W	Copper Creek	Lafayette	31
25-F	North Fork Popple River	Clark	99
25-C	Blue River	Grant	86
25-L	West Branch Milwaukee River	Fond du Lac	85
26-W	Trib. to Otter Creek	Lafayette	25
26-F	Duck Creek	Adams	98
26-C	Hamann Creek	Marathon	68
26-L	Robinson Creek	Jackson	83
27-W	Smith Conley Creek	lowa	49
27-W	Hay Creek	Eau Claire	87
27-C	Sylvester Creek	Green	64
27-L	Left Foot Creek	Marinette	42
28-W	McClintock Creek	Lafayette	22
28-F	Sixmile Creek	Dane	86
28-C	Seas Branch	Vernon	36
28-L	Wood River	Burnett	60
29-W	Trib. to Plum Creek	Pierce	27
29-W	Pheasant Branch	Dane	61
29-C	Skinner Creek	Green	77
29-L	Harder Creek	Polk	27
30-W	Flint Creek	lowa	77
30-W	Big Drywood Creek	Chippewa	96
30-C	Sixmile Branch	Grant	64
30-L	Cedar Creek	Washington	70
31-W	Pine Creek	Pierce	70 41
31-W	Spring Brook	Langlade	81
31-C	Mounds Branch	Grant	44
31-L	Rock Creek	Polk	28
	Trempealeau River	Buffalo	40
32-W	Blackhawk Creek	Rock	
32-F			91 85
32-C	South Fork Hay River	Dunn	85
32-L	Tarr Creek	Monroe	58
32-L	Tarr Creek	Monroe	100
32-L	Silver Creek	Monroe	82

<u>ID</u> 33-C	<u>Description</u> Artus Creek	<u>County</u> Marathon	Area (sq. km.)
33-L	Washington Creek	Taylor	34
34-W	Fennimore Fork	Grant	29
34-F	Cunningham Creek	Clark	99
34-C	Narrows Creek	Sauk	84
34-L	Oconomowoc River	Washington	76
35-W	Stevens Creek	Rock	37
35-F	Lightning Creek	Barron	55
35-C	Blockhouse Creek	Grant	94
35-L	Sand Creek	Monroe	55
36-W	Kinnickinnic River	Pierce	27
36-F	Koshkonong Creek	Dane	73
36-C	Wolf Creek	Lafayette	74
36-L	Bashaw Brook	Burnett	93
37-W	Rush River	Pierce	28
37-F	Rubicon River	Dodge	85
37-C	Springville Branch Bad Axe River	Vernon	57
37-L	North Branch Little River	Oconto	69
38-W	Little Plum Creek	Pepin	26
38-F	South Branch Rock River	Fond du Lac	100
38-C	Neshota River	Brown	87
38-L	Mukwonago River	Waukesha	29
39-W	Beaver Creek	Dunn	46
39-F	Little Tamarack Creek	Trempealeau	37
39-C	Silver Creek	Dodge	59
39-L	Hartman Creek	Waupaca	21
40-W	Mud Creek	Winnebago	66
40-F	Turtle Creek	Barron	95
40-C	Big River	Pierce	54
40-L	Glenn Creek	Jackson	42
41-W	Reads Creek	Vernon	50
41-F	Spring Brook	Rock	50
41-C	Boice Creek	Grant	94
41-L	Black River	Taylor	92
42-W	Kittleson Valley Creek	Iowa	86
42-F	Wolf River	Chippewa	98
42-C	East River	Brown	98
42-L	Trib. to Yellow River	Wood	98
43-W	Bear Creek	Vernon	62
43-F	Mill Creek	Dodge	58
43-C	Cazenovia Branch	Sauk	87
43-L	Silver Creek	Washington	24
44-W	Camp Creek	Richland	42
44-F	Pine Creek	Taylor	96
44-C	Moore Creek	Monroe	94
44-L	Beaver Creek	Taylor	44
45-W	Waumandee Creek	Buffalo	52
45-F	North Branch O'Neill Creek	Clark	91

<u>ID</u> 45-C	<u>Description</u> Jack Creek	<u>County</u> Clark	Area (sq. km.)
45-L	Tyler Forks	Iron	100
46-L	Trib to Morrison Creek	Jackson	22
46-W	Traverse Valley Creek	Trempealeau	58
46-F	South Fork Eau Claire River	Clark	98
46-C	Whiteside Creek	Lafayette	52
47-W	Spring Creek	Pierce	21
47-F	Meadows Creek	Barron	84
47-C	Rock River	Rock	37
47-L	Rock Creek	Jefferson	61
48-W	Trib. to Trempealeau River	Trempealeau	22
48-F	Rocky Run	Columbia	84
48-C	Soda Creek	Marathon	23
48-L	Waupee Creek	Oconto	66
49-W	Neshonoc Creek	La Crosse	35
49-F	Plum Creek	Dodge	61
49-C	Platte River	Grant	98
49-L	Marengo River	Bayfield	92
49-L	Morgan Creek	Bayfield	32
50-W	Trib. to Mississippi River	Pierce	25
50-F	Dill Creek	Marathon	96
50-C	Bishop Branch	Vernon	52
50-L	Deerskin River	Vilas	75
51-W	Trib. to Trempealeau River	Trempealeau	21
51-F	Trib. to West Branch Rock River	Dodge	55
51-C	Madden Branch	Lafayette	58
51-L	Horse Creek	Polk	91
52-W	Porcupine Creek	Pepin	30
52-F	Jambo Creek	Manitowoc	56
52-C	Lotz Creek	Chippewa	21
52-L	Black Otter Creek	Outagamie	55
53-W	Hackett Branch	Grant	25
53-F	Bear Creek	Portage	51
53-C	Spring Creek	Green	45
53-L	Schoenick Creek	Shawano	38
54-W	Elk Creek	Buffalo	46
54-F	West Branch Big Eau Pleine River	Marathon	89
54-C	Pigeon Creek	Grant	57
54-L	Clear Creek	Jackson	65
55-W	Kickapoo River	Monroe	97
55-F	Rocky Run	Marathon	41
55-C	Tenmile Creek	St Croix	55
55-L	Somo River	Lincoln	71
55-L	Trib. to South Fork Jump River	Price	23
55-L	South Fork Jump River	Price	46
56-W	Crooked Creek	Grant	45
56-F	Bear Creek	Marathon	94

<u>ID</u> 56-C	<u>Description</u> Dougherty Creek	<u>County</u> Lafayette	Area (sq. km.)
56-L	Perry Creek	Jackson	52
57-W	Vance Creek	Dunn	42
57-F	Sugar Creek	Walworth	71
57-C	Little Hemlock Creek	Wood	49
57-L	Spirit Creek	Burnett	65
57-L	Trib. to Wood River	Burnett	20
58-W	Pine Creek	Trempealeau	28
58-F	Randall Creek	Marathon	81
58-C	Rountree Branch	Grant	36
58-L	East Branch Milwaukee River	Fond du Lac	87
59-W	Trib. to Fennimore Fork	Grant	41
59-F	Honey Creek	Green	86
59-C	Trib. to Mississippi River	Pierce	34
59-L	South Branch Little Wolf	Waupaca	72
60-W	Mill Creek	Richland	72
60-F	Yahara River	Dane	93
60-C	Searles Creek	Green	52
60-L	White River	Waushara	49
60-L	White River	Waushara	48
61-W	Trib. to Coon Creek	Vernon	32
61-F	North Branch Tenmile Creek	Portage	85
61-C	West Branch Sugar River	Dane	85
61-L	Trib. to Morrison Creek	Jackson	22
62-W	Barr Creek	Sheboygan	24
62-F	Taylor Creek	Rock	80
62-C	Wood Branch	Lafayette	48
62-L	North Branch Little Wolf	Waupaca	89
63-W	Trib. to Trempealeau River	Buffalo	26
63-F	Des Plaines River	Kenosha	55
63-C	Rocky Run	Clark	38
63-L	South Branch Pigeon River	Waupaca	98
63-L	North Branch Pigeon River	Waupaca	65
64-W	East Branch Pecatonica River	Lafayette	22
64-F	Spring Creek	Columbia	92
64-C	Big Green River	Grant	89
64-L	Minnow Creek	Ashland	27
65-W	Trib. to Waumandee Creek	Buffalo	37
65-F	East Twin River	Kewaunee	84
65-C	Tainter Creek	Crawford	99
65-L	Starks Creek	Oneida	26
66-W	East Branch Mill Creek	Richland	26
66-F	Dorn Creek	Dane	32
66-C	Trib. to Fourteenmile Creek	Adams	28
66-L	Brant Creek	Lincoln	41
66-L	Little Somo River	Lincoln	100
67-W	Trib. to Rush Creek	Crawford	36

<u>ID</u> 67-F	<u>Description</u> East Branch Rock River	County Washington	Area (sq. km.)
67-C	Little Rib River	Marathon	57
67-L	Whitewater Creek	Walworth	62
68-W	North Creek	Trempealeau	28
68-F	Mullet River	Sheboygan	83
68-C	No Name Creek	Washington	42
68-L	Montreal River	Iron	51
69-W	Richland Creek	Crawford	66
69-F	Drewek Creek	Marathon	22
69-C	South Branch O'Neill Creek	Clark	61
69-L	Linzy Creek	Oconto	59
70-W	Trib. to Tainter Creek	Crawford	29
70-F	Token Creek	Dane	71
70-C	Hornby Creek	Vernon	53
70-L	Little Peshtigo River	Marinette	80
71-W	Sand Creek	Crawford	25
71-F	Nelson Creek	Clark	72
71-C	Trimbelle River	Pierce	92
71-L	North Fork Jump River	Price	70
71-L 72-W	West Branch Mill Creek	Richland	35
72-VV 72-F	Rock Creek	Clark	66
72-1 72-C	McAdam Branch	Grant	35
72-U	Pine Lake Creek	Oneida	87
72-L 73-W	Borst Valley Creek	Trempealeau	53
73-vv 73-F	Sevenmile Creek	Juneau	53 54
73-C	Engle Creek	Barron	25
73-C 73-L	•	Walworth	23 84
73-L 74-W	Honey Creek Van Dyne Creek		24
74-vv 74-F	Rowan Creek	Winnebago Columbia	24 77
74-C	Bower Creek	Brown	92
74-C 74-L	South Branch Yellow River	Juneau	92 79
74-L 75-W	Trib. to Milwaukee River	Ozaukee	33
75-vv 75-F	Brighton Creek	Kenosha	73
75-C	De Neveu Creek	Fond du Lac	
75-C 75-L	Hawkins Creek	Jackson	58 61
	Morrison Creek		39
75-L 76-W	Trib to Kickapoo River	Jackson Crawford	27
76-VV 76-F	Spring Brook	Walworth	41
76-F 76-C	Brush Creek	Vernon	82
76-C 76-L	Red River	Menominee	78
76-L 77-W		Buffalo	
	Trout Run Creek		21
77-F	Little Plover River	Portage	33
77-C	Menominee River	Grant	53 26
77-L 78-W	Squaw Creek	Forest	26 37
	Elk Creek	Vernon	37
78-F	Potato Creek	Marathon	30
78-C	Rogers Branch	Grant	67

<u>ID</u> 78-L	<u>Description</u> Dickey Creek	<u>County</u> Jackson	<u>Area (sq. km.)</u> 36
79-W	Sawmill Creek	Lafayette	59
79-F	Hemlock Creek	Wood	57
79-C	West Branch Little Sugar	Green	89
79-L	Lunch Creek	Marquette	58
80-W	Cherry Branch	Lafayette	24
80-F	Trib. to Cedar Creek	Washington	60
80-C	Borah Creek	Grant	45
80-L	West Branch Eau Claire River	Langlade	85
80-L	Sucker Creek	Langlade	20
81-W	Gran Grae Creek	Crawford	46
81-F	Otter Creek	Rock	99
81-C	Sugar River	Green	56
81-L	South Branch Neenah Creek	Marquette	81
82-W	Coon Creek	Vernon	21
82-F	Sugar River	Dane	94
82-C	Big Sandy Creek	Marathon	47
82-L	Jay Creek	Monroe	59
83-W	Sawyer Creek	Winnebago	37
83-F	Little Drywood Creek	Chippewa	88
83-C	OK Creek	Green	23
83-L	Little Mackay Creek	Washburn	92
84-W	Halls Branch	Crawford	34
84-F	Hoosier Creek	Racine	54
84-C	Squaw Creek	Marathon	53
84-L	Pecore Creek	Oconto	87
85-W	Ash Creek	Richland	48
85-F	Ore Creek	Walworth	49
85-C	Porky Creek	Marathon	21
85-L	Otter Creek	Marinette	84
85-L	Colburn Creek	Forest	25
86-W	Turton Creek	Trempealeau	62
86-F	Maunesha River	Dane	95
86-C	Bears Grass Creek	Eau Claire	73
86-L	Long Lake Creek	Iron	61
87-W	Halfway Creek	La Crosse	85
87-F	Fourmile Creek	Marathon	64
87-C	Rush Creek	Crawford	79
87-L	Christie Brook	Oconto	36
88-W	Trib. to Platte River	Grant	36
88-F	Little Turtle Creek	Walworth	98
88-C	Wildcat Creek	Dodge	74
88-L	Gudegast Creek	Oneida	49
88-L	Jennie Webber Creek	Oneida	64
89-W	Harvey Creek	Buffalo	95
89-F	Fenwood Creek	Marathon	53
89-C	Little Grant River	Grant	52

<u>ID</u> 89-L	<u>Description</u> Neenah Creek	County Marquette	Area (sq. km.)
90-W	Arkansaw Creek	Pepin	59
90-F	Scarboro Creek	Kewaunee	56
90-C	Bear Creek	Clark	30
90-L	Hay Creek	Washburn	35
91-W	Hardies Creek	Trempealeau	21
91-F	Kummel Creek	Dodge	79
91-C	East Branch Fond Du Lac River	Fond du Lac	65
91-L	Bear Creek	Taylor	27
92-W	Cleaver Creek	Juneau	62
92-F	Allen Creek	Rock	92
92-C	Martin Branch	Grant	57
92-L	Butternut Creek	Ashland	69
93-W	Pigeon Creek	Trempealeau	97
93-F	South Fork Lemonweir River	Monroe	93
93-C	West Branch Baraboo River	Vernon	95
93-L	Wolf Creek	Marinette	38
94-W	Warner Creek	Vernon	64
94-F	Seeley Creek	Sauk	85
94-C	Wild Creek	Marathon	28
94-L	Trib. to Wolf River	Menominee	64
95-W	Trib. to Apple Creek	Outagamie	26
95-F	Little Bear Creek	Barron	42
95-C	Bear Creek	Juneau	95
95-L	Big Weirgor Creek	Rusk	79
96-W	Pompey Pillar Creek	Iowa	51
96-F	Trib. to Red Cedar River	Barron	67
96-C	Norwegian Creek	Clark	35
96-L	Casey Creek	Washburn	84
97-W	Trib. to Mormon Creek	Vernon	40
97-F	Duck Creek	Jefferson	81
97-C	Prahl Creek	Marathon	35
97-L	Middle Inlet	Marinette	85
98-W	Bruce Valley Creek	Trempealeau	29
98-F	Shaw Brook	Dodge	91
98-C	Timber Coulee Creek	Vernon	91
98-L	Pike Lake Creek	Marathon	26
99-W	Sugar Creek	Crawford	66
99-F	Trib. to Buena Vista Creek	Portage	47
99-C	Mosquito Creek	Wood	52
99-L	Fox Creek	Polk	57
100-W	Bogus Creek	Pepin	29
100-F	Brewer Creek	Juneau	35
100-C	Hickey Creek	Barron	27
100-L	Bog Brook	Forest	27
101-W	Otter Creek	Vernon	29
101-F	Silver Creek	Manitowoc	64
			-

<u>ID</u> 101-C	<u>Description</u> Billings Creek	<u>County</u> Vernon	Area (sq. km.)
101-L	Radley Creek	Waupaca	81
101-L	Emmons Creek	Waupaca	70
102-W	Hawkins Creek	Richland	40
102-F	East Fork Hemlock Creek	Wood	33
102-C	Trib. to Sugar River	Rock	31
102-L	Hydes Creek	Waupaca	32
103-W	Trib to Chippewa River	Pepin	29
103-F	Lower Pine Creek	Dunn	94
103-C	Roxbury Creek	Dane	67
103-L	Dead Creek	Monroe	37
104-W	Trout Creek	Buffalo	31
104-F	North Branch Honey Creek	Sauk	94
104-C	Coon Branch	Lafayette	23
104-L	Eighteenmile Creek	Bayfield	81
105-W	Trib to Lake Michigan	Kewaunee	25
105-F	Noisy Creek	Marathon	33
105-C	Roger Creek	Chippewa	27
105-L	Little Deerskin River	Vilas	27
105-L	Blackjack Creek	Vilas	37
106-W	Annis Creek	Dunn	55
106-F	Black Creek	Marathon	73
106-C	Markham Creek	Rock	27
106-L	Minnesuing Creek	Douglas	56
107-W	Cook Creek	Monroe	23
107-F	Trib. to Des Plaines River	Kenosha	61
107-C	Mormon Creek	La Crosse	96
107-L	Squaw Creek	Price	63
108-W	Willow Branch	Grant	21
108-F	Cold Spring Creek	Dodge	37
108-C	East Branch Shioc River	Shawano	45
108-L	First South Branch Oconto	Menominee	82
109-W	Wilson Creek	Dunn	97
109-F	South Fork Popple River	Clark	64
109-C	West Fork Little Rib River	Marathon	61
109-L	North Branch Pemebonwon River	Marinette	92
110-W	Trib to S. Fork Bad Axe River	Vernon	24
110-F	North Fork Bob Creek	Chippewa	93
110-C	Burgy Creek	Green	65
110-L	North Fork Skinner Creek	Rusk	52
111-W	Pine Creek	Sauk	39
111-F	Marsh Creek	Rock	90
111-C	Plum Creek	Brown	94
111-L	Big Pine Creek	Lincoln	70
112-W	Little Willow Creek	Richland	36
112-F	Duncan Creek	Chippewa	90
112-C	County Line Creek	Marathon	34

<u><b>ID</b></u> 112-L	<u>Description</u> Noisy Creek	<u>County</u> Oneida	Area (sq. km.) 96
113-W	Ashwaubenon Creek	Brown	76
113-F	Trout Creek	Dunn	82
113-C	Trib. to Little Eau Pleine River	Marathon	35
113-L	Weso Creek	Oconto	29
114-W	Chimney Rock Creek	Trempealeau	62
114-F	Saunders Creek	Rock	69
114-C	Little Richard Creek	Green	25
114-L	Fay Lake Outlet	Florence	55
115-W	Weister Creek	Vernon	55
115-F	Spring Brook	Dodge	66
115-C	Knapp Creek	Richland	94
115-L	Dryden Creek	Ashland	64
116-W	McCartney Branch	Grant	20
116-F	Big Roche a Cri Creek	Adams	69
116-C	Dorrity Creek	Barron	22
116-L	Murphy Creek	Marinette	25
117-W	Melancthon Creek	Richland	38
117-F	Dead Creek	Dodge	81
117-C	Beaver Creek	Wood	28
117-L	Swamsauger Creek	Oneida	44
118-W	French Creek	Trempealeau	58
118-F	Trib. to Little Eau Pleine River	Portage	52
118-C	Thompson Valley Creek	Eau Claire	34
118-L	Little Thornapple River	Rusk	94
119-W	Davis Creek	Jackson	21
119-F	Yellow River	Wood	96
119-C	Silver Creek	Marathon	20
119-L	Yellow River	Washburn	86
120-W	Harrison Creek	Vernon	31
120-F	East Fork Popple River	Clark	53
120-C	East Branch Yellow River	Wood	41
120-L	Kelly Brook	Oconto	72
121-W	Millville Creek	Grant	61
121-F	North Fork Eau Claire River	Clark	99
121-C	Mud Creek	Dodge	36
121-L	Chippanazie Creek	Washburn	78
122-W	Elk Creek	Trempealeau	92
122-F	Tisch Mills Creek	Manitowoc	33
122-C	Coon Creek	Vernon	49
122-L	North Branch Oconto River	Forest	96
123-W	Rowley Creek	Sauk	34
123-F	Mud Creek	Jefferson	38
123-C	Apple Creek	Outagamie	93
123-L	Ericson Creek	Douglas	29
123-L	Amnicon River	Douglas	73
124-W	Willow Creek	Richland	95

<u><b>ID</b></u> 124-F	<u>Description</u> Starkweather Creek	<u>County</u> Dane	<u>Area (sq. km.)</u> 55
124-C	West Branch Pine River	Richland	99
124-L	Stevens Creek	Florence	76
125-W	Little Sugar River	Green	100
125-F	Crawfish River	Columbia	65
125-C	Devil Creek	Lincoln	71
125-L	Upper Tamarack River	Douglas	60
125-L	Toad Creek	Douglas	27
126-W	Mud Branch	Lafayette	25
126-F	Paint Creek	Chippewa	92
126-C	Sherman Creek	Eau Claire	60
126-L	Squirrel River	Oneida	88
127-W	Plum Creek	Crawford	43
127-F	Calamus Creek	Dodge	61
127-C	Trib. to Koshkonong Creek	Dane	21
127-L	Twin Lakes Creek	Oneida	54
128-W	Picatee Creek	Crawford	27
128-F	Lomira Creek	Dodge	54
128-C	Seymour Creek	Juneau	56
128-L	Beaver Brook	Washburn	63
129-W	Trib. to Waumandee Creek	Buffalo	31
129-F	North Branch Pensaukee River	Oconto	92
129-C	Beaver Creek	Eau Claire	33
129-L	Papoose Creek	Vilas	21
130-W	South Fork Kinnickinnic River	Pierce	46
130-F	Mud Creek	Dane	61
130-C	Trout Creek	Brown	40
130-L	Second South Branch Oconto River	Oconto	68
131-W	Lane Creek	Grant	48
131-F	Badger Mill Creek	Dane	87
131-C	Leggett Creek	Grant	49
131-L	Pioneer Creek	Vilas	93
132-W	Big Creek	Trempealeau	45
132-F	North Branch Milwaukee River	Sheboygan	97
132-C	Plum Creek	Sauk	39
132-L	Nixon Creek	Vilas	36
133-W	Trib. to De Neveu Creek	Winnebago	30
133-F	Muskrat Creek	Eau Claire	87
133-C	Hay Creek	Chippewa	45
133-L	Brule Creek	Forest	97
134-W	Garners Creek	Outagamie	30
134-F	Town Drain	Green Lake	40
134-C	Robbins Creek	Columbia	22
134-L	Tom Doyle Creek	Oneida	26
135-W	Lowery Creek	Iowa	35
135-F	West Branch Shioc	Shawano	77
135-C	Butler Creek	Dodge	48
		<b>-</b>	

<u>ID</u> 135-L	<u>Description</u> Brunsweiler River	<u>County</u> Ashland	Area (sq. km.)
136-W	Hutchinson Creek	Buffalo	21
136-F	Ahnapee River	Door	93
136-C	East Branch Big Sandy Creek	Marathon	26
136-L	Sailor Creek	Price	96
137-W	Trib. to South Fork Bad Axe River	Vernon	27
137-F	Oldens Creek	Marathon	48
137-C	Mount Vernon Creek	Dane	43
137-L	Ghost Creek	Sawyer	31
137-L	Christy Creek	Sawyer	27
138-W	Little Suamico River	Oconto	68
138-F	Kewaunee River	Kewaunee	85
138-C	Sterling Creek	Clark	20
138-L	Trib. to Bear River	Vilas	50
139-W	Little Waumandee Creek	Buffalo	93
139-F	North Fork Trade River	Burnett	48
139-C	Dutchman Creek	Brown	78
139-L	Squaw Creek	Price	109
140-W	Fall Creek	Dunn	29
140-F	Mud Creek	Monroe	47
140-C	East Fork Raccoon Creek	Rock	43
140-L	Swamp Creek	Forest	94
141-W	Trib. to Fennimore Creek	Grant	21
141-F	South Branch Manitowoc River	Calumet	73
141-C	Pensaukee River	Shawano	91
141-L	Price Creek	Sawyer	56
142-W	Pine Creek	Crawford	70
142-F	Fordham Creek	Adams	72
142-C	Willow Creek	Rock	60
142-L	East Branch Lily River	Langlade	52
143-W	Trib. to Wisconsin River	Crawford	21
143-F	Black Creek	Manitowoc	62
143-C	Gilbert Creek	Dunn	95
143-L	Gull Creek	Washburn	26
144-F	Piscasaw Creek	Walworth	35
144-W	Sandy Creek	Grant	53
144-C	Pine River	Richland	97
144-L	Turtle River	Iron	74
145-W	Rossman Creek	Buffalo	22
145-F	Spring Creek	Calumet	52
145-C	Fancy Creek	Richland	75
145-L	Little Roche a Cri Creek	Adams	68
146-W	Taycheedah Creek	Fond du Lac	42
146-F	Pumpkinseed Creek	Waushara	56
146-C	Otter Creek	Eau Claire	66
146-L	Sand Creek	Sawyer	73
147-W	Trib. to Sugar River	Dane	22

<u>ID</u> 147-F	<u>Description</u> Klawitter Creek	<u>County</u> Marquette	Area (sq. km.) 82
147-C	Trout Run	Jackson	44
147-L	Totagatic River	Sawyer	73
148-W	Du Charme Creek	Crawford	35
148-F	Pratt Creek	Dodge	73
148-C	Fisher Creek	Sheboygan	31
148-L	Springstead Creek	Price	42
149-W	West Fork Knapp Creek	Richland	48
149-F	Mole Brook	Marathon	34
149-C	Sevenmile Creek	Fond du Lac	59
149-L	Fishtrap Creek	Sawyer	45
150-W	Big Slough	Trempealeau	22
150-F	East Branch Little Black	Taylor	58
150-C	Trib. to South Fork Eau Claire	Clark	24
150-L	Wolf River	Forest	84
151-W	Trib. to Trempealeau River	Jackson	25
151-F	Eagle Creek	Racine	43
151-C	Johnson Creek	Jefferson	83
151-L	Rice Creek	Vilas	70
152-W	Citron Creek	Crawford	41
152-F	Beaver Creek	Dodge	76
152-C	Sheboygan River	Fond du Lac	76
152-L	Wilson Creek	Sawyer	56
153-W	Weedons Creek	Sheboygan	23
153-F	Soft Maple Creek	Rusk	96
153-C	Gill Creek	Dodge	31
153-L	Enterprise Creek	Oneida	73
154-W	Trout Creek	Iowa	44
154-F	Casco Creek	Kewaunee	44
154-C	Twomile Creek	Wood	45
154-L	Nixon Creek	Vilas	30
155-W	Parsons Creek	Fond du Lac	20
155-F	Little Kickapoo Creek	Crawford	35
155-C	Trib. to Baraboo River	Sauk	28
155-L	East Branch Eau Claire River	Langlade	63
156-W	Squaw Creek	Jackson	54
156-F	Rocky Creek	Wood	60
156-C	Hills Creek	Juneau	44
156-L	Mud Creek	Oneida	91
157-W	North Fork Buffalo River	Trempealeau	78
157-F	Kohlsville River	Washington	52
157-C	Puff Creek	Wood	34
157-L	Little Turtle River	Iron	45
158-W	Big Creek	Sauk	75
158-F	Dunlap Creek	Dane	36
158-C	Bull Branch	Grant	30
158-L	Johnson Creek	Vilas	31

<u><b>ID</b></u> 159-W	<u>Description</u> Trib. to Mississippi River	<u>County</u> Crawford	<u>Area (sq. km.)</u>
159-F	Daly Creek	Oconto	72
159-C	Kankapot Creek	Outagamie	69
159-L	Little Bear Creek	Iron	95
160-W	Copper Creek	Crawford	70
160-F	Casper Creek	Dodge	36
160-C	Cameron Creek	Clark	40
160-L	Spread Eagle Outfit	Florence	24
161-W	Pine Creek	Trempealeau	34
161-F	Lyndon Creek	Juneau	55
161-C	Goggle-Eye Creek	Clark	24
161-L	West Branch Wolf River	Menominee	90
162-W	Sucker Creek	Ozaukee	36
162-F	Silver Creek	Kewaunee	64
162-C	West Creek	Eau Claire	50
162-L	West Fork Chippewa River	Sawyer	72
163-W	Spring Creek	Buffalo	40
163-F	Little Eau Pleine River	Marathon	77
163-C	Bolen Creek	Dunn	38
163-L	South Branch Oconto	Langlade	92
164-W	Mill Creek	Jackson	21
164-F	Shoulder Creek	Rusk	69
164-C	Juda Branch	Green	48
164-L	West Torch River	Ashland	74
165-W	North Fork Beaver Creek	Trempealeau	87
165-F	Honey Creek	Sauk	63
165-C	Bass Creek	Rock	45
165-L	Hay Creek	Price	70
166-L	Trib to Yellow River	Burnett	81
166-W	Trib. to Duck Creek	Outagamie	21
166-F	Onion River	Sheboygan	72
166-C	Trib. to Root River	Racine	99
167-W	Little Green River	Grant	43
167-F	Bark River	Waukesha	71
167-C	Spring Brook	Rock	24
167-L	Lynch Creek	Sawyer	31
168-W	Bostwick Creek	La Crosse	96
168-F	Milwaukee River	Fond du Lac	78
168-C	Trib. to Wisconsin River	Wood	26
168-L	Lily River	Forest	80
169-W	Trib. to Fox River	Winnebago	22
169-F	Trib. to Rock River	Rock	32
169-C	Douglas Creek	Jackson	63
169-L	Connors Creek	Sawyer	38
170-W	Sand Branch	Grant	22
170-F	White River	Walworth	79
170-C	Bridge Creek	Eau Claire	93

<u>ID</u> 170-L	<u>Description</u> Sevenmile Creek	<u>County</u> Oneida	Area (sq. km.)
171-W	Mill Creek	Buffalo	38
171-F	Trib. to Little Turtle Creek	Rock	58
171-C	Jim Creek	Chippewa	31
171-L	Tamarack Creek	Vilas	72
172-W	Trib. to Kickapoo River	Monroe	26
172-F	Tomorrow River	Portage	98
172-C	Trib. to Rock River	Jefferson	41
172-L	Rice Creek	Vilas	65
173-W	Sioux Creek	Barron	28
173-F	Raccoon Creek	Rock	66
173-C	Trib. to Little River	Oconto	22
173-L	Mukwonago River	Waukesha	70
174-W	Trib. to Honey Creek	Sauk	28
174-F	North Branch Crawfish River	Columbia	77
174-C	Trib. to Fox River	Racine	40
174-L	McKenzie Creek	Washburn	38
175-W	Reynolds Coulee Creek	Trempealeau	21
175-F	Sweeny Pond	Barron	24
175-C	Crawfish River	Dodge	31
175-L	Portage Creek	Vilas	47
176-W	North Fork Clam River	Burnett	69
176-F	Root River	Kenosha	27
176-C	Little Trappe River	Marathon	24
176-L	Muskellunge Creek	Oneida	38
177-W	North Branch Manitowoc River	Calumet	28
177-F	Brick Creek	Clark	47
177-C	Dawson Creek	Dodge	25
177-L	Caves Creek	Marquette	26
177-L	Westfield Creek	Marquette	91
177-L	Tagatz Creek	Marquette	58
178-W	Oak Creek	Milwaukee	73
178-F	Deer Tail Creek	Rusk	97
178-C	McGinnis Creek	Marathon	62
178-L	Rocky Run	Oneida	77
179-W	West Branch Fond Du Lac River	Fond du Lac	20
179-F	Mill Creek	Shawano	83
179-C	Fall Creek	Eau Claire	46
179-L	Pickerel Creek	Langlade	110
180-W	Potter Creek	Brown	28
180-F	Rat River	Winnebago	83
180-C	Little Elk Creek	Dunn	42
180-L	Ninemile Creek	Vilas	20
181-W	Missouri Creek	Pepin	82
181-F	Little River	Marinette	39
181-C	Trib. to Rock River	Jefferson	32
181-L	Foulds Creek	Price	51

<u><b>ID</b></u> 182-W	<u>Description</u> Pine Creek	<u>County</u> Manitowoc	Area (sq. km.)
182-F	Fourmile Creek	Barron	43
182-C	Spring Creek	Dane	21
182-L	Upper Inlet	Marinette	41
183-W	Pensaukee River	Shawano	21
183-F	Maple Creek	Outagamie	69
183-C	Big Slough	Columbia	91
183-L	Scuppernong River	Jefferson	93
184-W	Onion River	Sheboygan	45
184-F	Little Menomonee River	Milwaukee	73
184-C	Soper Creek	Monroe	97
184-L	Loon Creek	Shawano	36
185-W	Trib to Cranberry Creek	Juneau	45
185-F	Quaderer Creek	Barron	29
185-C	Baraboo River	Juneau	90
185-L	Kaubashine Creek	Oneida	27
186-W	Babb Creek	Sauk	25
186-F	Saunders Creek	Rock	29
186-C	North Branch Trempealeau River	Jackson	56
186-L	Rat River	Forest	92
187-W	Irving Creek	Dunn	30
187-F	Grand River	Green Lake	50
187-C	Sauk Creek	Ozaukee	81
187-L	South Branch Pike River	Marinette	53
188-W	Twin Creek	Sauk	31
188-F	Stony Brook	Dodge	70
188-C	Pike River	Kenosha	47
188-L	Rice Creek	Price	64
189-W	Sanders Creek	Grant	44
189-F	McCann Creek	Chippewa	71
189-C	Little Sandy Creek	Marathon	31
189-L	Neptune Creek	Oneida	27
190-W	Prentice Creek	Columbia	32
190-F	Spring Creek	Jefferson	25
190-C	Ninemile Creek	Eau Claire	22
190-L	Shell Creek	Washburn	49
191-W	Trib. to Apple River	St Croix	51
191-F	Fisher River	Chippewa	94
191-C	Suamico River	Brown	76
191-L	Lost Creek	Vilas	57
192-W	Willow River	St Croix	35
192-F	Kinnickinnic River	Milwaukee	70
192-C	Arrowhead River	Winnebago	80
192-L	Trib. to Brill River	Washburn	46
193-W	Fly Creek	Trempealeau	26
193-F	Spring Brook	Rock	31
193-C	Deer Creek	Jefferson	27

<u>ID</u> 193-L	<u>Description</u> Moose Creek	<u>County</u> Iron	Area (sq. km.)
194-W	Daggets Creek	Winnebago	30
194-F	Potters Creek	Waupaca	70
194-C	Black River	Jackson	24
194-L	Julia Creek	Oneida	45
195-W	Little Bear Creek	Buffalo	94
195-F	Spring Creek	Portage	47
195-C	Black River	Sheboygan	39
195-L	Randall Creek	Iron	23
196-W	Mill Creek	Richland	25
196-F	Dry Creek	Adams	33
196-C	Elk Creek	Chippewa	86
196-L	Whalen Creek	Washburn	21
197-W	Trib. to Little LaCrosse River	Monroe	23
197-F	Menomonee River	Waukesha	90
197-C	Ross Crossing Creek	Green	28
197-L	Indian Chain Creek	Oneida	23
197-L	Kathan Creek	Oneida	21
198-W	Halfway Prairie Creek	Dane	71
198-F	Mud Creek	Manitowoc	73
198-C	Bear Creek	Richland	91
198-L	Tomahawk Creek	Oneida	21
199-W	Trout Creek	Crawford	23
199-F	Story Creek	Green	59
199-C	Jim Moore Creek	Marathon	21
199-L	Long Lake Branch	Bayfield	87
200-W	Council Creek	Monroe	36
200-F	Trib. to North Branch Crawfish River	Columbia	28
200-C	Black Earth Creek	Dane	86
200-L	Slim Creek	Washburn	38
201-W	Fish Creek	La Crosse	39
201-F	Sawyer Creek	Washburn	88
201-C	Fourmile Creek	Portage	97
201-L	Buckaton Creek	Vilas	45
202-W	East Branch Blue Mounds Creek	Dane	83
202-F	Elm Creek	Wood	95
202-C	Little LaCrosse River	Monroe	75
202-L	Pelican River	Oneida	66
203-W	Trib. to Hemlock Creek	Wood	25
203-F	Hog Creek	Marathon	42
203-C	Pewaukee River	Waukesha	99
203-L	Loon Creek	Burnett	74
204-W	Onion River	Sheboygan	22
204-F	Fox River	Marquette	39
204-C	Devils River	Manitowoc	88
204-L	North Branch Peshtigo Brook	Oconto	72
205-W	North Fork Beaver Creek	Trempealeau	89

<u>ID</u> 205-F	<u>Description</u> Carter Creek	<u>County</u> Adams	Area (sq. km.)
205-C	South Branch Trempealeau	Jackson	57
205-L	Webb Creek	Burnett	63
206-W	Trib. to De Neveu Creek	Winnebago	31
206-F	Fountain Creek	Juneau	28
206-C	Chippewa River	Dunn	39
206-L	Snake Creek	Douglas	33
207-W	Trump Coulee Creek	Trempealeau	27
207-F	Eightmile Creek	Winnebago	82
207-C	Knights Creek	Dunn	84
207-L	Swamp Creek	Oneida	29
208-W	Johns Creek	Dunn	22
208-F	East Branch Honey Creek	Sauk	63
208-C	Trib. to Rock River	Jefferson	21
208-L	Fourmile Creek	Oneida	24
209-W	Trib. to East Twin River	Kewaunee	22
209-F	Trib. to O'Neil Creek	Chippewa	49
209-C	Duck Creek	Outagamie	94
209-L	Denomie Creek	Ashland	48
210-W	Trib. to Mississippi River	Crawford	23
210-F	Muddy Creek	Dunn	78
210-C	Francis Creek	Manitowoc	35
210-L	Rice Creek	Oneida	48
211-W	Tamarack Creek	Trempealeau	96
211-F	Brown Creek	Barron	26
211-C	Allen Creek	Marquette	26
211-L	Eagle River	Oneida	82
212-F	Dutch Gap Canal	Kenosha	32
212-L	Trib to Manitowish River	Iron	61
212-W	Sneed Creek	Iowa	75
212-C	Trib. to Bass Creek	Rock	38
213-W	West Branch Blue Mounds Creek	Iowa	50
213-F	Trib. to Rock River	Jefferson	32
213-C	Trib. to Kinnickinnic River	St Croix	37
213-L	Trib. to Link Creek	Oneida	78
214-W	Dutch Creek	La Crosse	50
214-F	South Branch Beaver Brook	Polk	61
214-C	Trib. to Alto Creek	Dodge	20
214-L	Plum Creek	Vilas	92
215-W	Little Manitowoc River	Manitowoc	35
215-F	Trib. to Wolf River	Outagamie	48
215-C	Trib. to Rock River	Jefferson	22
215-L	Trib. to Fox River	Racine	20
216-W	Marsh Creek	Iowa	92
216-F	Web Creek	Price	83
216-C	Pigeon River	Manitowoc	53
217-W	Trib. to Baraboo River	Sauk	21

<u><b>ID</b></u> 217-F	<u>Description</u> Galloway Creek	<u>County</u> Jefferson	Area (sq. km.)
217-C	Little River	Oconto	32
218-W	Trib. to Hemlock Creek	Wood	20
218-F	Stillson Creek	Chippewa	20
218-C	Lowes Creek	Eau Claire	93
219-W	Fleming Creek	La Crosse	92
219-F	Iron Creek	Dunn	51
219-C	Kriwanek Creek	Manitowoc	31
220-W	Trib. to Blue Mounds Creek	Iowa	20
220-F	Trib. to Crawfish River	Columbia	40
220-C	Dell Creek	Sauk	92
221-W	Rocky Run	Portage	24
221-F	Trib. to Turtle Creek	Walworth	40
221-C	Fourteenmile Creek	Adams	85
222-W	Wilson Creek	Sauk	41
222-F	Potato Creek	Rusk	76
222-C	Point Creek	Manitowoc	57
223-W	Corning Creek	Adams	38
223-F	Deer Creek	Jefferson	41
223-C	Black Creek	Outagamie	93
224-W	Johnson Creek	Marathon	79
224-F	Stony Creek	Kewaunee	62
224-C	Nolan Creek	Dodge	23
225-W	Taylor Creek	Eau Claire	20
225-F	Fox River	Columbia	90
225-C	Wildcat Creek	Dodge	22
226-W	Fischer Creek	Manitowoc	30
226-F	Beaver Creek	Marathon	21
226-C	Eighteenmile Creek	Dunn	75
227-W	Little Lemonweir River	Juneau	91
227-F	Mud Creek	Manitowoc	76
227-C	Beaver Creek	Eau Claire	47
228-W	Big Cain Creek	Marathon	28
228-F	West Branch Fond Du Lac River	Fond du Lac	75
228-C	Como Creek	Walworth	43
229-W	Spencer Creek	Monroe	26
229-F	Bundy Creek	Marinette	88
229-C	Branch River	Brown	88
230-W	Devils Creek	Ashland	40
230-F	Alto Creek	Dodge	39
230-C	Hulburt Creek	Sauk	37
231-W	Big Beaver Creek	Dunn	50
231-F	Pickerel Creek	Shawano	36
231-C	Baker Creek	Dodge	36
232-W	Rajek Creek	Lincoln	24
232-F	Trib. to Fox River	Green Lake	76
232-C	Spring Brook	Winnebago	57

<u>ID</u> 233-F	<u>Description</u> Trib to Channel Lake	<u>County</u> Kenosha	Area (sq. km.)
233-W	Trib. to West Branch Fond Du Lac River	Fond du Lac	27
233-C	Mud Run	Dodge	27
234-W	Barnes Creek	Lincoln	24
234-F	Bear Creek	Outagamie	77
234-C	Trib. to Alto Creek	Dodge	41
235-W	Indian Creek	Burnett	38
235-F	Babit Creek	Taylor	24
235-C	Trib. to Duck Creek	Outagamie	45
236-W	Trib to Wisconsin River	Sauk	34
236-F	Nine Springs Creek	Dane	32
236-C	Cranberry Creek	Dunn	81
237-W	King Creek	Trempealeau	33
237-F	Little Eau Claire River	Marathon	100
237-C	Pebble Creek	Waukesha	48
238-W	Trib. to Wisconsin River	Grant	26
238-F	East Fork Black River	Wood	99
238-C	Molash Creek	Manitowoc	45
239-W	Tamarack Creek	Buffalo	46
239-F	Hay Creek	Chippewa	89
239-C	Spring Creek	Walworth	24
240-C	Trib to Green Bay	Brown	31
240-W	Pony Creek	Shawano	32
240-F	Whitefish Bay Creek	Door	67
241-W	Trib. to Little LaCrosse River	Monroe	29
241-F	Elder Creek	Chippewa	51
241-C	Kinnickinnic River	St Croix	25
242-W	Byrds Creek	Richland	29
242-F	Duck Creek	Brown	32
242-C	Yellow River	Wood	20
243-W	Pike Creek	Kenosha	46
243-F	Walla Walla Creek	Waupaca	54
243-C	Liberty Creek	Green	33
244-W	South Fork Paint Creek	Chippewa	27
244-F	Shivering Sands Creek	Door	31
244-C	Duck Creek	Brown	20
245-W	Sevenmile Creek	Sheboygan	30
245-F	Trib. to Henderson Creek	Winnebago	36
245-C	Trib. to Bear Creek	Outagamie	60
246-W	Horse Creek	Richland	23
246-F	Blake Creek	Waupaca	97
246-C	Hoods Creek	Racine	40
247-W	Trappers Creek	Taylor	32
247-F	School Section Creek	Shawano	37
247-C	Como Creek	Chippewa	27
248-W	Baird Creek	Brown	49
248-F	Meeme River	Manitowoc	55

<u>ID</u> 248-C	<u>Description</u> Kirchner Creek	County Oconto	Area (sq. km.)
249-W	Kennedy Creek	Marathon	25
249-F	Trib. to Milwaukee River	Fond du Lac	41
249-C	Trib. to Kinnickinnic River	Pierce	22
250-W	Trib. to Mississippi River	Buffalo	39
250-F	Black Creek	Eau Claire	96
250-C	Tibbet Creek	Oconto	38
251-W	South Fork Buffalo River	Trempealeau	81
251-F	Fox River	Waukesha	61
251-C	Herman Creek	Outagamie	63
252-W	French Creek	Jackson	71
252-F	Shaw Creek	Waupaca	40
252-C	North Branch Manitowoc River	Calumet	78
253-W	Pine Creek	Jackson	28
253-F	Silver Creek	Sheboygan	51
253-C	Otter Creek	Sheboygan	30
254-W	Rocky Run	Wood	91
254-F	Grand River	Green Lake	83
254-C	Thomas Slough	Oconto	37
255-C	Trib to Lake Michigan	Door	26
255-W	Little Bear Creek	Richland	37
255-F	Alder Creek	Winnebago	38
256-W	Clear Creek	Rusk	31
256-F	Red River	Kewaunee	43
256-C	Bear Creek	Pepin	86
257-W	Town Line Creek	Jackson	22
257-F	Keyes Creek	Door	30
257-C	Sand Creek	Dunn	53
258-W	Conlan Creek	Clark	22
258-F	Willow Creek	Fond du Lac	21
258-C	Pokegama Creek	Barron	95
259-W	Vosse Coulee Creek	Jackson	25
259-F	Hinkson Creek	Columbia	49
259-C	Grand River	Marquette	31
260-W	Popple Creek	Dunn	21
260-F	Lincoln Creek	Milwaukee	56
260-C	Irish Creek	Dodge	24
261-W	Little Suamico River	Oconto	30
261-F	Trib. to Rock River	Jefferson	31
261-C	Roaring Creek	Jackson	24
262-W	Sand Creek	Monroe	31
262-F	Levitt Creek	Taylor	81
262-C	Trib. to Branch River	Manitowoc	31
263-W	Lakes Coulee Creek	Trempealeau	33
263-F	Hatton Creek	Waupaca	78
263-C	New Channel	La Crosse	41
264-W	Three Springs Creek	Door	22

<u>ID</u> 264-F	<u>Description</u> Whitcomb Creek	<u>County</u> Waupaca	Area (sq. km.)
264-C	Stony Creek	Jackson	44
265-W	Tank Creek	Jackson	33
265-F	Cedar Creek	Calumet	65
265-C	Trib. to Milwaukee River	Washington	31
266-W	North Branch Pine River	Lincoln	62
266-F	Trib. to Sheboygan River	Fond du Lac	51
266-C	Becky Creek	Rusk	26
267-W	La Crosse River	La Crosse	72
267-F	Rio Creek	Kewaunee	63
267-C	Killsnake River	Calumet	83
268-F	Trib to Sheboygan River	Manitowoc	42
268-W	Trib. to Cedar Creek	Washington	21
268-C	Mink Creek	Sheboygan	50
269-W	Twentymile Creek	Bayfield	47
269-F	Bear Creek	Portage	56
269-C	Trib. to Badfish Creek	Dane	26
270-W	Pine River	Langlade	66
270-F	Hayes Creek	Oconto	38
270-C	Trib. to Fourteenmile Creek	Adams	28
271-W	Stony Creek	Jackson	25
271-F	Little Black River	Taylor	58
271-C	Black Brook	St Croix	43
272-W	Trib. to Plover River	Portage	34
272-F	Rat River	Winnebago	33
272-C	Trib. to Duck Creek	Outagamie	59
273-W	German Creek	Barron	22
273-F	Lilly Bay Creek	Door	44
273-C	Trib. to Fox River	Winnebago	57
274-F	Trib to Green Bay	Oconto	26
274-W	Devils Creek	Ashland	27
274-C	Crossman Creek	Sauk	53
275-W	Cramer Creek	Price	20
275-F	Belle Fountain Creek	Green Lake	89
275-C	Trib. to North Branch Milwaukee River	Washington	36
276-W	Spring Brook	Ashland	27
276-F	Rice Creek	Polk	22
276-C	Big Creek	La Crosse	49
277-W	Wood Creek	Taylor	81
277-F	Black Creek	Green Lake	61
277-C	Little West Branch Wolf River	Menominee	88
278-C	Trib to Fox River	Green Lake	29
278-W	Boomer Creek	Iron	42
278-F	Trib. to Yahara River	Dane	71
279-W	Knuteson Creek	Sawyer	67
279-F	Trib to Beaver Creek	Dodge	27
279-C	Trib to S. Branch Manitowoc River	Calumet	22

<u>ID</u>	<u>Description</u>	County	Area (sq. km.)
280-F	Trib to Green Bay	Brown	28
280-W	Yellow River	Barron	98
280-C	Beaver Creek	Monroe	45
281-W	Trib to Little LaCrosse River	Monroe	20
281-C	Trib to West Branch Milwaukee River	Fond du Lac	31
281-F	Pine Creek	Calumet	74
282-F	Trib to Pine River	Waushara	42
282-W	Vermont Creek	Dane	39
282-C	Poplar Creek	Waukesha	66
283-W	Trib to Oconto River	Oconto	21
283-C	Trib to Wolf River	Winnebago	22
283-F	Peplin Creek	Marathon	49
284-W	Klein Creek	Adams	47
284-F	Trib to Wisconsin River	Portage	28
284-C	Big Rock Creek	Polk	29
285-W	Devils Creek	Rusk	65
285-C	Trib to Milwaukee River	Ozaukee	23
285-F	Lost Creek	Portage	37
286-W	Jader Creek	Bayfield	21
286-F	Little Creek	Waupaca	34
286-C	Sugar Creek	Door	38
287-F	Peterson Creek	Kenosha	25
287-W	Rock Creek	Jackson	89
287-C	Farmers Valley Creek	Monroe	61
288-F	Rose Brook	Shawano	33
288-C	Copper Creek	Sauk	21
288-W	Lawrence Creek	Iron	35
289-F	Trib to Bear Creek	Outagamie	33
289-C	Trib to Branch River	Manitowoc	23
289-W	Baldwin Creek	Lincoln	31
290-F	Trib to Wisconsin River	Adams	48
290-C	Trib to Little River	Oconto	21
290-W	Schramm Creek	Bayfield	49
291-W	Spirit River	Lincoln	99
291-F	Underwood Creek	Milwaukee	55
291-C	Black Creek	Clark	40
292-F	Trib to Sheboygan River	Sheboygan	25
292-C	Trib to Trout Creek	Marinette	23
292-W	Tiger Creek	Shawano	49
293-W	Dent Creek	Shawano	20
293-F	Pine River	Waushara	97
293-C	Hay Creek	Sauk	24
294-C	Trib to Cedar Creek	Washington	20
294-W	Potato River	Iron	83
294-F	Bull Brook	Polk	66
295-W	Fisher Creek	Florence	31
295-F	Kroenke Creek	Shawano	31
		2	<b>O</b> .

<u>ID</u> 295-C	<u>Description</u> Trib to Onion River	<u>County</u> Sheboygan	Area (sq. km.)
296-W	Silver Creek	Taylor	81
296-C	Trib to Little River	Oconto	22
296-F	Husher Creek	Racine	29
297-W	Rock Creek	Dunn	95
297-F	Toad Creek	Outagamie	42
297-C	Webster Creek	Juneau	41
298-F	O'Keefe Creek	Marquette	61
298-C	Apple River	Polk	84
298-W	Joe Snow Creek	Lincoln	29
299-F	Trib to Yahara River	Dane	26
299-C	Hibbard Creek	Door	54
299-W	McCloud Creek	Langlade	43
300-C	Trib to Mud Creek	Manitowoc	28
300-W	Little Elk River	Price	79
300-F	Bear Creek	Outagamie	42
301-F	Trib to Balsam Branch	Polk	50
301-C	Stony Creek	Washington	54
301-W	Coon Creek	Dunn	40
302-W	N. Branch Beaver Creek	Marinette	49
302-C	Trib to Sheboygan River	Fond du Lac	25
302-F	Mouse Creek	Waupaca	21
303-C	Trib to Grand River	Green Lake	22
303-W	Holt Creek	Marathon	41
303-F	North Branch Beaver Brook	Polk	64
304-W	Silver Creek	Shawano	44
304-F	Lau Creek	Dodge	23
304-C	Turner Creek	Wood	68
305-W	Trib to Wisconsin River	Juneau	22
305-F	Trib to Fox River	Green Lake	48
305-C	Alder Creek	Rusk	31
306-W	McKenzie Creek	Taylor	63
306-F	Willow Creek	Waushara	21
306-C	Otter Creek	Chippewa	88
307-F	Trib to Fox River	Winnebago	33
307-W	Owl Creek	Wood	21
307-C	Fivemile Creek	Clark	95
308-W	South Fish Creek	Bayfield	77
308-F	Beaver Creek	Barron	76
308-C	Mink Creek	Taylor	39
309-C	Mukwonago River	Waukesha	40
309-W	Lambs Creek	Dunn	47
309-F	Tenmile Creek	Barron	77
310-F	Trib to White River	Walworth	43
310-W	Hay Creek	Dunn	44
310-C	Allen Creek	Jefferson	30
311-W	Moose Ear Creek	Barron	83

<u>ID</u> 311-F	<u>Description</u> Genesee Creek	<u>County</u> Waukesha	Area (sq. km.)
311-C	Trade River	Polk	70
312-W	Trib to South Fish Creek	Bayfield	24
312-C	Bradley Creek	Waupaca	26
312-F	Battle Creek	Jefferson	24
313-W	Blueberry Creek	Douglas	47
313-F	Mud Brook	Marinette	26
313-C	Comet Creek	Waupaca	89
314-W	Sevenmile Creeek	Wood	39
314-F	Gilson Creek	Brown	23
314-C	Pebble Brook	Waukesha	47
315-F	Yahara River	Dane	30
315-C	Little Oconomowoc River	Waukesha	29
315-W	Twin Creek	Marinette	41
316-F	Heins Creek	Door	40
316-C	Cedar Creek	Washington	35
316-W	Hay Creek	Rusk	27
317-W	North Fork Copper River	Lincoln	95
317-F	Trib to Sheboygan River	Sheboygan	41
317-C	Fairbanks Creek	Adams	51
318-F	Skunk Creek	Rusk	29
318-W	Skulen Creek	Marathon	32
318-C	Butternut Creek	Polk	47
319-W	Bingham Creek	Adams	32
319-F	Hay Creek	Taylor	70
319-C	Browns Creek	Eau Claire	31
320-W	Trib to East Branch Eau Claire River	Langlade	30
320-C	Mecan River	Waushara	93
320-F	Larson Creek	Door	24
321-C	Trib to Montello River	Marquette	25
321-W	Carpenter Creek	Waushara	32
321-F	Straight River	Polk	82
322-W	Holmes Creek	Price	51
322-F	Trib to Otter Creek	Jefferson	25
322-C	Wolf Creek	Polk	90
323-W	Oxbo Creek	Lincoln	21
323-F	Trib to Little Peshtigo River	Marinette	23
323-C	Middle Branch Embarass River	Shawano	86
324-W	Trib to Clam River	Burnett	29
324-F	South Fork Main Creek	Rusk	93
324-C	West Branch Red River	Shawano	79
325-W	Dead Horse Creek	Adams	84
325-F	Trib to Muddy Creek	Dunn	23
325-C	North Branch Embarrass River	Shawano	79
326-W	Silver Creek	Ashland	24
326-F	Rice Creek	Barron	56
326-C	Otter Creek	Dunn	95

<u>ID</u> 327-C	<u>Description</u> Spranger Creek	<u>County</u> Shawano	<u>Area (sq. km.)</u> 37
327-W	Squaw Creek	Lincoln	48
327-F	Tank Creek	La Crosse	23
328-F	Stony Brook	Calumet	28
328-W	Joe Creek	Taylor	22
328-C	Sinking Creek	Dunn	32
329-W	Trib to Lemonweir River	Juneau	49
329-F	Clam River	Polk	93
329-C	Paradise Creek	Taylor	47
330-W	Spaulding Creek	Waupaca	26
330-F	Willow Creek	Waushara	20
330-C	Flume Creek	Portage	96
331-F	Little Jump River	Rusk	20
331-C	South Branch Embarrass River	Shawano	79
331-W	Allen Creek	Forest	50
332-W	Spring Creek	Douglas	29
332-F	Hay Creek	Wood	88
332-C	Auburn Lake Creek	Fond du Lac	35
333-W	Spring Lake Creek	Sawyer	38
333-F	Trib to Puckaway Lake	Green Lake	25
333-C	Dandy Creek	Monroe	28
334-W	Little Weirgor Creek	Sawyer	99
334-F	Bassett Creek	Kenosha	23
334-C	Peterson Creek	Waupaca	70
335-W	Levis Creek	Jackson	97
335-F	Johnson Creek	Manitowoc	23
335-C	Knapp Creek	Polk	25
336-W	Saint Croix River	Douglas	92
336-F	Trib to Cedar Creek	Ozaukee	26
336-C	South Fork Clam River	Burnett	47
337-W	Evergreen River	Menominee	79
337-C	Trib to Mud Creek	Manitowoc	24
337-F	Duchess Creek	Shawano	28
338-W	Little Hay Meadow Creek	Lincoln	64
338-F	Trib to Bark River	Jefferson	29
338-C	Gardner Creek	Shawano	23
339-F	Trib to Mecan River	Marquette	23
339-W	Nichol Creek	Waupaca	27
339-C	Sucker Creek	Green Lake	54
340-F	Trib to Grand River	Green Lake	20
340-C	Trib to Branch River	Manitowoc	25
340-W	North Fork Spirit River	Lincoln	93
341-W	Douglas Creek	Price	74
341-F	Mud Creek	Rusk	90
341-C	Nace Creek	Waupaca	30
342-W	Kurt Creek	Wood	41
342-F	Ox Creek	Marquette	36

<u>ID</u> 342-C	<u>Description</u> Black Brook	County Langlade	Area (sq. km.)
343-F	Friday Creek	Polk	28
343-W	North Fork Thunder River	Marinette	78
343-C	Copper River	Lincoln	97
344-F	Little Pine Creek	Marquette	26
344-W	South Fork Thunder River	Marinette	57
344-C	Lower Middle Inlet	Marinette	78
345-W	Holmes Creek	Marinette	41
345-C	Packard Creek	Shawano	54
345-F	O'Neil Creek	Chippewa	44
346-W	Hanson Creek	Bayfield	27
346-F	Scuppernong Creek	Waukesha	52
346-C	Manley Creek	Sauk	34
347-C	Wedde Creek	Marquette	63
347-W	Miscauno Creek	Marinette	60
347-F	McKenzie Creek	Polk	44
348-W	Medicine Brook	Marinette	28
348-F	Hay Meadow Creek	Portage	72
348-C	Little Wolf River	Marathon	64
349-W	Montagne Creek	Florence	39
349-F	Rice Creek	Rusk	108
349-C	Christmas Creek	Chippewa	31
350-C	Pammel Creek	Vernon	41
350-W	Deer Creek	Ashland	28
350-F	Mosquito Creek	Waupaca	41
351-W	Black Alder Creek	Lincoln	30
351-F	Bull Junior Creek	Marathon	98
351-C	Cedar Creek	Marathon	21
352-F	Osceola Creek	Polk	34
352-W	Eddy Creek	Sawyer	31
352-C	Pigeon Creek	Barron	24
353-W	Trib to Robinson Creek	Jackson	29
353-C	Plover River	Marathon	52
353-F	Snake Creek	Green Lake	28
354-W	Big Hay Meadow Creek	Lincoln	77
354-C	Little Jump River	Rusk	51
354-F	Trappe River	Marathon	79
355-F	Trib to Grand River	Green Lake	21
355-W	Lemke Creek	Taylor	52
355-C	Mollies Creek	Jackson	35
356-W	Mondeaux Creek	Price	38
356-F	French Creek	Columbia	62
356-C	Indian Creek	Jackson	37
357-W	Hines Creek	Oconto	23
357-F	Black Creek	Marathon	35
357-C	Upper Middle Inlet	Marinette	73
358-F	Hay Creek	Clark	95

<u>ID</u> 358-W	<u>Description</u> Little Bois Brule River	<u>County</u> Douglas	Area (sq. km.)
358-C	Gravelly Brook	Marinette	40
359-F	Stuntz Brook	Washburn	53
359-C	Clearwater Creek	Langlade	30
359-W	Green Meadow Creek	Lincoln	52
360-F	Trib to Mud Lake	Door	25
360-W	Alder Creek	Iron	58
360-C	North Fork Main Creek	Rusk	55
361-W	Clemens Creek	Burnett	24
361-F	Trib to Apple River	Polk	24
361-C	Trib to Lemonweir River	Juneau	46
362-F	Trib. to Waupaca River	Portage	54
362-C	Bruce Creek	Waushara	38
362-W	Little Wausaukee Creek	Marinette	30
363-C	Horse Creek	Eau Claire	34
363-W	Sullivan Creek	Marinette	27
363-F	Middle Fork Main Creek	Rusk	60
364-F	Cedar Springs Creek	Waushara	29
364-C	Trib. to Wisconsin River	Juneau	54
364-W	Hay Creek	Sawyer	44
365-F	Trib. to South Branch Little Wolf River	Waupaca	25
365-C	Willow Creek	Waushara	78
365-W	Spring Creek	Washburn	29
366-W	Bad River	Ashland	21
366-F	Skinner Creek	Rusk	83
366-C	Sucker Creek	Barron	33
367-W	Trib. to Marengo River	Bayfield	23
367-F	Rice Bed Creek	Polk	42
367-C	Godfrey Creek	Washburn	38
368-F	Peshtigo Brook	Oconto	33
368-C	Little Silver Creek	Waushara	43
368-W	Rock Creek	Sawyer	26
369-C	Smith Lake Creek	Sawyer	32
369-W	Cap Creek	Bayfield	29
369-F	Twin Creek	Rusk	51
370-C	Trib. to Menominee River	Marinette	40
370-W	Frog Creek	Washburn	75
370-F	Kenyon Creek	Sawyer	64
371-W	Trib. to Eau Claire River	Douglas	39
371-F	South Branch Peshtigo River	Forest	53
371-C	Logemanns Creek	Shawano	26
372-W	Wausaukee River	Marinette	92
372-F	Trib. to Mecan River	Marquette	26
372-C	Oshkosh Creek	Menominee	28
373-C	Smith Creek	Price	28
373-W	Oronto Creek	Iron	45
373-F	Middle Branch Peshtigo River	Forest	42

<u><b>ID</b></u> 374-W	<u>Description</u> Dalles Creek	<u>County</u> Menominee	<u>Area (sq. km.)</u>
374-F	Needle Creek	Price	39
374-C	Murray Creek	Price	28
375-W	LeRoy Creek	Florence	31
375-F	Little Popple River	Florence	92
375-C	Summit Creek	Sawyer	37
376-W	Jackson Creek	Waupaca	21
376-F	North Fork Yellow River	Taylor	98
376-C	Crescent Creek	Lincoln	44
377-F	Page Creek	Marquette	25
377-W	Kakagon River	Ashland	87
377-C	Little Frog Creek	Washburn	45
378-W	North Branch Pike River	Marinette	96
378-F	North Fork Clam River	Burnett	36
378-C	Silver Creek	Juneau	29
379-W	Squaw Creek	Marinette	23
379-F	White Creek	Adams	20
379-C	Hemlock Creek	Barron	68
380-W	Swift Creek	Sawyer	23
380-F	Little Eau Claire River	Portage	24
380-C	Miller Creek	Shawano	66
381-C	Mondeaux River	Taylor	77
381-W	Slough Creek	Marinette	28
381-F	Mackay Creek	Washburn	50
382-W	New Wood River	Lincoln	100
382-F	Crawford Creek	Douglas	21
382-C	White Creek	Jackson	27
383-F	Alder Creek	Rusk	46
383-C	Crazy Horse Creek	Rusk	51
383-W	Spikehorn Creek	Marinette	35
384-W	Handsaw Creek	Marinette	25
384-F	Armstrong Creek	Forest	95
384-C	Hunting River	Langlade	93
385-W	Little West Branch Creek	Menominee	100
385-F	Logging Creek	Polk	81
385-C	South Fork Yellow River	Taylor	58
386-F	Bean Brook	Washburn	70
386-W	Chases Brook	Burnett	95
386-C	Lepage Creek	Florence	30
387-F	Squaw Lake Creek	Sawyer	67
387-W	Miller Creek	Douglas	22
387-C	Pine Creek	Bayfield	45
388-F	Trib. to Yellow River	Burnett	22
388-W	North Branch Prairie River	Lincoln	98
388-C	Hay Creek	Price	27
389-F	Deer Creek	Price	26
389-C	Spring Creek	Taylor	20

<u>ID</u> 389-W	<u>Description</u> Little Waupee Creek	<u>County</u> Oconto	Area (sq. km.)
390-W	Trib. to Wolf River	Langlade	26
390-F	Beaver Creek	Price	27
390-C	Hay Creek	Price	76
391-W	Elk River	Price	87
391-F	Carpenter Creek	Price	32
391-C	Bearskin Creek	Oneida	71
392-W	Trib. to Little Yellow River	Juneau	24
392-F	Patterson Creek	Price	21
392-C	Pine Creek	Price	66
393-F	Black Brook	Burnett	66
393-C	South Branch Pemebonwon	Marinette	99
393-W	Trout Brook	Ashland	29
394-W	Trib. to North Fork Jump River	Price	20
394-F	Schraum Creek	Ashland	38
394-C	Little Pine Creek	Lincoln	90
395-W	Scott Creek	Lincoln	31
395-F	Crooked Creek	Rusk	24
395-C	Musser Creek	Price	27
396-F	Bosner Creek	Ashland	28
396-C	Mosquito Brook	Sawyer	39
396-W	Big Brook	Bayfield	70
397-F	Trib. to Namekagon River	Washburn	39
397-C	Trib. to Pine River	Florence	25
397-W	Castle Creek	Bayfield	28
398-W	Sheosh Creek	Douglas	35
398-F	Nail Creek	Rusk	67
398-C	Hobbles Creek	Price	76
399-C	Chicog Creek	Washburn	63
399-W	Averill Creek	Lincoln	45
399-F	Little Mondeaux Creek	Price	58
400-W	Little South Branch Pike	Marinette	73
400-F	Lamon Tangue Creek	Florence	63
400-C	Copper Creek	Douglas	46
401-W	Buckley Creek	Douglas	22
401-F	Iron River	Bayfield	83
401-C	Elvoy Creek	Forest	50
402-W	Pipestone Creek	Sawyer	21
402-C	Trib. to East Branch Eau Claire River	Langlade	23
402-F	Pokegama River	Douglas	84
403-W	Little Thornapple River	Sawyer	32
403-F	Section Twenty Creek	Sawyer	22
403-C	Ounce River	Douglas	97
404-C	Thornapple River	Sawyer	85
404-W	Landwehr Creek	Lincoln	23
404-F	Poplar River	Douglas	94
405-F	Trib. to North Fork Wood River	Burnett	77

<u><b>ID</b></u> 405-W	<u>Description</u> Davidson Creek	<u>County</u> Jackson	Area (sq. km.)
405-C	Weasel Creek	Sawyer	62
406-C	Smith Creek	Price	30
406-W	Eagle Creek	Marinette	80
406-F	Bluff Creek	Douglas	52
407-W	Chase Creek	Price	31
407-F	Mishonagon Creek	Vilas	46
407-C	Knowles Creek	Oconto	22
408-F	Hoffman Creek	Ashland	24
408-C	Hay Creek	Burnett	38
408-W	Hendricks Creek	Florence	33
409-F	Nelson Creek	Price	25
409-W	Camp F Creek	Marinette	71
409-C	North Branch Peshtigo River	Forest	90
410-W	Wisconsin Creek	Florence	30
410-F	Swamp Creek	Iron	78
410-C	Larson Creek	Bayfield	28
411-F	Steve Creek	Price	29
411-W	Otter Creek	Forest	75
411-C	Popple Creek	Price	48
412-W	Deer Creek	Ashland	24
412-F	Rock Creek	Price	26
412-C	Dead Creel	Sawyer	78
413-W	Crotte Creek	Douglas	52
413-F	Bardon Creek	Douglas	35
413-C	Vaughn Creek	Ashland	71
414-W	Middle River	Douglas	92
414-C	Johnson Creek	Florence	22
414-F	Camp Eight Creek	Forest	43
415-W	Woods Creek	Florence	85
415-F	Fivemile Creek	Washburn	57
415-C	Torpee Creek	Forest	34
416-W	Halley Creek	Forest	26
416-F	Bear Creek	Douglas	24
416-C	South Branch Popple River	Florence	88
417-W	South Branch Presque Isle River	Vilas	42
417-F	Hill Creek	Bayfield	90
417-C	Iron River	Ashland	78
418-W	George Ladd Creek	Sawyer	23
418-F	Monico Creek	Oneida	67
418-C	Magee Creek	Ashland	48
419-W	West Fork Montreal River	Iron	89
419-F	Muskeg Creek	Bayfield	63
419-C	Ninemile Creek	Langlade	58
420-F	Bear Creek	Oneida	21
420-W	Tupper Creek	Sawyer	70
420-C	Thompson Creek	Douglas	51

<u><b>ID</b></u> 421-W	<u>Description</u> Raspberry River	<u>County</u> Bayfield	Area (sq. km.)
421-F	Pearson Creek	Douglas	29
421-C	Laymans Creek	Iron	51
422-F	Halls Creek	Florence	23
422-W	Bois Brule River	Douglas	138
422-C	Log Creek	Sawyer	57
423-W	Hay Creek	Burnett	35
423-F	Hay Creek	Lincoln	21
423-C	Haymeadow Creek	Oneida	57
424-C	Lauterman Creek	Florence	25
424-W	K C Creek	Marinette	46
424-F	Whittlesey Creek	Bayfield	63
425-W	Muskrat Creek	Vilas	26
425-F	Little George Creek	Oneida	22
425-C	Weber Creek	Iron	21
426-C	Little Popple River	Florence	32
426-W	Coffee Creek	Lincoln	25
426-F	North Branch Pine River	Forest	93
427-F	Cole Creek	Douglas	21
427-C	Balsam Creek	Douglas	79
427-W	Squaw Creek	Sawyer	23
428-W	Berry Creek	Lincoln	23
428-F	Kingstone Creek	Forest	21
428-C	Little Willow Creek	Oneida	67
429-F	Muskellunge Creek	Ashland	26
429-W	Armstrong Creek	Lincoln	31
429-C	Jones Creek	Forest	51
430-W	East Fork Chippewa River	Ashland	77
430-F	Indian Creek	Oneida	26
430-C	Skanawan Creek	Lincoln	24
431-W	Hoffman Creek	Price	28
431-F	Flag River	Bayfield	90
431-C	Fourmile Creek	Bayfield	27
432-W	Trib. to South Branch Presque Isle River	Vilas	55
432-F	McCaslin Brook	Oconto	37
432-C	Prairie River	Langlade	66
433-W	Cranberry Creek	Douglas	26
433-F	East Branch Presque Isle River	Vilas	48
433-C	McDonald Creek	Forest	33
434-W	Brown Creek	Oneida	28
434-F	Beecher Creek	Marinette	26
434-C	Brunet River	Sawyer	35
435-W	East Torch River	Ashland	37
435-F	Mud Creek	Forest	30
435-C	Threemile Creek	Price	21
436-W	Deer Creek	Sawyer	23
436-F	Silver Creek	Douglas	39

<u>ID</u> 436-C	<u>Description</u> Bad River	<u>County</u> Ashland	<u>Area (sq. km.)</u> 91
437-W	Devils Creek	Sawyer	21
437-F	South Fork White River	Bayfield	73
437-C	Rocky Run	Ashland	34
438-W	Cowan Creek	Polk	28
438-F	Sioux River	Bayfield	86
438-C	Popple River	Forest	72
439-W	Moose River	Douglas	84
439-F	Hungry Run	Ashland	32
439-C	Moose River	Ashland	60
440-W	Siphon Creek	Vilas	22
440-C	North Fish Creek	Bayfield	37
440-F	Smith Creek	Douglas	21
441-F	East Fork Iron River	Bayfield	72
441-C	Spider Creek	Langlade	25
441-W	Hackett Creek	Rusk	29
442-W	Allequash Creek	Vilas	22
442-F	Kolin Creek	Bayfield	29
442-C	Fish Creek	Bayfield	39
443-W	Riley Creek	Price	24
443-F	Reefer Creek	Bayfield	30
443-C	Meadow Creek	Ashland	50
444-F	Trib. to Johnson Creek	Vilas	25
444-W	Red Cliff Creek	Bayfield	21
444-C	Little Amnicon River	Douglas	55
445-F	Johnson Creek	Oneida	27
445-W	Bergen Creek	Washburn	79
445-C	Pine River	Forest	76
446-W	Spruce River	Douglas	71
446-F	South Branch Pike River	Marinette	30
446-C	Haymeadow Creek	Vilas	37
447-W	Upper Ox Creek	Douglas	31
447-F	Elm Creek	Ashland	21
447-C	Little Sioux River	Bayfield	55
448-W	Bootjack Creek	Oneida	27
448-F	Stevenson Creek	Vilas	21
448-C	Sand River	Bayfield	79
449-W	Trib. to Moose River	Sawyer	20
449-F	Lenawee Creek	Bayfield	24
449-C	Siskiwit River	Bayfield	63
450-W	Garland Creek	Vilas	20
450-F	Bark River	Bayfield	28
450-C	Pikes Creek	Bayfield	84
451-W	Trib. to Saint Croix River	Douglas	24
451-F	Lost Creek Number One	Bayfield	25
451-C	Trib. to East Fork Cranberry River	Bayfield	21
452-W	East Fork Cranberry River	Bayfield	97

#### **Appendix C**

## Tables—Economic Impacts of Alternative Management Practices on Selected Wisconsin Farms

- Table 1. BASE and Two No-Till SNAP-Plus Simulations for Case 1.
- Table 2. Comparison of Corn Grain SNAP-Plus Simulations for Case 1.
- Table 3. Aggregate Farm Results Over Full Rotation for Case 1.
- Table 4. Comparison of BASE and Two SNAP-Plus Simulations for Case 2.
- Table 5. Aggregate Farm Results Over Full Rotation for Case 2.
- Table 6. Comparison of BASE and Two SNAP-Plus Simulations for Case 3.
- Table 7. Aggregate Farm Results Over Full Rotation for Case 3.

Table 1. BASE and Two No-Till SNAP-Plus Simulations for Case 1.

Corn Silage,	Corn Silage, Spring Chisel, Avg Yields, Winter Spreading  Corn Silage, No-Till, Avg Yields, Winter Spreading  Corn Silage, No-Till, Avg Yields, Winter Spreading															
	BA	SE Scenari	o			NT/	WinterSprea	ding: Ola-A	A-A-A-Csl-C	Csl		NT/N	NoWinterSpre	eading: Ola	-A-A-A-Csl	-Csl
Costs	Revenues	Profits	Soil Loss	P-Index		Costs	Revenues	Profits	Soil Loss	P-Index		Costs	Revenues	Profits	Soil Loss	P-Index
\$1,996	\$2,153	\$157	7.7	21.4		\$1,932	\$2,153	\$221	1	3.4		\$1,932	\$2,153	\$221	1.0	3.4
\$3,559	\$3,968	\$409	7.5	18.8		\$3,474	\$3,968	\$494	1	4.1		\$3,474	\$3,968	\$494	1.0	4.1
\$7,983	\$8,610	\$627	7.5	21.3		\$7,728	\$8,610	\$882	1	4.6		\$7,728	\$8,610	\$882	1.0	4.6
\$7,983	\$8,610	\$627	7.6	17.8		\$7,728	\$8,610	\$882	1	2.8		\$7,728	\$8,610	\$882	1.0	2.8
			-													
\$7,983	\$8,610	\$627	7.6	21.7		\$7,728	\$8,610	\$882	1	4.6		\$7,728	\$8,610	\$882	1.0	4.6
\$7,983	\$8,610	\$627	7.6	22.6		\$7,728	\$8,610	\$882	1	3.5		\$7,728	\$8,610	\$882	1.0	3.5
\$7,983	\$8,610	\$627	7.5	23.4		\$7,728	\$8,610	\$882	1	5.0		\$7,728	\$8,610	\$882	1.0	5.0
\$3,992	\$4,305	\$313	7.6	19.5		\$3,864	\$4,305	\$441	1	3.0		\$3,864	\$4,305	\$441	1.0	3.0
\$3,559	\$3,968	\$409	5.1	16.8		\$3,474	\$3,968	\$494	0.6	6.0		\$3,474	\$3,968	\$494	0.6	2.4
\$3,992	\$4,305	\$313	5.0	17.4		\$3,864	\$4,305	\$441	0.6	6.2		\$3,864	\$4,305	\$441	0.6	2.5
\$3,992	\$4,305	\$313	5.0	16.5		\$3,864	\$4,305	\$441	0.6	6.2	Ī	\$3,864	\$4,305	\$441	0.6	3.1
\$3,559	\$3,968	\$409	2.1	7.3		\$3,474	\$3,968	\$494	0.3	2.9	Ī	\$3,474	\$3,968	\$494	0.3	2.1
\$3,992	\$4,305	\$313	5.0	20.8		\$3,864	\$4,305	\$441	0.6	5.4	Ī	\$3,864	\$4,305	\$441	0.6	3.5
\$1,780	\$1,984	\$205	5.1	20.8		\$1,737	\$1,984	\$247	0.6	6.5	Ī	\$1,737	\$1,984	\$247	0.6	3.4
\$3,992	\$4,305	\$313	5.0	16.5		\$3,864	\$4,305	\$441	0.6	6.1	Ī	\$3,864	\$4,305	\$441	0.6	2.5
			2.1					<u> </u>		3.1					0.3	1.8
\$5,987	\$6,458	\$470		7.5		\$5,796	\$6,458	\$662	0.3		Ī	\$5,796	\$6,458	\$662		
\$3,992	\$4,305	\$313	2.1	6.7		\$3,864	\$4,305	\$441	0.3	2.9	Ī	\$3,864	\$4,305	\$441	0.3	2.1
\$3,992	\$4,305	\$313	4.9	15.4		\$3,864	\$4,305	\$441	0.6	3.6		\$3,864	\$4,305	\$441	0.6	3.6
\$5,987	\$6,458	\$470	5.0	15.7		\$5,796	\$6,458	\$662	0.6	2.8		\$5,796	\$6,458	\$662	0.6	2.8
\$3,992	\$4,305	\$313	5.0	14.6		\$3,864	\$4,305	\$441	0.6	3.6		\$3,864	\$4,305	\$441	0.6	3.6
\$3,992	\$4,305	\$313	7.6	21.1		\$3,864	\$4,305	\$441	1	3.1		\$3,864	\$4,305	\$441	1.0	3.1
\$8,898	\$9,921	\$1,023	7.6	17.7		\$8,685	\$9,921	\$1,236	1	3.0		\$8,685	\$9,921	\$1,236	1.0	3.0
\$1,996	\$2,153	\$157	9.1	22.7		\$1,932	\$2,153	\$221	1.6	4.3		\$1,932	\$2,153	\$221	1.6	4.3
\$3,992	\$4,305	\$313	7.4	22.2		\$3,864	\$4,305	\$441	1	4.5		\$3,864	\$4,305	\$441	1.0	4.5
\$5,987	\$6,458	\$470	7.6	16.8		\$5,796	\$6,458	\$662	1	2.5		\$5,796	\$6,458	\$662	1.0	2.5
\$5,339	\$5,953	\$614	5.0	12.9		\$5,211	\$5,953	\$742	0.6	3.1		\$5,211	\$5,953	\$742	0.6	3.1
\$1,996	\$2,153	\$157	7.8	16.7		\$1,932	\$2,153	\$221	1.6	3.7		\$1,932	\$2,153	\$221	1.6	3.7
\$12,457	\$13,889	\$1,433	7.6	15.7		\$12,159	\$13,889	\$1,730	1	2.3		\$12,159	\$13,889	\$1,730	1.0	2.3
\$5,987	\$6,458	\$470	5.0	11.5		\$5,796	\$6,458	\$662	0.6	2.7		\$5,796	\$6,458	\$662	0.6	2.7
\$1,996	\$2,153	\$157	7.8	20.5		\$1,932	\$2,153	\$221	1.6	5.3		\$1,932	\$2,153	\$221	1.6	5.3
\$15,966	\$17,220	\$1,254	7.6	15.1		\$15,456	\$17,220	\$1,764	1	2.3	Ī	\$15,456	\$17,220	\$1,764	1.0	2.3
\$3,992	\$4,305	\$313	8.0	20.0		\$3,864	\$4,305	\$441	1.6	4.4	Ī	\$3,864	\$4,305	\$441	1.6	4.4
\$7,983	\$8,610	\$627	7.6	17.4		\$7,728	\$8,610	\$882	1	2.9	Ī	\$7,728	\$8,610	\$882	1.0	2.9
\$7,118	\$7,937	\$819	7.5	17.0		\$6,948	\$7,937	\$989	1	3.8	Ī	\$6,948	\$7,937	\$989	1.0	3.8
\$7,983	\$8,610	\$627	5.1	9.8		\$7,728	\$8,610	\$882	0.6	1.6	Ī	\$7,728	\$8,610	\$882	0.6	1.6
\$5,987	\$6,458	\$470	7.8	18.8		\$5,796	\$6,458	\$662	1.6	4.8	Ī	\$5,796	\$6,458	\$662	1.6	4.8
\$7,983	\$8,610	\$627	7.6	15.7		\$7,728	\$8,610	\$882	1	2.4	Ī	\$7,728	\$8,610	\$882	1.0	2.4
\$7,983	\$8,610	\$627	11.4	25.6		\$7,728	\$8,610	\$882	2.1	5.2	Ī	\$7,728	\$8,610	\$882	2.1	5.2
\$5,339	\$5,953	\$614	11.5	24.9		\$5,211	\$5,953	\$742	2.2	5.1	Ī	\$5,211	\$5,953	\$742	2.2	5.1
\$3,559	\$3,968	\$409	11.2	25.2		\$3,474	\$3,968	\$494	2.2	5.8	Ī	\$3,474	\$3,968	\$494	2.2	5.8
\$1,780	\$1,984	\$205	11.5	24.9		\$1,737	\$1,984	\$247	2.2	5.0	Ī	\$1,737	\$1,984	\$247	2.2	5.0
\$5,339	\$5,953	\$614	7.6	17.8		\$5,211	\$5,953	\$742	1	4.0	Ī	\$5,211	\$5,953	\$742	1.0	4.0
\$5,987	\$6,458	\$470	7.8	12.5		\$5,796	\$6,458	\$662	1	2.0	Ī	\$5,796	\$6,458	\$662	1.0	2.0
\$3,992	\$4,305	\$313	7.7	13.8		\$3,864	\$4,305	\$441	1	3.5	Ī	\$3,864	\$4,305	\$441	1.0	3.5
\$1,780	\$1,984	\$205	1.8	3.0		\$1,737	\$1,984	\$247	0.2	0.9	Ī	\$1,737	\$1,984	\$247	0.2	0.9
\$3,559	\$3,968	\$409	7.9	23.2		\$3,474	\$3,968	\$494	1.6	5.8	Ī	\$3,474		\$494	1.6	5.8
\$0	\$4,440	\$4,440	1.9	6.7		\$5,474	\$4,440	\$4,440	2.1	5.4	Ī	\$0,474		\$4,440	2.1	5.4
										5.4						
\$0	\$5,328	\$5,328	1.9	7.0		\$0	\$5,328	\$5,328	2.1	5.7		\$0	\$5,328	\$5,328	2.1	5.7
\$247,245	\$278,897	\$31,652	6.6	16.6		\$239,886	\$278,897	\$39,011	1.1	3.8		\$239,886	\$278,897	\$39,011	1.1	3.5
Δhhre		А	1.0	1.0	$\overline{}$			N TO	_	. * 11	$\sim$	1 0	1	• .1	10 10	

Abbreviations: A = alfalfa, Cg = corn grain, NT = no-till, Ola = Oatlage with alfalfa seeding spring.

 $Table\ 2.\ Comparison\ of\ Corn\ Grain\ SNAP-Plus\ Simulations\ for\ Case\ 1.$ 

Corn	Corn Grain, Spring Chisel, Avg Yields  Corn Grain, No-Till, Avg Yields, Winter Spreading  Corn Grain, No-Till, Avg Yields, No Winter Spreading												
SC/W	interSpreading:	Ola-A-A-A	Cg-Cg	NT/W	VinterSpreading	g: Ola-A-A-A	Cg-Cg	NT/Ne	NT/NoWinterSpreading: Ola-A-A-A-Cg-Cg				
Costs	Revenues	Profits	P-Index	Costs	Revenues	Profits	P-Index	Costs	Revenues	Profits	P-Index		
\$1,839	\$1,941	\$102	8.1	\$1,775	\$1,941	\$166	2.2	\$1,775	\$1,941	\$166	2.2		
\$3,245	\$3,544	\$300	8.3	\$3,159	\$3,544	\$385	3.1	\$3,159	\$3,544	\$385	3.1		
\$7,354	\$7,762	\$408	9.4	\$7,099	\$7,762	\$663	3.6	\$7,099	\$7,762	\$663	3.6		
\$7,354	\$7,762	\$408	6.8	\$7,099	\$7,762	\$663	1.8	\$7,099	\$7,762	\$663	1.8		
\$7,354	\$7,762	\$408	9.5	\$7,099	\$7,762	\$663	3.5	\$7,099	\$7,762	\$663	3.5		
\$7,354	\$7,762	\$408	8.6	\$7,099	\$7,762	\$663	2.3	\$7,099	\$7,762	\$663	2.3		
\$7,354	\$7,762	\$408	10.2	\$7,099	\$7,762	\$663	3.8	\$7,099	\$7,762	\$663	3.8		
\$3,677	\$3,881	\$204	7.5	\$3,549	\$3,881	\$332	2.0	\$3,549	\$3,881	\$332	2.0		
\$3,245	\$3,544	\$300	8.9	\$3,159	\$3,544	\$385	5.4	\$3,159	\$3,544	\$385	2.3		
\$3,677	\$3,881	\$204	9.1	\$3,549	\$3,881	\$332	5.5	\$3,549	\$3,881	\$332	2.4		
\$3,677	\$3,881	\$204	9.0	\$3,549	\$3,881	\$332	5.6	\$3,549	\$3,881	\$332	2.5		
\$3,245	\$3,544	\$300	4.2	\$3,159	\$3,544	\$385	2.7	\$3,159	\$3,544	\$385	1.9		
\$3,677	\$3,881	\$204	9.7	\$3,549	\$3,881	\$332	5.0	\$3,549	\$3,881	\$332	2.2		
\$1,622	\$1,772	\$150	10.3	\$1,580	\$1,772	\$192	5.7	\$1,580	\$1,772	\$192	3.0		
\$3,677	\$3,881	\$204	8.9	\$3,549	\$3,881	\$332	5.5	\$3,549	\$3,881	\$332	2.4		
\$5,516	\$5,822	\$306	4.4	\$5,324	\$5,822	\$497	2.8	\$5,324	\$5,822	\$497	1.5		
\$3,677	\$3,881	\$204	4.0	\$3,549	\$3,881	\$332	2.7	\$3,549	\$3,881	\$332	1.3		
\$3,677	\$3,881	\$204	7.0	\$3,549	\$3,881	\$332	2.9	\$3,549	\$3,881	\$332	2.9		
\$5,516	\$5,822	\$306	6.3	\$5,324	\$5,822	\$497	2.0	\$5,324		\$497	2.0		
\$3,677	\$3,881	\$204	6.8	\$3,549	\$3,881	\$332	3.0	\$3,549		\$332	3.0		
\$3,677	\$3,881	\$204	7.9	\$3,549	\$3,881	\$332	2.0	\$3,549		\$332	2.0		
\$8,111	\$8,861	\$750	7.0	\$7,899	\$8,861	\$962	2.0	\$7,899	\$8,861	\$962	2.0		
\$1,839	\$1,941	\$102	7.6	\$1,775	\$1,941	\$166	2.0	\$1,775		\$166	2.0		
\$3,677	\$3,881	\$204	9.6	\$3,549	\$3,881	\$332	3.4	\$3,549		\$332	3.4		
\$5,516	\$5,822	\$306	4.3	\$5,324	\$5,822	\$497	1.6	\$5,324	\$5,822	\$497	1.6		
\$4,867	\$5,317	\$450	5.9	\$4,739	\$5,317	\$577	2.5	\$4,739		\$577	2.5		
\$1,839	\$1,941	\$102	6.5	\$1,775	\$1,941	\$166	1.8	\$1,775		\$166	1.8		
\$11,356	\$12,405	\$1,050	5.8	\$11,058	\$12,405	\$1,347	1.4	\$11,058	\$12,405	\$1,347	1.4		
\$5,516	\$5,822	\$306	5.1	\$5,324	\$5,822	\$497	2.2	\$5,324	\$5,822	\$497	2.2		
\$1,839	\$1,941	\$102	8.6	\$1,775	\$1,941	\$166	3.1	\$1,775		\$166	3.1		
\$14,708	\$15,524	\$816	5.7	\$14,198	\$15,524	\$1,327	1.5	\$14,198		\$1,327	1.5		
\$3,677	\$3,881	\$204	7.6	\$3,549	\$3,881	\$332	2.1	\$3,549	\$3,881	\$332	2.1		
\$7,354	\$7,762	\$408	6.7	\$7,099	\$7,762	\$663	2.0	\$7,099		\$663	2.0		
\$6,489	\$7,762	\$600	7.5	\$6,319	\$7,782	\$770	2.9	\$6,319		\$770	2.9		
\$7,354	\$7,762	\$408	3.8	\$7,099	\$7,762	\$663	1.1	\$7,099	\$7,762	\$663	1.1		
\$5,516	\$5,822	\$306	7.9	\$5,324	\$5,822	\$497	2.8	\$5,324	\$5,822	\$497	2.8		
		\$408					1.6				1.6		
\$7,354	\$7,762		6.0	\$7,099	\$7,762	\$663		\$7,099		\$663			
\$7,354 \$4,867	\$7,762 \$5,317	\$408 \$450	9.3	\$7,099 \$4,739	\$7,762 \$5,317	\$663 \$577	2.3	\$7,099 \$4,739		\$663 \$577	2.3		
			8.9										
\$3,245	\$3,544	\$300	9.9	\$3,159	\$3,544	\$385	2.9	\$3,159		\$385	2.9		
\$1,622	\$1,772	\$150	8.9	\$1,580	\$1,772	\$192	2.1	\$1,580		\$192	2.1		
\$4,867	\$5,317	\$450	7.9	\$4,739	\$5,317	\$577	3.1	\$4,739		\$577	3.1		
\$5,516	\$5,822	\$306	4.8	\$5,324	\$5,822	\$497	1.3	\$5,324		\$497	1.3		
\$3,677	\$3,881	\$204	6.5	\$3,549	\$3,881	\$332	2.9	\$3,549		\$332	2.9		
\$1,622	\$1,772	\$150	1.5	\$1,580	\$1,772	\$192	0.7	\$1,580		\$192	0.7		
\$3,245	\$3,544	\$300	9.5	\$3,159	\$3,544	\$385	3.2	\$3,159		\$385	3.2		
\$0	\$4,440	\$4,440	5.4	\$0	\$4,440	\$4,440	5.4	\$0	\$4,440	\$4,440	5.4		
\$0	\$5,328	\$5,328	5.7	\$0	\$5,328	\$5,328	5.7	\$0		\$5,328	5.7		
\$227,114	\$251,761	\$24,647	7.1	\$219,755	\$251,761	\$32,006	2.8	\$219,755	\$251,761	\$32,006	2.5		

Abbreviations: A = alfalfa, Cg = corn grain, NT = no-till, Ola = Oatlage with alfalfa seeding spring.

Table 3. Aggregate Farm Results Over Full Rotation for Case 1.

Weighted Farm Averages Over Full Rotation (six years)	Change in Profits (Loss)	Change in P Index	Profit (Loss) per Change in P Index		
Corn Silage, No-Till, Average Yields, Winter Spreading	\$7,359	(12.7)	\$15.05		
Corn Silage, No-Till, Average Yields, No Winter Spreading	\$7,359	(13.0)	\$14.54		
Corn Grain, Spring Chisel, Average Yields	(\$7,005)	(9.5)	(\$19.63)		
Corn Grain, No-Till, Average Yields, Winter Spreading	\$355	(13.7)	\$0.64		
Corn Grain, No-Till, Average Yields, No Winter Spreading	\$355	(14.0)	\$0.59		

 $Table\ 4.\ Comparison\ of\ BASE\ and\ Two\ SNAP-Plus\ Simulations\ for\ Case\ 2.$ 

OFG 14: BASE					OFG 14: Cg-43 and Csl-11					OFG 14: NoTill-43 (Csl)			
Costs	Revenues	Profits	P-Index		Costs	Revenues	Profits	P-Index		Costs	Revenues	Profits	P-Index
\$12,017	\$11,848	-\$169	2.6		\$12,017	\$11,848	-\$169	2.9		\$12,017	\$11,848	-\$169	2.6
\$4,604	\$9,507	\$4,903	1.5	1	\$4,604	\$9,507	\$4,903	1.5	1	\$4,604	\$9,507	\$4,903	1.5
\$10,301	\$10,156	-\$145	3.3	1	\$10,301	\$10,156	-\$145	3.3	1	\$10,301	\$10,156	-\$145	3.3
\$32,504	\$30,996	-\$1,508	4.0	1	\$32,504	\$30,996	-\$1,508	4.0	1	\$32,504	\$30,996	-\$1,508	4.0
\$4,604	\$9,507	\$4,903	1.4	1	\$4,604	\$9,507	\$4,903	1.4	1	\$4,604	\$9,507	\$4,903	1.4
\$12,888	\$11,855	-\$1,033	2.0	1	\$12,888	\$11,855	-\$1,033	2.0	1	\$12,888	\$11,855	-\$1,033	2.0
\$20,601	\$20,311	-\$290	3.8	1	\$20,601	\$20,311	-\$290	3.8		\$20,601	\$20,311	-\$290	3.8
\$2,302	\$4,753	\$2,452	2.6	1	\$2,302	\$4,753	\$2,452	2.6		\$2,302	\$4,753	\$2,452	2.6
\$4,604	\$9,507	\$4,903	0.9	1	\$4,604	\$9,507	\$4,903	0.9		\$4,604	\$9,507	\$4,903	0.9
\$41,202	\$40,622	-\$580	2.9		\$41,202	\$40,622	-\$580	2.9		\$41,202	\$40,622	-\$580	2.9
\$4,604	\$9,507	\$4,903	1.2	1	\$4,604	\$9,507	\$4,903	1.2		\$4,604	\$9,507	\$4,903	1.2
\$10,735	\$9,244	-\$1,490	3.6	]	\$10,735	\$9,244	-\$1,490	3.6	1	\$10,735	\$9,244	-\$1,490	3.6
\$14,578	\$30,105	\$15,527	1.6	L	\$14,578	\$30,105	\$15,527	1.6	L	\$14,578	\$30,105	\$15,527	1.6
\$8,946	\$7,704	-\$1,242	2.7		\$8,055	\$7,410	-\$646	3.9		\$8,946	\$7,704	-\$1,242	2.7
\$6,444	\$5,928	-\$517	2.3		\$6,444	\$5,928	-\$517	2.3		\$6,444	\$5,928	-\$517	2.3
\$12,888	\$11,855	-\$1,033	5.4	1	\$12,888	\$11,855	-\$1,033	5.4		\$12,888	\$11,855	-\$1,033	5.4
\$18,884	\$18,619	-\$266	3.1		\$18,884	\$18,619	-\$266	3.1		\$18,884	\$18,619	-\$266	3.1
\$20,601	\$20,311	-\$290	1.7		\$20,601	\$20,311	-\$290	1.7		\$20,601	\$20,311	-\$290	1.7
\$5,951	\$5,342	-\$609	2.1		\$5,951	\$5,342	-\$609	2.1		\$5,951	\$5,342	-\$609	2.1
\$7,673	\$15,845	\$8,172	1.2		\$7,673	\$15,845	\$8,172	1.2		\$7,673	\$15,845	\$8,172	1.2
\$8,946	\$7,704	-\$1,242	1.8		\$8,946	\$7,704	-\$1,242	1.8		\$8,946	\$7,704	-\$1,242	1.8
\$7,786	\$6,536	-\$1,250	2.2		\$7,786	\$6,536	-\$1,250	2.2		\$7,786	\$6,536	-\$1,250	2.2
\$17,854	\$16,027	-\$1,827	3.1		\$17,854	\$16,027	-\$1,827	3.1		\$17,854	\$16,027	-\$1,827	3.1
\$11,509	\$23,767	\$12,258	1.5		\$11,509	\$23,767	\$12,258	1.5		\$11,509	\$23,767	\$12,258	1.5
\$44,636	\$44,008	-\$628	2.4		\$44,636	\$44,008	-\$628	2.4		\$44,636	\$44,008	-\$628	2.4
\$15,451	\$15,233	-\$217	1.6		\$15,451	\$15,233	-\$217	1.6		\$15,451	\$15,233	-\$217	1.6
\$18,884	\$18,619	-\$266	1.7		\$18,884	\$18,619	-\$266	1.7		\$18,884	\$18,619	-\$266	1.7
\$58,370	\$57,548	-\$821	2.0		\$58,370	\$57,548	-\$821	2.0		\$58,370	\$57,548	-\$821	2.0
\$18,884	\$18,619	-\$266	3.7		\$18,884	\$18,619	-\$266	3.7		\$18,884	\$18,619	-\$266	3.7
\$10,301	\$10,156	-\$145	2.5		\$10,301	\$10,156	-\$145	2.5		\$10,301	\$10,156	-\$145	2.5
\$23,357	\$19,608	-\$3,749	2.3		\$23,357	\$19,608	-\$3,749	2.3		\$23,357	\$19,608	-\$3,749	2.3
\$15,871	\$14,246	-\$1,624	4.0		\$15,871	\$14,246	-\$1,624	4.0		\$15,871	\$14,246	-\$1,624	4.0
\$128,210	\$122,262	-\$5,948	2.9		\$128,210	\$122,262	-\$5,948	2.9		\$128,210	\$122,262	-\$5,948	2.9
\$18,058	\$17,220	-\$838	2.4	4	\$18,058	\$17,220	-\$838	2.4	Į.	\$18,058	\$17,220	-\$838	2.4
\$23,357	\$19,608	-\$3,749	4.8	1	\$23,357	\$19,608	-\$3,749	4.8	l	\$23,357	\$19,608	-\$3,749	4.8
\$3,069	\$6,338	\$3,269	0.9	1	\$3,069	\$6,338	\$3,269	0.9	l	\$3,069	\$6,338	\$3,269	0.9
\$14,499	\$13,337	-\$1,162	1.7	1	\$14,499	\$13,337	-\$1,162	1.7	l	\$14,499	\$13,337	-\$1,162	1.7
\$13,734	\$13,541	-\$193	4.3	-	\$13,734	\$13,541	-\$193	4.3	l	\$13,734	\$13,541	-\$193	4.3
\$13,734	\$13,541	-\$193	3.3	1	\$13,734	\$13,541	-\$193	3.3	ł	\$13,734	\$13,541	-\$193	3.3
\$14,499	\$13,337	-\$1,162	2.7	-	\$14,499	\$13,337	-\$1,162	2.7	ł	\$14,499	\$13,337	-\$1,162	2.7
\$5,150	\$5,078	-\$72	3.2	-	\$5,150	\$5,078	-\$72	3.2	ł	\$5,150	\$5,078	-\$72	3.2
\$32,618	\$32,159	-\$459	3.0	-	\$32,618	\$32,159	-\$459	3.0	ł	\$32,618	\$32,159	-\$459	3.0
\$17,854 \$41,202	\$16,027 \$40,622	-\$1,827 -\$580	3.1	1	\$17,854 \$41,202	\$16,027 \$40,622	-\$1,827 -\$580	3.1	ł	\$17,854 \$41,202	\$16,027 \$40,622	-\$1,827 -\$580	3.1
			3.5	$\vdash$					<del>                                     </del>		, .		
\$54,173	\$51,660	-\$2,513	2.8	$\vdash$	\$54,173	\$51,660	-\$2,513	2.8	_	\$54,173	\$51,660	-\$2,513	2.8
\$11,716	\$10,995	-\$721	10.6		\$10,301	\$10,156	-\$145	4.5		\$11,369	\$10,995	-\$374	4.0
\$25,777	\$23,710	-\$2,066	4.3	-	\$25,777	\$23,710	-\$2,066	4.3	ł	\$25,777	\$23,710	-\$2,066	4.3
\$29,185	\$28,774	-\$411	5.7	-	\$29,185	\$28,774	-\$411	5.5	ł	\$29,185	\$28,774	-\$411	5.7
\$22,250	\$45,950	\$23,700	2.6	$\vdash$	\$22,250	\$45,950	\$23,700	2.6		\$22,250	\$45,950	\$23,700	2.6
\$977,861	\$1,019,752	\$41,891	2.9		\$975,555	\$1,018,618	\$43,063	2.8		\$977,514	\$1,019,752	\$42,238	2.8

Abbreviations: Cg = corn grain, Csl = corn silage, OFG = on farmer's ground.

Table 5. Aggregate Farm Results Over Full Rotation for Case 2.

Weighted Farm Averages Over	Change in	Change in P Index	Profit (Loss) per		
Full Rotation (six years)	Profits (Loss)		Change in P Index		
OFG 14: Cg-43 and Csl-11	\$1,172	(0.1)	\$3.12		
OFG 14: No-Till-43 (Csl)	\$347	(0.1)	\$0.51		

Abbreviations: Cg = corn grain, Csl = corn silage, OFG = on farmer's ground.

Table 6. Comparison of BASE and Two SNAP–Plus Simulations for Case 3.

OFG 16: BASE					OFG 16: 10&16, Csl==>Cg					OFG 16: +9b&11&13, Pg ==> PRg			
Costs	Revenues	Profits	P-Index		Costs	Revenues	Profits	P-Index		Costs	Revenues	Profits	P-Index
\$21,711	\$38,453	\$16,742	3.7		\$21,711	\$38,453	\$16,742	3.7		\$21,711	\$38,453	\$16,742	3.7
\$0	\$5,051	\$5,051	1.4		\$0	\$5,051	\$5,051	1.4	1 I	\$0	\$5,051	\$5,051	1.8
\$0	\$6,734	\$6,734	1.4		\$0	\$6,734	\$6,734	1.4	1 I	\$0	\$6,734	\$6,734	1.8
\$0	\$5,051	\$5,051	1.4		\$0	\$5,051	\$5,051	1.4	1 I	\$0	\$5,051	\$5,051	1.4
\$0	\$1,684	\$1,684	3.9		\$0	\$1,684	\$1,684	3.9	lĺ	\$0	\$1,684	\$1,684	4.1
\$0	\$842	\$842	4.6		\$0	\$842	\$842	4.6	lĺ	\$0	\$842	\$842	4.8
\$0	\$842	\$842	0.4		\$0	\$842	\$842	0.4	1 I	\$0	\$842	\$842	0.6
\$0	\$842	\$842	3.9		\$0	\$842	\$842	3.9	1 I	\$0	\$842	\$842	4.1
\$25,958	\$36,393	\$10,435	5.9	1	\$25,958	\$36,393	\$10,435	5.9	1 I	\$25,958	\$36,393	\$10,435	5.9
\$12,979	\$18,196	\$5,217	7.1		\$10,856	\$19,227	\$8,371	3.8	l i	\$10,856	\$19,227	\$8,371	3.8
\$0	\$2,525	\$2,525	8.5		\$0	\$2,525	\$2,525	8.5		\$0	\$4,545	\$4,545	5.6
\$9,650	\$17,090	\$7,441	2.4		\$9,650	\$17,090	\$7,441	2.4	1 I	\$9,650	\$17,090	\$7,441	2.4
\$0	\$3,367	\$3,367	6.4		\$0	\$3,367	\$3,367	6.4		\$0	\$6,061	\$6,061	2.9
\$18,093	\$32,045	\$13,952	3.5		\$18,093	\$32,045	\$13,952	3.5	1 [	\$18,093	\$32,045	\$13,952	3.5
\$8,653	\$12,131	\$3,478	9.8		\$7,237	\$12,818	\$5,581	3.0	1 [	\$7,237	\$12,818	\$5,581	3.0
\$8,443	\$14,954	\$6,511	2.7		\$8,443	\$14,954	\$6,511	2.7	l í	\$8,443	\$14,954	\$6,511	2.7
\$12,062	\$21,363	\$9,301	2.5		\$12,062	\$21,363	\$9,301	2.5	l i	\$12,062	\$21,363	\$9,301	2.5
\$9,650	\$17,090	\$7,441	2.3		\$9,650	\$17,090	\$7,441	2.3	l i	\$9,650	\$17,090	\$7,441	2.3
\$6,031	\$10,682	\$4,651	4.5	1	\$6,031	\$10,682	\$4,651	4.5	lĺ	\$6,031	\$10,682	\$4,651	4.5
\$2,412	\$4,273	\$1,860	3.0	1	\$2,412	\$4,273	\$1,860	3.0	l í	\$2,412	\$4,273	\$1,860	3.0
\$2,412	\$4,273	\$1,860	3.8	1	\$2,412	\$4,273	\$1,860	3.8	lĺ	\$2,412	\$4,273	\$1,860	3.8
\$4,825	\$8,545	\$3,720	3.4	1	\$4,825	\$8,545	\$3,720	3.4	lĺ	\$4,825	\$8,545	\$3,720	3.4
\$3,619	\$6,409	\$2,790	3.6		\$3,619	\$6,409	\$2,790	3.6	lĺ	\$3,619	\$6,409	\$2,790	3.6
\$4,825	\$8,545	\$3,720	2.6	1	\$4,825	\$8,545	\$3,720	2.6	lĺ	\$4,825	\$8,545	\$3,720	2.6
\$9,650	\$17,321	\$7,671	2.4	1	\$9,650	\$17,321	\$7,671	2.4	1 i	\$9,650	\$17,321	\$7,671	2.4
\$38,598	\$68,362	\$29,764	1.4	1	\$38,598	\$68,362	\$29,764	1.4	1 i	\$38,598	\$68,362	\$29,764	1.4
\$3,619	\$6,409	\$2,790	4.0	1	\$3,619	\$6,409	\$2,790	4.0	1 i	\$3,619	\$6,409	\$2,790	4.0
\$203,187	\$369,469	\$166,281	3.5		\$199,649	\$371,186	\$171,537	3.1		\$199,649	\$375,900	\$176,251	3.0

Abbreviations: Cg = corn grain, Csl = corn silage, OFG= on farmer's ground, PRg = pasture rotational, grass.

Table 7. Aggregate Farm Results Over Full Rotation for Case 3.

Weighted Farm Averages Over	Change in Profits	Change in P Index	Profit (Loss) per		
Full Rotation (seven years)	(Loss)		Change in P Index		
OFG 16: 10&16, Csl==>Cg	\$5,256	(0.4)	\$53.89		
OFG 16: +9b≪&13, Pg==>PRg	\$9,970	(0.4)	\$80.54		

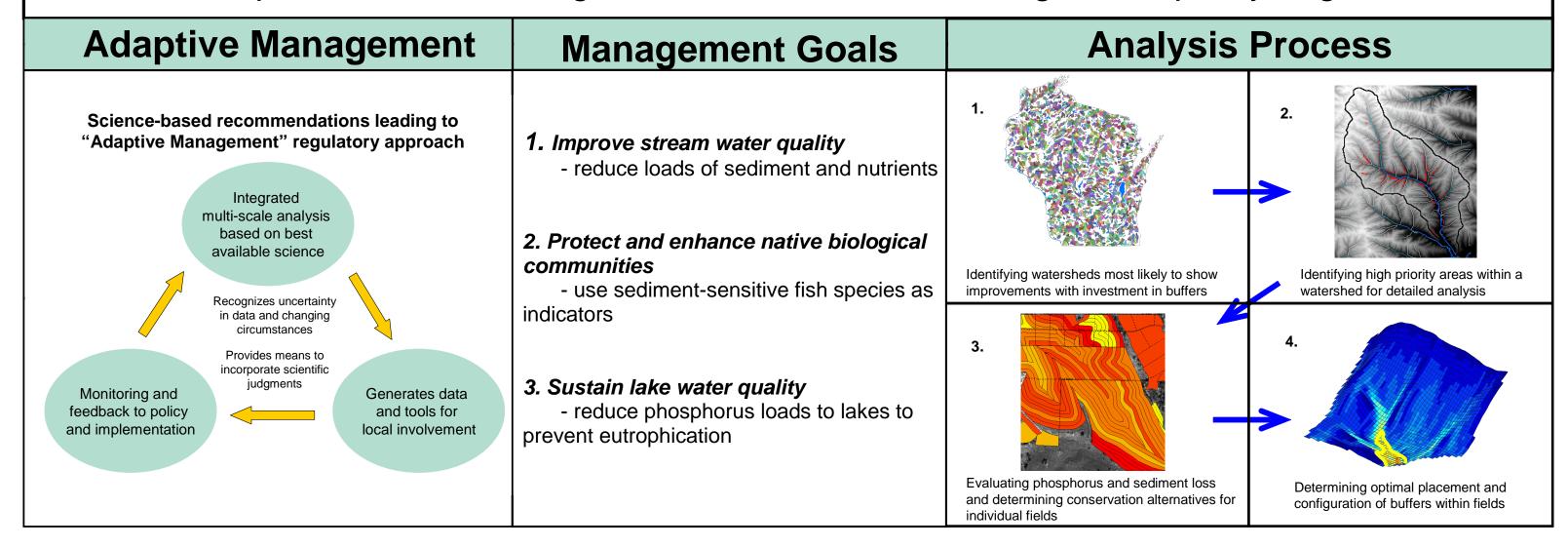
Abbreviations: Cg = corn grain, Csl = corn silage, OFG = on farmer's ground, PRg = pasture rotational, grass.

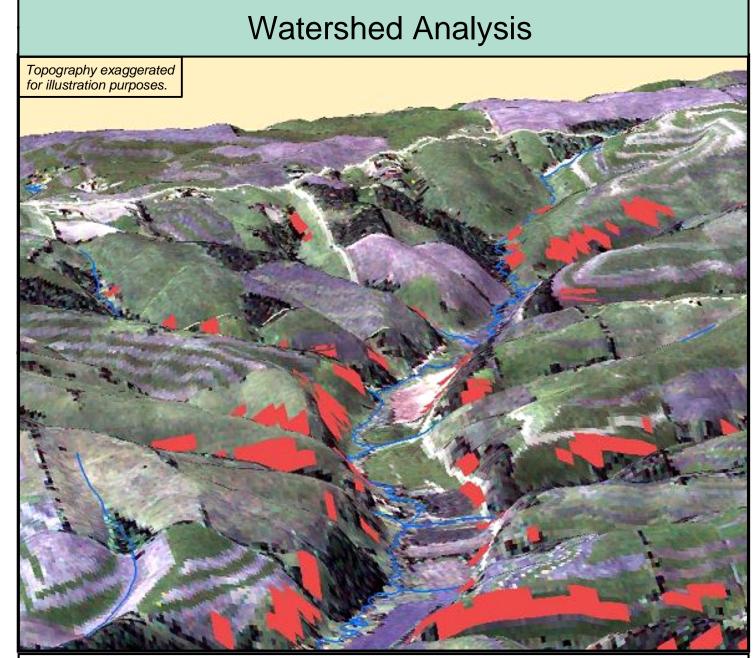
# Wisconsin Buffer Initiative

# WBI Statewide Watershed Ranking Map Legend WBI Watershed Ranking Process Wisconsin watersheds between 8 and 40 square miles are separately evaluated on their potential to accomplish the three WBI management goals. A ranked list of watershed is created. Top-ranked groups of watersheds have the greatest likelihood of responding to buffers and related conservation practices.

# Wisconsin Buffer Initiative Project Description

The Wisconsin Buffer Initiative (WBI) is an effort to make science-based recommendations to the Wisconsin Department of Natural Resources for the development of state rules governing agricultural pollution. Under the guidance of a broadly representative advisory committee, researchers at the University of Wisconsin-Madison provide ideas and approaches at four scales. The overall goal is to identify areas where buffers, in conjunction with other conservation practices, have the greatest likelihood of reducing water quality degradation.





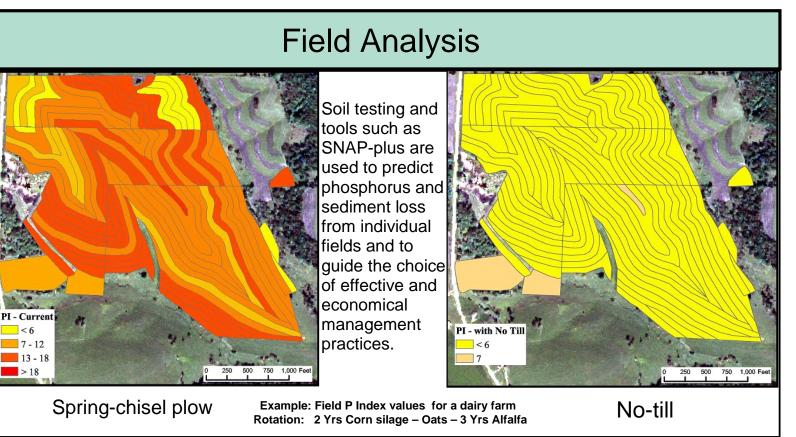
Soil and phosphorus delivery are calculated for all fields in a selected watershed, beginning with the areas that are most vulnerable to erosion (shown above in red) according to a GIS-based USLE analysis.

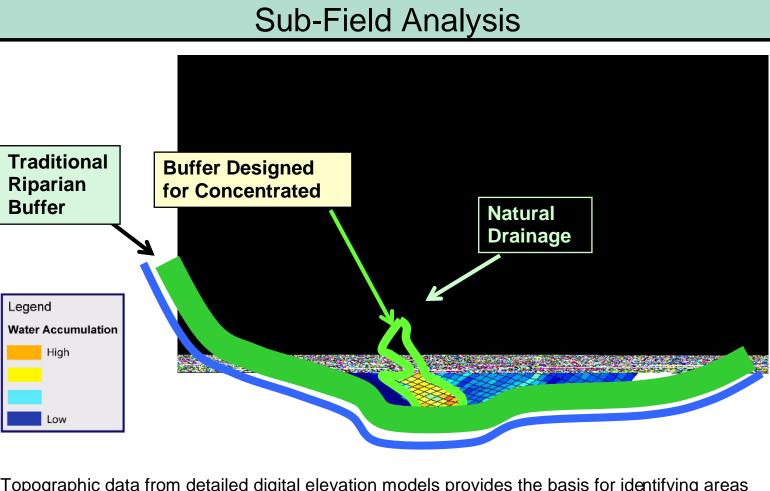
### **Credits and Contacts**

- WBI Principal Investigator: Pete Nowak, University of Wisconsin-Madison
- WBI Researchers: Laura Good, John Norman, Larry Bundy, Jake Vander Zanden, Jeff Maxted, Matt Diebel, Christine Molling – University of Wisconsin-Madison
- Map Design and Development: Steve Ventura and Ben Webb, University of Wisconsin-

Natural Resource Conservation Service

This research was funding in part through a grant from the USDA Website: http://www.drs.wisc.edu/wbi/ November, 2005





Topographic data from detailed digital elevation models provides the basis for identifying areas of convergent flow and locating buffers for greatest affect. Grass waterways and buffers in these areas are more effective and may be more economical than simple "ribbons" of grass between fields and streams (riparian buffers).